# Canadä <br> (2) Ontario 

Canada-United States-Ontario-Michigan Border Transportation Partnership

## Practical Alternatives Evaluation Working Paper

Air Quality Impact<br>Assessment

## Preface

The Detroit River International Crossing (DRIC) Environmental Assessment Study is being conducted by a partnership of the federal, state and provincial governments in Canada and the United States in accordance with the requirements of the Canadian Environmental Assessment Act (CEAA), the Ontario Environmental Assessment Act (OEAA), and the U.S. National Environmental Policy Act (NEPA). In 2006, the Canadian and U.S. Study Teams completed an assessment of illustrative crossing, plaza and access road alternatives. This assessment is documented in two reports: Generation and Assessment of Illustrative Alternatives Report - Draft November 2006) (Canadian side) and Evaluation of Illustrative Alternatives Report (December 2006) (U.S. side). The results of this assessment led to the identification of an Area of Continued Analysis (ACA) as shown in Exhibit 1.

Within the ACA, practical alternatives were developed for the crossings, plazas and access routes alternatives. The evaluation of practical crossing, plaza and access road alternatives is based on the following seven factors:

- Changes to Air Quality
- Protection of Community and Neighbourhood Characteristics
- Consistency with Existing and Planned Land Use
- Protection of Cultural Resources
- Protection of the Natural Environment

Improvements to Regional Mobility
Cost and Constructability
This report pertains to the Changes to Air Quality factor and is one of several reports that will be used in support of the evaluation of practical alternatives and the selection of the technically and environmentally preferred alternative. This report will form a part of the environmental assessment documentation for this study.

Additional documentation pertaining to the evaluation of practical alternatives is available for viewing/downloading at the study website (www.partnershipborderstudy.com).

# Practical Alternatives Evaluation Working Paper 

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### 1.0 INTRODUCTION

Changes to Air Quality is one of the seven factors being used to assess the potential effects of the various transportation improvement alternatives currently being studied by the Detroit River International Crossing (DRIC) study team.

Due to the proximity to the Canada-U.S. border and the resulting high rate of traffic through the City of Windsor, vehicular emissions and their effect on existing air quality are of concern in the Windsor-Essex area In addition. The City of Windsor also has a relatively high fraction of diesel powered transport trucks that are used to move goods into and out of Canada. Diesel exhaust is highly visible and odourous, and there is increasing evidence that there are health effects associated with it. Thus, a primary objective of the Air Quality Assessment is to have a transportation solution that not only improves transportation in the Windsor-Essex area, but also improves the overall air quality in the local area, if possible.

This report outlines the methodology and tools used to conduct the Air Quality Assessment, and presents the results and evaluation of the each of the alternatives studied.

## $1.1 \quad$ Practical Alternatives Under Assessment

Five Practical Alternatives for the Access Road were presented in the public in March 2006 at the second round of DRIC Public Information Open Houses $(\mathrm{PIOH})$. These represent the best route alternatives developed from the Illustrative Alternatives with input from the public. The Alternatives are all located within the Area of Continued Analysis (ACA) as is shown in Figure 1.1.


The five Practical Alternatives for the Access Road are as follows:

- Alternative 1A - At grade freeway with one-way local access service roads located along each side;
- Alternative 1B - Depressed (below) grade freeway with one-way local access service drives located at grade along each side;
- Alternative 2A - At grade freeway with two-way local access service roads located along the approximate existing Huron Church Road / Highway 3 corridor;
- Alternative 2B - Depressed grade freeway with two-way local access service roads located at grade along the approximate Huron Church Road / Highway 3 corridor; and
- Alternative 3 - Tunnelled freeway with two-way local access service roads located at-grade along the approximate Huron Church Road / Highway 3 corridor.

In addition to these five alternatives, Alternatives 1A - 2B have two different alignment options (Option 1 \& Option 2) between St.Clair College and Howard


Also, four separate ventilation options were studied for Alternative 3. These are as follows:

- VB1A - use of two separate ventilation buildings to circulate and remove air from the tunnel. One vent building located approximately $1 / 3 r d$ of the distance from the south tunnel entrance/exit at the present Highway 401 terminus at Highway 3; the second vent building located approximately $1 / 3$ rd of the distance from the north tunnel entrance and exit, which is half way between Malden Rd. and Huron Church Road.
- VB1B - use of two separate ventilation buildings at the main tunnel entrance/exits to circulate and remove air from the tunnel. One vent building located approximately at the present Highway 401 terminus at

Highway 3; the second vent building located approximately half way between Malden Rd. and Huron Church Road.

- VB1C - use of a single ventilation building at the approximate half way point of the tunnel to circulate and remove air from the tunnel. One vent building located in the vicinity of Todd Lane/Cabana Rd.
- Jet Fans - use of multiple jet fans located in the tunnel interior to continuously circulate the tunnel air; assumes no vent buildings required.

The locations of the three vent building options are shown on Figure 1.3 below.


Four Plaza Alternatives and three river Crossing Alternatives were also examined, in various combinations. Each Plaza Alternative typically had several potential Crossing Alternatives, as follows:

- Plaza A

0 to Crossing A
0 to Crossing B
0 to Crossing C

- Plaza B
o to Crossing C
- Plaza B1
o to Crossing B
- Plaza C
o to Crossing C
The details of each of Access Road, Plaza and Crossing Alternatives are presented in Figure 1.4 below.
figure 1.4 Summary of Practical Alternatives


Potential air quality effects of the five Practical Alternatives for connecting routes, four Tunnel Ventilation Alternatives and seven combinations of Plaza/Crossing Alternatives were assessed using a combination of existing air monitoring data in combination with air dispersion modelling. Air dispersion modelling must be used to assess the impacts of future changes, such as implementation of the alternatives, in addition to changes in fuels, vehicle technologies and traffic volumes. The model being used is specifically designed
to assess impacts from roads and highways. The model incorporates the differences between moving vehicles, and queued vehicles that are idling, as well as differences in roads that are "at grade", depressed and bridges.

Two indicator pollutants were selected for this phase of the analysis to represent one gaseous compound and one particulate compound. These are Particulate Matter less than 2.5 microns $\left(\mathrm{PM}_{2.5}\right)$ and Nitrogen Oxides $\left(\mathrm{NO}_{\mathrm{x}}\right)$. Changes in the total predicted concentrations of these two air pollutants were compared for each alternative, as well as to existing conditions and a future "do nothing" condition.

## 1.2

## Area of Investigation

Since air quality does not respect local boundaries, a relatively broad area was included in the Air Quality Assessment. This comprised an approximate 10 km x 10 km area in West Windsor, from just south of the present Highway 401 terminus at Highway 3, 10 km north and 10 km west to the Detroit River. This is approximately the area depicted in Figure 1.1 that was presented earlier.

Potential air quality effects from roadways decrease with increasing distance from the roadway. Therefore the greatest effects will occur immediately adjacent to the roadway. For assessment of the potential affects on air quality of the Access Road Alternatives and Crossing Alternatives, an area located within 250 m on either side of the Right of Way (ROW) of each proposed Alternative was studied. Similar to the connecting route alternatives, the Plaza Alternatives were assessed within 250 m of the proposed facility property lines.

### 2.0 Existing Environmental Conditions

Assessment of the existing environmental conditions in the Windsor area is an important first step in the analysis of the various alternatiyes being studied. The existing conditions represent the benchmark to which future changes must be added (such as future traffic growth without implementation of any project related Alternatives). This forms the baseline conditions, and is also known as the No Build Alternative. All future changes related to the project are added to the existing conditions and evaluated against the baseline condition.

## 2.1

Climate and Meteorological Data

Characterization of the existing climate and meteorological conditions in the vicinity of the Huron Church Road / Highway 3 corridor is important because these are the main forces driving contaminant transport (dispersion) in the atmosphere. The direction and speed of the wind dictates the location and distance from the source that the pollutants may travel. The factors that influence the contaminant mixing in the atmosphere are described below.

The Windsor-Essex area has a middle latitude humid continental climate affected by Lake Erie and Lake St.Clair. The region is characterized by pronounced seasonal differences of weather and by a highly variable day-to-day weather pattern. Some periods in summer are essentially humid tropical (high temperatures, high humidity, afternoon thunderstorms, etc.). Some periods in winter are effectively polar (very cold, clear, dry). Precipitation occurs throughout the year.

The surface meteorological data used in the air dispersion modelling was obtained from the Windsor Airport meteorological station (2000 - 2004) which is approximately $5-7 \mathrm{~km}$ from the Huron Church Road / Highway 3 corridor. It is well exposed and represents the general wind flow pattern in the vicinity of the corridor since the area is generally flat. The upper air measurements used are from the closest upper air station (Pontiac, MI), which is located approximately 30 km to the northwest of the DRIC study area. In order to be considered representative, the wind and temperature data should be obtained from within 100 km of the study area, and the upper air data (which is a regional parameter) should be within 300 km . The stations used for this study are well within these parameters.

### 2.1.1 $\quad$ Near-Surface Temperature

Temperature and precipitation normals for the Windsor Airport (1971-2000) are presented in Table 2.1. "Normals" is the term commonly used for values of climatic elements averaged over a fixed standard period of years (usually 30 years).

Temperature near the surface of the earth controls the buoyant component of turbulence (vertical motion). Heat from the earth's surface heats the air near the ground causing it to rise. This mechanism reaches a maximum in early afternoon and is at a minimum near sunrise. This affects the dispersion of air pollutants through the influence of "thermal mixing" as the air mass rises.

Table 2.1 indicates that the mean (averaged over 30 years) daily minimum temperature is $-8.1^{\circ} \mathrm{C}$ in January and daily maximum temperature is $28^{\circ} \mathrm{C}$ in July at the Windsor Airport site. The annual mean temperature is $9.4^{\circ} \mathrm{C}$.

|  | Temperature | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec | Year |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Daily Average ( ${ }^{\circ} \mathrm{C}$ ) | -4.5 | -3.2 | 2 | 8.2 | 14.9 | 20 | 23 | 21.6 | 17 | 11 | 4.6 | -1.5 | 9.4 |
|  | Standard Deviation | 2.9 | 2.7 | 2.1 | 1.6 | 2.1 | 1.3 | 1.1 | 1.2 | 1.3 | 1.7 | 1.7 | 2.7 | 0.8 |
|  | Daily Maximum ( ${ }^{\circ} \mathrm{C}$ ) | -0.9 | 0.6 | 6.4 | 13 | 20.5 | 25 | 28 | 26.6 | 23 | 16 | 8.3 | 1.9 | 14 |
|  | Daily Minimum ( ${ }^{\circ} \mathrm{C}$ ) | -8.1 | -7 | -2.4 | 3 | 9.3 | 15 | 17 | 16.6 | 12 | 6.2 | 0.9 | -4.8 | 4.9 |
|  | Precipitation |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Rainfall (mm) | 29 | 33 | 55.6 | 81 | 80.7 | 90 | 82 | 79.7 | 96 | 64 | 67 | 47 | 805.2 |
|  | Snowfall (cm) | 35 | 28 | 20.6 | 4.3 | 0 | 0 | 0 | 0 | 0 | 0.7 | 8.3 | 30 | 126.6 |
|  | Precipitation (mm) | 58 | 57 | 75 | 85 | 80.8 | 90 | 82 | 79.7 | 96 | 65 | 76 | 75 | 918.3 |
|  | Days with Rainfall |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | $>=0.2 \mathrm{~mm}$ | 5.7 | 5.6 | 9.4 | 12 | 11.8 | 11 | 10 | 10 | 11 | 11 | 11 | 7.9 | 115.7 |
|  | Days With Snowfall |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | $>=0.2 \mathrm{~cm}$ | 13 | 9.1 | 6.7 | 2.3 | 0.03 | 0 | 0 | 0 | 0 | 0.3 | 3.8 | 10 | 45 |
|  | Days with Precipitation |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | $>=0.2 \mathrm{~mm}$ | 15 | 12 | 13.9 | 13 | 11.8 | 11 | 10 | 10 | 11 | 11 | 13 | 15 | 146.7 |
|  | Wind |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Days with Winds >= $52 \mathrm{~km} / \mathrm{hr}$ | 1.9 | 1.4 | 2.5 | 1.8 | 1.1 | 0.9 | 0.7 | 0.3 | 0.4 | 0.5 | 1.2 | 1.2 | 14 |
|  | Days with Winds >= $63 \mathrm{~km} / \mathrm{hr}$ | 0.6 | 0.4 | 0.7 | 0.7 | 0.5 | 0.3 | 0.4 | 0.2 | 0.1 | 0.2 | 0.3 | 0.3 | 4.7 |

Source: Environment Canada website, http://www.climate.weatheroffice.ec.gc.ca/climate_normals/index_e.html

The meteorological file used in the air dispersion modeling for this project requires hourly temperatures for each day in the year.

### 2.1.2 <br> Precipitation

Precipitation acts as an atmospheric cleansing mechanism, as contaminants in the air are generally washed out by precipitation. More precipitation produces more washout. For this study, the role of precipitation in the removal of pollutants
from the air was not considered, thereby generally providing conservatively high ground level concentrations.

As shown in Table 2.1 above, the Windsor area normally receives a total of 918.3 mm of precipitation per year, including 805.2 mm of rainfall and 126.6 cm of snowfall. The maximum mean monthly rainfalh is 96.2 mm , which occurs in September.
2.1.3 Atmospheric Stability

Normally, temperature decreases with increasing height above sea level. The relationship of the actual vertical temperature to the near-surface temperature determines the atmosphere's ability to resist or enhance vertical motion. The amount of vertical motion is a measure of the stability of the atmosphere.

The atmosphere can have three general stability states - unstable, neutral and stable. The stability scale normally used for air quality simulations varies from very unstable (A) through neutral (D) to very stable (F). The stability class distribution for the Windsor Airport station for the period 2000-2004 is presented in Table 2.2. At this station, neutral stability conditions \{D (neutral) + C (near neutral) $\}$ occur approximately $68 \%$ of the time and stable conditions (E, F) about $28 \%$ of the time. Stable conditions can produce higher concentrations of contaminants because of reduced turbulent mixing.

TABLE 2.2 - Stability CLass Distribution - Windsor Airport (2000-2004)

| Stability Class | \% Frequency |  |  |  |  |  | Descriptor |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathbf{2 0 0 0} \mathbf{- 2 0 0 4}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ |  |
| A | 0.5 | 0.4 | 0.8 | 0.6 | 0.4 | 0.4 | Unstable |
| B | 4.2 | 3.6 | 4.6 | 4.4 | 4.4 | 3.9 |  |
| C | 10.1 | 10.6 | 10.3 | 9.8 | 9.9 | 9.9 | Neutral |
| D | 57.0 | 56.0 | 56.2 | 57.1 | 57.0 | 58.6 |  |
| E | 13.3 | 13.6 | 14.0 | 13.2 | 12.8 | 13.1 | Stable |
| F | 14.9 | 15.8 | 14.2 | 15.0 | 15.5 | 14.1 |  |

The meteorological file used in the air dispersion modeling for this project requires hourly stability classes for each day in the year.

### 2.1.4 <br> Wind Direction

Wind direction is reported as the direction from which the wind blows and is based on surface ( 10 meter) observations. In general terms, if the wind does not
blow toward a receptor, there will be no impact from an upwind emission source. The wind blows in all directions with varying frequencies. Certain directions occur more frequently than others. These are known as the prevailing wind directions.

Figure 2.1 presents a wind rose for the Windsor Airport for the years 2000-2004. The prevailing wind is from the southwest, primarily during the summer months, with winds blowing from the west through southwest directions (i.e. from Southeast Michigan) approximately $32 \%$ of the time.

FIGURE 2.1 - WIND ROSE - WINDSOR AIRPORT (2000-2004)


$$
\begin{array}{ll}
\text { Percentage of Calms } \\
2000-2004 & 3.6 \% \\
2003 & 4.3 \%
\end{array}
$$

### 2.1.5 $\quad$ Wind Speed

Contaminant concentrations decrease with increasing wind speed as a result of atmospheric mixing. The wind speed used in the air quality modelling is based on surface observations from the Windsor Airport. Wind speed increases with height as surface friction is reduced. Variation of wind speed with height is built into the dispersion model used in this assessment. When wind speeds are high, there is good dispersion of gases and particles, but more potential for resuspension of surface dust. When wind speeds are near zero, the primary mechanism of pollutant transport away from a source is via diffusion, which can lead to very high pollutant concentrations near the ground. Calms were recorded $4.3 \%$ of the time at the Windsor Airport meteorological station (Figure 2.1) during 2003 compared with $3.6 \%$ for the 2000-2004 period.

The meteorological file used in the air dispersion modeling for this project requires hourly wind speed and directions for each day in the year.

Mixing Height
Another very important parameter in the dispersion of contaminants from a source is the "mixing height". This is the vertical extent through which the plume can be mixed. With a higher mixing height, there is a larger volume of air available within which the pollutants can mix, which results in lower concentrations. With a lower mixing height, the plume may become trapped resulting in higher concentrations.

The concept of mixing height is founded on the principle that heat transferred to the atmosphere at the earth's surface results in convection, vigorous vertical mixing and the establishment of a dry-adiabatic lapse rate [Holzworth 1967]. For annual and 24 hour average concentrations, the mixing height does not have much effect on the modelled ground level concentrations [Young \& Radonjic 1993]. For 1 hour average concentrations, however, mixing height is very important. The use of variable mixing heights, that are as close to the actual conditions as possible, improves the ability of the model to accurately predict downwind concentrations. For the sources that are close to the ground, the mixing heights do not play a major role.

The closest station having the upper air data necessary for this study is the Pontiac, Michigan. The mixing height data for each day in the 5 -year meteorological period (2000-2004) was developed using the Holzworth methodology. The surface values and the mean monthly minimum (morning) and
maximum (afternoon) mixing heights were then pre-processed through the U.S. EPA meteorological pre-processor (PCRAMMET) [U.S. EPA 1998] which combines surface and upper air measurements to create the hourly mixing heights which are required by the dispersion model. Missing data was filled in by interpolation. There were no significant blocks of data missing from this meteorological data set.

## Assessment Criteria



Environment Canada and the Ontario Ministry of the Environment (MOE) have set air quality objectives, and air quality standards and criteria, respectively for various air pollutants.

Ontario Regulation 419 (O.Reg.419) of the Ontario Environmental Protection Act (EPA) defines maximum concentration levels for various air contaminants at a Point of Impingement (POI), arising from an industrial facility or similar operation. The POI is generally defined as the off property location where the maximum concentration resulting from a facility emission occurs. However, if there is a child care facility, health care facility, senior's home or educational facility on the property in question these locations become the designated POI location.

Facility property boundaries are most often used as the POI. With the exception of the ventilation buildings assessed for Alternative 3, the emissions in this assessment are from open, public sources, and thus are not subject to MOE POI standards and criteria. The ventilation buildings will be assessed against the POI criteria during the analysis of the Technically and Environmentally Preferred Alternative in order to determine the necessary property footprint based on the design.

In addition, Section 14 of the Ontario Environmental Protection Act (EPA) prohibits a facility or operation to cause an adverse effect. The definition of "adverse effect" in the EPA includes:

1. impairment of the quality of the natural environment for any use that can be made of it; and,
2. loss of enjoyment of normal use of property.
O.Reg. 419 also sets desirable Ambient Air Quality Criteria (AAQCs) for various pollutants. The AAQCs are used to assess air quality and potential changes to it. The Standards Development Branch of the MOE publishes a set of guideline
limits in Summary of O.Reg. 419 Standards and Point of Impingement Guidelines \& Ambient Air Quality Criteria (AAQCs) [MOE, 2005].

Federal Air Quality Objectives encompass three levels of air quality objectives: maximum desirable level (MDL), maximum acceptable level (MAL) and maximum tolerable level (MTL). The MAL is intended to provide adequate protection against effects on soil, water, vegetation, materials, visibility, personal comfort and well-being. The MAL is considered to be a realistic objective. When the MAL is exceeded, control action by a regulatory agency is indicated. Table 2.3 summarizes the applicable available criteria from the MOE and Environment Canada.

TABLE 2.3 AIR QUALITY CRITERIA FOR PM 2.5 AND NO

| ntaminant | Averaging Time | MOE AAQC $\mu \mathrm{g} / \mathrm{m}^{3}(\mathrm{ppb})$ | Federal AQ Objective or Maximum Acceptable Level (MAL) $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ |
| :---: | :---: | :---: | :---: |
| $\begin{gathered} \mathrm{NO}_{\mathrm{x}} \\ \left(\text { as } \mathrm{NO}_{2}\right) \end{gathered}$ | 1 h | 400 (200) |  |
|  | 24h | 200 (100) |  |
|  | Annual |  | 1001 |
|  | 24 |  | 30 * |
| $\mathrm{NO}_{x}$ nin $\mathrm{PM}_{2.5}$ inclu respirable <br> ${ }^{1}$ MAL is for $\mathrm{NO}_{2}$ <br> - Indicates no criterion available |  |  |  |

Emissions of $\mathrm{NO}_{\mathrm{x}}$ and $\mathrm{PM}_{2.5}$ from the vehicles traveling on the freeway and the local service roads, other local arterial roadways, local industry and transboundary pollution from the southeastern United States have the greatest potential to impact local air quality. $\mathrm{NO}_{\mathrm{x}}$ is the sum of nitrogen dioxide $\left(\mathrm{NO}_{2}\right)$ plus nitric oxide (NO). At present, there is no annual AAQC for $\mathrm{NO}_{\mathrm{x}}$, but there is a MAL for $\mathrm{NO}_{2}$. The assessment was conservatively completed assuming that $100 \%$ of the $\mathrm{NO}_{x}$ is $\mathrm{NO}_{2}$. Typically, $\mathrm{NO}_{2}$ comprises approximately $60 \%$ of total $\mathrm{NO}_{\mathrm{x}}$. With respect to $\mathrm{PM}_{2.5}$, the MOE does not currently have an AAQC for $\mathrm{PM}_{2.5}$. Instead, they have adopted the Canada Wide Standard (CWS) for $\mathrm{PM}_{2.5}$, which is a Federal air quality objective that comes into force in 2010. Unlike the POI criteria in Ontario Regulation 419, it is not a legally enforceable standard that can be applied to specific sources. However, non-attainment of the CWS may indicate that regional action is required to reduce emissions.

## 2.3 <br> Existing Air Pollutant Concentrations

The Ontario Ministry of the Environment (MOE) measures air contaminants at various locations throughout Ontario, and reports on the state of Ontario's air quality on an annual basis. These reports are known as "Air Quality in Ontario" reports.

The existing air quality is greatly influenced by local and long range (crossborder) contaminants generated in upwind urban and industrial areas. The predominant wind directions in Windsor are from the west to southwest, which bring contaminants from the heavily industrialized areas of Detroit, nearby communities and beyond. Air quality impacts in the area are dominated by the substances that combine to produce smog or acid rain. This includes both $\mathrm{NO}_{\mathrm{x}}$ and $\mathrm{PM}_{2.5}$.

Figure 2.2 presents a breakdown of $\mathrm{PM}_{2.5}$ emissions in Southwestern Ontario in 2005.


### 2.3.1 $\mid$ Ambient Monitoring Data

The MOE has historically operated a number of ambient air monitoring stations in Windsor. However, in recent years the number of fully operational stations has been reduced to two. These stations are located at:

1) 467 University Ave. (Station \#060204 C);
2) College / South St. (Station \#060211R);

To assess the existing air pollutant concentrations in the area, monitoring data from these two stations were obtained from the MOE [MOE 2000-2005]. The MOE AAQCs are based on $\mathrm{NO}_{2}$ measurements rather than total NOx, thus the $\mathrm{NO}_{2}$ data has been presented. Tables 2.4 and 2.5 present a summary of the measurements for $\mathrm{NO}_{2}$ and $\mathrm{PM}_{2.5}$ respectively.

| Station ID | Station Location | Averaging Peri | Nitrogen Dioxide ( $\mu \mathrm{g} / \mathrm{m}^{3}$ ) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | AAQC | Year |  |  |  |  |
|  |  |  |  | 2001 | 2002 | 2003 | 2004 | 2005 |
| \#060211-R | College / South St. | Mean | - | 39 | 37 | $\mathrm{INS}^{+}$ | 33 | 32 |
|  |  | 90th Percentile | - | 66 | 62 | 69 | 62 | 62 |
|  |  | 1-Hour Maximum | 400 | 130 | 175 | 182 | 176 | 133 |
|  |  | 24-Hour Maximum | 200 | 83 | 116 | 92 | 79 | 109 |
| \#060204-C | 467 <br> University Ave. | Mean | - | 36 | 36 | INS | 34 | 32 |
|  |  | 90 ${ }^{\text {th }}$ Percentile | - | 62 | 60 | 73 | 68 | 62 |
|  |  | 1-Hour Maximum | 400 | 163 | 130 | 150 | 182 | 124 |
|  |  | 24-Hour Maximum | 200 | 77 | 86 | 94 | 90 | 100 |

${ }^{+}$INS $=$Insufficient data available to compute a representative average

Table 2.5-Five Year Summary of MOE Monitoring Results - PM 2.5
2.3.1.2 Existing Air Pollutant Concentrations in the Huron Church Rd/Hwy 3 Corridor

As part of the Environmental Assessment process, the DRIC team has established two ambient air monitoring stations in the study ACA, along the existing Huron Church/Talbot Rd. corridor. The purpose of the monitoring program is to collect data on the total pollutant concentrations that are routinely observed in the corridor, rather than specifically determine the fraction that originates from the roadway. This information will firmly establish the baseline air contaminant concentrations in the vicinity of the route. The monitoring program commenced in September 2006 and will continue until the end of September 2007.

The data are being used to:

- Establish current conditions within the corridor;
- Assist in determining background air concentrations of the pollutants being measured; and
- Benchmark the air dispersion modelling.

Table 2.6 presents a summary of the $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{\mathrm{x}}$ measurements collected from the two DRIC stations from October $1^{\text {st }}, 2006$ through December 31 $1^{\text {st }}, 2006$.

Table 2.6 Summary of DRIC $1^{\text {st }}$ Quarter Monitoring Results

| Pollutant | Averaging Time | Ontario Public Health Laboratory (OPHL) | St. Clair College (SCC) |
| :---: | :---: | :---: | :---: |
| $\begin{aligned} & \mathrm{NOx} \\ & (1-\mathrm{hr}) \end{aligned}$ | Max | 319 | 345 |
|  | Min | 0 | 0 |
|  | Average | $36 \square$ | 23 |
|  | 90th percentile | 71 | 47 |
| $\begin{aligned} & \mathrm{NOx} \\ & (24-\mathrm{hr}) \end{aligned}$ | Max | 144 | 149 |
|  | Min | 2 | 1 |
|  | Average | - 36 | 23 |
|  | 90th percentile | - 57 | 39 |
| $\begin{aligned} & \text { PM2.5 } \\ & \text { (24-hr) } \end{aligned}$ | Max | - 48 | 42 |
|  | Min | 8 | 8 |
|  | Average | 21 | 20 |
|  | 90th percentile | 32 | 29 |

Note that the values collected at the DRIC monitoring stations are somewhat higher than those collected at the MOE monitoring stations. This is expected since the DRIC monitoring stations are located within the corridor, whereas the MOE stations are not. Thus the MOE stations are not influenced by the same volumes of traffic.

## Contribution from Upwind / Background Sources

Air dispersion models provide an estimate of the air pollutant concentrations resulting from emission sources that are specifically included in the model set-up and inputs. Concentrations resulting from other, upwind sources are not included, but must be considered when assessing total expected air pollutant concentrations against relevant standards and guidelines. This is typically done by adding a "background component" to all model predicted results. The Ontario Ministry of the Environment (MOE) generally advocates the use of 90th percentile air pollutant concentrations, obtained from ambient air monitoring stations, for this purpose, as this is typically representative of background concentrations. The 90th percentile concentrations are then added to all model results.

Data on the existing air pollutant concentrations in the Windsor area were obtained from the two Ontario Ministry of the Environment (MOE) air monitoring stations located on College Street and on University Avenue. It is important to note that these stations are impacted by vehicle emissions from local
roadways (University Avenue in particular) which results in somewhat higher concentrations than would be seen if only background sources were captured. If these stations were located within the Huron Church Road / Highway 3 corridor, they would not reflect upwind background conditions, and corridor vehicle emissions would be "double counted" since they are being modelled. However, the MOE two stations are far enough away from the corridor to effectively remove "double counting" of the emissions.

The 24-hour 90th percentile $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{\mathrm{x}}$ concentrations measured at these stations in the past four years were conservatively selected as the interim background concentrations, which were added to all model predicted results.

As can be seen in Tables 2.4 and 2.5 above, the average $90^{\text {th }}$ percentile measured concentrations at the MOE stations are 69 and $73 \mathrm{ug} / \mathrm{m}^{3}$, and 23 and $21 \mathrm{ug} / \mathrm{m}^{3}$ for 1-hour $\mathrm{NO}_{\mathrm{x}}$ and $\mathrm{PM}_{2.5}$, respectively. The interim (first quarter) data from the two DRIC air monitoring stations were used in conjunction with the MOE monitoring data to select the background concentrations. As discussed above, since the DRIC monitors are located so close to the roadway (i.e. within the corridor), it is not appropriate to use the $90^{\text {th }}$ percentile values from these stations as a measure of comparison for upwind, background sources. The DRIC monitors are primarily being affected by emissions from the corridor, and thus use of the $90^{\text {th }}$ percentile values would effectively "double count" the emissions. Thus, the average measured concentrations were used. As shown in Table 2.6, the average measured concentrations at the DRIC stations for the first quarter of monitoring data (Oct $1-$ Dec $31^{\text {st }}$, 2006) are $20 \mu \mathrm{~g} / \mathrm{m}^{3}$ and $30 \mu \mathrm{~g} / \mathrm{m}^{3}$ for $\mathrm{PM}_{2.5}$ and NOx respectively. Thus, a value of $70 \mu \mathrm{~g} / \mathrm{m}^{3}$ was selected as the background $\mathrm{NO}_{\mathrm{x}}$ concentration (based primarily on the MOE data, and a value of $20 \mu \mathrm{~g} / \mathrm{m}^{3}$ was selected as the background $\mathrm{PM}_{2.5}$ concentration, based primarily on the DRIC monitoring data.

Table 2.7 presents a summary of the selected background concentrations.

Table 2.7 Summary of Background Concentrations Used in DRIC AQ Assessment

| Pollutant | Averaging Time |  |  |
| :--- | :---: | :---: | :---: |
|  | 1-hour | 24-hour | Annual |
| $\mathrm{NO}_{\mathrm{x}}$ | $70 \mu \mathrm{~g} / \mathrm{m}^{3}$ | $70 \mu \mathrm{~g} / \mathrm{m}^{3}$ | - |
| $\mathrm{PM}_{2.5}$ | - | $20 \mu \mathrm{~g} / \mathrm{m}^{3}$ | $9 \mu \mathrm{~g} / \mathrm{m}^{3}$ |

### 3.0 AIR DISPERSION Modelling

Atmospheric dispersion modelling is an essential step in the air quality assessment process as it is the only way to evaluate the impact of future changes in air pollutant emission sources. With respect to the Detroit River International Crossing project, these changes include implementation of alternatives, changes in fuels, vehicle technologies and traffic volumes.

Dispersion modelling is used to predict atmospheric concentrations of pollutants at specific receptors downwind of the source of pollutants over specific averaging times (i.e. annual, daily, hourly). The process involves using a computer model to mimic the way pollutants are emitted from sources, and how the atmosphere disperses them. The model takes emissions from a source, estimates how high into the atmosphere they will go, how widely they will spread and how far they will travel based on hourly meteorological data. The model then outputs the pattern of concentrations that will occur at receptors located downwind of the source for various averaging times.

In general, the maximum air pollutant concentrations (rather than average concentrations) that are predicted to occur over specific time periods at each receptor are typically used to assess the impact of worst case meteorological conditions. This usually occurs during periods with light wind speeds, when atmospheric dispersion is poor.

## $3.1 \quad$ Assessment Methodology

A large amount of data was required to complete the Air Quality Assessment in support of the evaluation of Practical Alternatives. This included data on existing air pollutant concentrations in the Windsor area, existing and future traffic volumes on the Huron Church Rd./Highway 3 corridor for each connecting route Alternative and Future No-Build scenarios, meteorological conditions in the Windsor area, and geographic information such as the location co-ordinates of roadways and sensitive receptors.

The necessary data was obtained from various sources, including other DRIC team members (i.e. traffic consultant, survey/mapping consultant), Environment Canada and the Ontario Ministry of the Environment (MOE).

The analysis was completed using the following approach:

1. Characterize Existing Environmental Conditions
a. Acquire Meteorological Data
b. Compile data on existing $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{\mathrm{x}}$ concentration
c. Determine background concentrations
2. Acquire data on current and future car and truck traffic volumes
a. Input to model - traffic data for existing and future conditions, including access road, plaza and crossing alternatives
3. Calculate pollutant emission factors for the highway corridor for existing and future conditions
a. Input to model - vehicle emissions for each road considered in the assessment, for both $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{\mathrm{x}}$
4. Use air dispersion modelling (CAL3QHCR) with meteorological data from Windsor Airport to determine future air pollutant concentrations in the vicinity of the corridor (essentially all of west Windsor) and at sensitive receptor locations (such as schools)

For the analysis of practical alternatives, an air dispersion model was set up for each of the alternative connecting routes, plazas, and crossings. The selected dispersion model was the CAL3QHCR model, which is specifically designed for roads and highways, and is approved for use in Ontario by the MOE. The model calculates emissions from moving vehicles differently from those that are queued and idling at intersections and customs plazas. The model also differentiates between "at grade", depressed and bridge sources.

The evaluation of Practical Alternative 3 required the assessment of tunnel ventilation buildings and emissions from the tunnel entrance and exit portals. The CAL3QHCR model is not appropriate for these emission sources, and thus another model was required. SENES evaluated the both the AERMOD and ISCST3 models for this purpose. While both models are appropriate to use in this assessment, the ISCST3 model was preferred since the same meteorological data file could be used for both models. Use of the AERMOD dispersion model would have required a different meteorological data file, which potentially could have introduced some inconsistencies since the outputs from both the CAL3QHCR and AERMOD/ISCST3 models were being combined. In order to avoid this potential problem, the ISCST3 air dispersion model was selected.

## 3.2

### 3.2.1

## Meteorological Data

Air dispersion models typically require the following inputs: hourly meteorological data, receptor locations, source characteristics, and emission rates.

The model results indicated that the meteorological data from 2003 resulted in the highest atmospheric concentrations for both contaminants evaluated $\left(\mathrm{NO}_{\mathrm{x}}\right.$ and $\mathrm{PM}_{2.5}$ ). Thus, the analysis for all alternatives was completed using this single year of data. The 2003 windrose overlain on the 5 -year wind rose (2000-2004) was presented in Figure 2.1. As can be seen in the figure, the 2003 windrose is similar to the 5 -year average, except that the 2003 wind speeds are lower in the quadrants from WSW to SSW, and slightly higher in the ENE quadrant. This is consistent with the model results (i.e. slightly higher predicted concentrations) since lower wind speeds results in poorer dispersion conditions.

### 3.2.2 <br> Receptors

A gridded network of receptors was created along the corridor at 100 m intervals that covered an area of 500 meters from the roadway on each side. In order to ensure that the worst-case effects were captured in the model results, several
grids with different receptor spacing were used within this area. The first two rows of receptors were placed at 50 m intervals from the ROW, followed by 100 m intervals up to 500 m away. Another grid with $500 \mathrm{~m} \times 500 \mathrm{~m}$ spacing was then overlaid to cover the rest of the modelling domain, which was essentially all of west Windsor. Any receptors that fell within the ROW were removed to prevent erroneous model results, as the models do not accurately predict air pollutant concentrations at locations on a source (i.e. on the roadway). Sensitive receptors (schools, churches, parks, etc.) were also identified and included in the model runs. A total of 2484 receptors were used in each model run completed for the analysis.

### 3.2.3


3.2.3.1

Roadway Segments Considered in the Assessment

The dispersion modeling analysis considered a large number of existing roads and roadway segments, in addition to new, or modified roads that will be constructed through implementation of the alternatives. These are as follows:

## Roads North of EC Row Expressway

Huron Church Road and all major intersecting roads along Huron Church were considered from the EC Row Expressway up to Riverside Drive. This includes
the existing Ambassador Plaza, and local roads in the immediate vicinity of the Plaza. The roads that were included in the assessment are listed below:

- Riverside Dr.
- University Ave.
- Wyandotte St.
- Patricia Rd. /

Union St. /
Sunset Ave.

- College Ave.
- Millen St.


In addition, all traffic on the Canadian side of the Ambassador Bridge and through the Ambassador Plaza was included in the assessment.

## Roads South of EC Row Expressway

Huron Church Road, Talbot Road/Highway 3 and all major intersections south of EC Row Expressway along the Huron Church / Highway 3 corridor were also included in the analysis. These are as follows:

Spring Garden Rd./
Labelle St.

- Lambton St./ Grand Marais Rd.
- Pulford St.
- Reddock Ave
- Todd Ln / Cabana

Rd.

- Huron Line
- Geraedt's Rd.
- Cousineau Rd. /


## Sandwich Pkwy

West

- Montgomery Dr.
- Surrey Dr.
- Grosvenor Rd.
- Howard Ave.
- Outer Dr.
- $6^{\text {th }}$ Concession
- Roseland Dr.
- Eastbourne Ave.
- North Talbot Rd.
- Tuson Way


## Roads in the Vicinity of Ojibway Parkway

The EC Row Expressway and Ojibway Parkway also formed part of the road network included in the assessment. A number of local roads in the vicinity of these major arteries were also assessed. They are as follows:

- EC Row
Expressway
- Ojibway Parkway
- Malden Rd.
- Matchette Rd.
- Broadway St. (E \& W)
- Chappus St.
- GN Booth Dr.
- Sandwich St.
- Prospect Ave.

In order to represent each roadway in the air dispersion model, the geographic co-ordinates of the first and last point of each roadway segment (which were often comprised of several links) for each traffic flow direction had to be coded into the model input files. This was done using ArcView GIS in combination with digital orthographic aerial photography and geo-referenced AutoCAD drawings of each alternative to manually select the start and end points of each of the over 700 roadway links included in the modeling. It is important to note that the roadway links for each connecting route alternative differed, due to variations in route alignments, locations of service roads, etc. Thus the coordinates for each connecting route alternative had to be coded manually for essentially all of the segments included in the models.

A map showing the network of existing roadways included in the analysis is shown in Figure 3.1.

Figure 3.1 - Modelled Road Network - Existing Roadways


\section*{| 3.2.3.2 | Traffic Volumes |
| :--- | :--- |}

Annual Average Daily Traffic (AADT) volumes for the roadways outlined above was provided by IBI Group, the DRIC team traffic consultant for existing, baseline conditions (2006) and the future no build cases for 2015, 2025 \& 2035. Traffic data was also provided for each connecting route, plaza and crossing alternative in each of these years, which reflects the anticipated changes resulting from implementation of the alternatives.

A selection of traffic volumes from the main routes considered in this assessment is presented below in Table 3.1 to illustrate the relative magnitude of the volumes. The full record of traffic data used in the assessment is presented in Appendix A These data form the basis of the emission calculations used in the dispersion modeling analysis.

Table 3.1 Summary of Traffic Volumes on Main Roads

| LOCATION | SECTION | SCENARIO | 24-HOUR AADT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 2006 |  | 2015 |  | 2025 |  | 2035 |  |
|  |  |  | CARS | TRUCKS | CARS | TRUCKS | CARS | TRUCKS | CARS | TRUCKS |
| Huron Church Rd / Talbot Road | North of ECR (Malden) | No Build | 46619 | 10495 | 51466 | 15109 | 50865 | 19582 | 50178 | 23384 |
|  |  | Alternatives | 0 | 0 | 58313 | 3352 | 60655 | 3876 | 63147 | 4592 |
|  | Grand Marais | No Build | 38142 | 10685 | 40771 | 15164 | 43485 | 18702 | 44116 | 22369 |
|  |  | Alternatives | 0 | 0 | 16732 | 245 | 18689 | 323 | 19884 | 351 |
|  | Todd/Cabbana | No Build | 33454 | 8049 | 35160 | 11484 | 37285 | 13728 | 38494 | 16010 |
|  |  | Alternatives | 0 | 0 | 15378 | 203 | 17269 | 227 | 18615 | 246 |
|  | Howard | No Build | 24217 | 6349 | 24229 | 9039 | 23549 | 11054 | 23159 | 13246 |
|  |  | Alternatives | 0 | 0 | 15282 | 21 | 16601 | 49 | 16979 | 73 |
| Hwy 401 Mainline | Todd/Cabbana to Grand Marais | Alternatives | 0 | 0 | 39481 | 11976 | 45994 | 16720 | 49632 | 20509 |

Daily profiles of car and truck traffic on different roadway types (i.e. highway, major arterial, local roads) were also provided, which were used to convert the AADTs into hourly volumes. These hourly volumes of domestic and international cars and trucks on each roadway segment were used to estimate emissions of $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{\mathrm{x}}$ from each source. Separate weekday and weekend traffic patterns were provided to SENES by IBI and used to represent actual expected traffic conditions. Idling traffic volumes and queue lengths were calculated by the CAL3QHCR air dispersion model based on the number of vehicles that approach an intersection, the signal timing and the capacity of each intersection. The vehicles approaching an intersection queue were conservatively assumed to be same as the free-flowing traffic volume.

### 3.2.3.3 $\mid$ Vehicle Emissions Estimates

Emissions from vehicles traveling on public roadways account for a significant portion of the smog producing air pollutants in North America. Although tailpipe emissions are the major source of gaseous pollutants (such as NOx) from these sources, they are not the major source of particulate emissions. In most cases, tailpipe emissions are a small fraction ( $<5 \%$ ) of the total particulate emissions from roadways during free-flow traffic conditions. As cars and trucks travel over the surface of a roadway, there are many sources that contribute to overall particulate emissions. These include road abrasion and degradation, tire \& brake wear, and soil/mud/debris that is deposited on the surface, in addition to tailpipe emissions. This is collectively known as surface resuspended particulate. When vehicles queue and idle, the particulate emissions are $100 \%$ from the tailpipe, as there are no emissions from the roadway surface if the vehicles are not moving. Idling cars emit approximately 4X more particulate than free-flowing cars, whereas idling diesel trucks emit over 25 X more particulate than free-flowing diesel trucks. However, they generally queue for shorter periods of time than they are free-flowing, unless the roadways are completely congested. However, the inclusion of queuing in the analysis is an important consideration.

Emission factors were developed separately for vehicle exhaust and surface roadway emissions (i.e. road dust) using Environment Canada’s MOBILE 6.2C model and USEPA emission factor methodologies (i.e.AP-42). Separate emission factors were developed for cars and trucks, and incorporate:

- regulatory changes in fuels and engine technologies;
- differences in Canadian and U.S. fuels and vehicles; and
- Canadian and U.S. fleet turnover rates.

Recent and on-going improvements in emission control technologies and fuels will combine to substantially reduce the emissions from transportation sources. As of June 2006, the maximum amount of sulphur in on-road diesel fuel was reduced from $500 \mathrm{mg} / \mathrm{kg}$ to $15 \mathrm{mg} / \mathrm{kg}$. These reductions were necessary for Canadian sulphur levels in on-road fuels to be consistent with U.S. levels, and to ensure that advanced emission control technologies on newer engines would be effective. In January 2007, additional engine standards for heavy-duty vehicles came into effect. These standards reduce NOx and particulate matter emissions by $60 \%$ and $90 \%$ respectively over existing levels, and require the incorporation of additional emission control technologies on these newer engines to effect these reductions.

Since the assessment spans a large area with a number of different types of roads, the development of the emission factors considered appropriate vehicle
speeds for each road type. Different emission factors were applied to each road based on the current or future assumed posted speed limits. The assessment also spans a long period of time, over which several regulated changes to fuel characteristics and vehicle engine technologies will occur. Although the effect of fuel changes on emissions occurs immediately following the implementation of the changes, technological changes require several years before the effects of the changes are fully observed. As such, the historical vehicle fleet turnover rates from the Detroit and Windsor areas were used to reflect the impacts of technological changes on vehicle emissions.

Table 3.2 presents a summary of the emission factors used in this assessment. Cars and trucks entering Canada from the U.S. were assumed to have U.S. vehicle and fuel characteristics, whereas cars and trucks exiting Canada were assumed to have Canadian vehicle and fuel characteristics. These assumptions are expected to adequately represent the fleet characteristics and emissions in the Windsor area, particularly on a daily basis, as some vehicles will both exit and enter on the same day. The complete database of emission factors, fleet turnover information and other assumptions used in the MOBILE6.2C model can be found in Appendix B. Sample calculations are presented in Appendix C.

Table 3.2 Summary of Emission Factors used in the Assessment

| Pollutant | Speed (km/h) | Surface Emissions (g/VKT) | Tailpipe Emission Factors (g/VKT) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Canadian Cars |  |  | Canadian Trucks |  |  | U.S. Cars |  |  | U.S. Trucks |  |  |
|  |  |  | 2015 | 2025 | 2035 | 2015 | 2025 | 2035 | 2015 | 2025 | 2035 | 2015 | 2025 | 2035 |
| $\mathrm{NO}_{\text {x }}$ | Idle* |  | 1.32 | 0.63 | 0.58 | 113.68 | 115.42 | 115.42 | 1.20 | 0.59 | 0.52 | 111.9 | 115.65 | 115.65 |
|  | 25 |  | 0.44 | 0.20 | 0.18 | 2.35 | 0.46 | 0.34 | 0.40 | 0.19 | 0.16 | 1.9 | 0.50 | 0.34 |
|  | 50 |  | 0.40 | 0.18 | 0.17 | 2.02 | 0.39 | 0.29 | 0.36 | 0.17 | 0.15 | 1.7 | 0.43 | 0.29 |
|  | 75 |  | 0.49 | 0.21 | 0.19 | 2.91 | 0.57 | 0.43 | 0.44 | 0.20 | 0.17 | 2.4 | 0.63 | 0.43 |
|  | 100 |  | 0.49 | 0.21 | 0.19 | 2.91 | 0.57 | 0.43 | 0.44 | 0.20 | 0.17 | 2.4 | 0.63 | 0.43 |
| $\mathbf{P M}_{2.5}$ | Idle* | 0 | 0.0086 | 0.0066 | 0.0065 | 1.0684 | 0.3140 | 0.1554 | 0.0086 | 0.0067 | 0.0065 | 1.1543 | 0.4342 | 0.1557 |
|  | 25 | 0.233* | 0.0021 | 0.0016 | 0.0016 | 0.0129 | 0.0062 | 0.0058 | 0.0021 | 0.0016 | 0.0016 | 0.0119 | 0.0063 | 0.0058 |
|  | 50 |  | 0.0021 | 0.0016 | 0.0016 | 0.0129 | 0.0062 | 0.0058 | 0.0021 | 0.0016 | 0.0016 | 0.0119 | 0.0063 | 0.0058 |
|  | 75 |  | 0.0021 | 0.0016 | 0.0016 | 0.0129 | 0.0062 | 0.0058 | 0.0021 | 0.0016 | 0.0016 | 0.0119 | 0.0063 | 0.0058 |
|  | 100 |  | 0.0021 | 0.0016 | 0.0016 | 0.0129 | 0.0062 | 0.0058 | 0.0021 | 0.0016 | 0.0016 | 0.0119 | 0.0063 | 0.0058 |

* $\mathrm{PM}_{2.5}$ surface emissions based on typical freeway link

In regards to traffic movements, the following additional assumptions were made:

- Vehicles on Highway 401 will be moving in a free-flowing state;
- Vehicles on service roads (and north of EC Row) will generally move in free-flow, but will queue at signalized intersections;
- Inbound vehicles at the customs plaza will queue at booths; and
- Outbound vehicles at the customs plaza will not queue.


### 3.2.3.4 $\mid$ Customs / Inspections Plazas

The traffic conditions at the customs plazas were modeled using the same queuing algorithm that was used for the intersections. Volumes of cars and trucks entering Canada from the U.S. as well those leaving Canada were provided to SENES by IBI for the years 2015, 2025, and 2035.

The amount of queuing at the plazas was estimated using the hourly traffic volume and the number of booths that are open during each hour, in addition to the average duration of each vehicle at a booth. The number of booths open in each hour was assumed to be a function of the traffic volume entering the plaza. Queues of cars and trucks form at car and truck booths respectively, and thus were modelled separately. Design information regarding plaza operations and vehicle timings were provided by Stantec.

With respect to plaza queuing, the following assumptions were used:

- Each truck requires 60 seconds at the customs booths
- Each car requires 45 seconds at the customs booths
- There is always queuing (idling) at the booth due to the one vehicle in the booth being inspected.
- Number of open booths assumed to be slightly less than capacity, such that some minimal queuing ( 2 or 3 cars or trucks) is always occurring at open booths.
- During periods where the capacity of the plaza is exceeded, longer queues form back towards the plaza entrance.

Groups of queue links were set up for each plaza car and truck lane based on an equal hourly distribution of free flow traffic through each booth that is open during a given hour. The groups extended back away from the booths to accommodate longer and longer queue lengths, as necessary. Each queue link was then manually "turned on" or "off" by calculating the number of vehicles queued at the open booths.

Based on the methodology and assumptions outlined above, and the inbound traffic volumes through the plaza provided by IBI, the maximum number of plaza booths open at any given time was 17 truck booths and 9 car booths at any of the new Customs/Inspection Plaza Alternatives.

The same methodology was applied to the Ambassador Plaza for the future nobuild scenarios and all of the connecting route alternatives. Using this approach, the queue lengths at the Ambassador often extended all the way back and onto

3.2.3.5

the Ambassador Bridge and Huron Church Road for the future no-build scenarios, which is what would be expected.
Tunnel Ventilation Buildings and Portal Emissions

The tunnel ventilation buildings are not a roadway source, and thus require the use of a different model. As described above, the ISCST3 model, which is used for assessing the impact of stationary emission sources such as industrial stacks, was used to model emissions from the tunnel entrance / exit portals and ventilation buildings. The conceptual design of the tunnel is based on the premise that emissions should not escape from the portals (i.e. exhaust flow is always greater than supply flow, such that air is continually drawn into the tunnel through the ramps and portals). However, there is a "piston effect" as cars drive out of the tunnel, which will result in some emissions from these areas. A total of $5 \%$ of the emissions were assumed to escape from the tunnel at these portal locations.

Based on the tunnel configuration, there are 10 locations where emissions may exit the tunnel. These are entrance/exit portals at on and off ramps, as well as two main entrance and egress locations (one at the approximate present terminus of Highway 401 [which is combined with an entrance portal] and one immediately west of the intersection of Huron Church Rd and EC Row Expressway). The main entrance and egress locations were assumed to be comprised of two separate tunnel "tubes". The $5 \%$ of the emissions that were assumed to escape from the portals were assumed to be evenly apportioned over these 10 locations. For the "Jet Fans" option, $100 \%$ of tunnel emissions were assumed to be emitted from these openings, and the emissions were evenly apportioned over the 10 locations.

As outlined earlier, there are three options for tunnel ventilation buildings (VBIA, VBIB, VBIC). Each of these has a slightly different conceptual design and thus each option was modelled to assess whether there are any differences in the potential affects to air quality. Mitigation options were not considered in this phase of the assessment.

The basic assumptions were as follows:

- The ventilation systems collect $95 \%$ of the total emissions from the tunnel
o All collected emissions were discharged from the vent stacks
o Vent building height is 18 m
o Stack height is 45 m (from the ground surface)
- Options VBIA \& VBIA have two ventilation buildings
o Emissions were apportioned equally between the two buildings
- Option VBIC has one ventilation building

The locations of each of the ventilation building options were presented earlier in Figure 1.3.

The ISCST3 model input files were completed and run for each of the tunnel ventilation scenarios. The hourly predicted concentrations from the vent buildings and portals were then added to the hourly predicted concentrations from the surface roadway sources (i.e. re-build Huron Church Road / Highway 3 corridor from the CAL3QHCR model) plus ambient background background concentrations to determine the total model predicted concentrations.

### 3.2.4 Model combinations



The work undertaken for this project required an assessment of local impacts, as well as an assessment of end-to-end solutions. The length of the model run times (i.e. computer time) and the number of possible combinations of connecting route, plaza and crossing alternatives would require an extraordinary amount of time effort to model each possible end-to-end combination. In addition, separate model runs are required for each pollutant $\left(\mathrm{PM}_{2.5}\right.$ and $\left.\mathrm{NO}_{\mathrm{x}}\right)$.

In order to complete all of the necessary model runs, the models were run in blocks of roadway/facility type. For each pollutant, separate runs were set up for each connecting route alternative, each plaza/crossing combination, and separate connections to the plazas from Highway 401. In addition, there are two alignment alternatives (Option $1 \&$ Option 2) for four of the connecting routes, and four tunnel ventilation options. Also, all model runs had to be completed for three horizon years (2015, 2025 \& 2035).

These model runs were completed on the same receptor network, and the results were output as hourly and/or daily values for the entire year of meteorology, at each receptor. The model results for each necessary combination of blocks were then added together to provide the hourly or daily maximum concentrations. A computer program was developed using the Linux operating system to overlay the necessary files. The combinations considered in this assessment are outlined below.

## Connecting Routes

- Future No-Build, Alternatives 1A (Opt $1 \& 2$ ), 1B (Opt $1 \& 2$ ), 2A (Opt $1 \& 2$ ), 2B (Opt $1 \& 2$ ), 3 (VB1A), 3 (VB1B), 3 (VB1C), 3 (jet fans) = 13 connecting route alternatives $\times 2$ pollutants $\times 3$ years $=78$ model runs


## Plazas \& Crossings

- Alternatives PA-A, PA-B, PA-C, PB-C, PB1-B, PB1-C, PC-C $=8$ combinations $\times 2$ pollutants $\times 3$ years $=48$ model runs


## Connections to Plazas

- Alternatives 1 A - PA, 1A - PB/C, 1B-PA, 1B-PB/C, 2A/2B-PA, 2A/2B$\mathrm{PB} / \mathrm{C}, 3-\mathrm{PA}, 3-\mathrm{PB} / \mathrm{C}=8$ alternatives $\times 2$ pollutants x 3 years $=48$ model runs

It should be noted that Huron Church Road north of EC Row Expressway and the Ambassador Bridge/Plaza were included in each model run for all of the connecting route alternatives.

As can be seen above, a total of 174 model runs were completed to evaluate the potential impacts of the proposed alternatives on air quality. If all end-to-end combinations were assessed, rather than the approach described above, almost double the number of model runs would have been required.

A model input file was prepared for each necessary run, as outlined above and run using one year of meteorological data (2003). The models were run on the Linux operating system, which offers more flexibility and memory in terms of processor use, file storage and manipulation of large data files. Data and file storage and management were significant issues in the completion of this project, since over 200 GB of numerical data were generated through the model runs alone.

Once the model runs were complete, the data was post-processed by adding the necessary data component results together (i.e. connecting route + connection to plazas + plaza/crossing) to form complete end-to-end results. The summed results were then imported into a GIS system for each combination such that the data could be interpreted in different areas along the connecting route, at various distances away from the ROW of each alternative.

## $4.0 \quad$ Overview of Model Results

As discussed earlier, air dispersion models calculate air pollutant concentrations at the receptor locations specified by the user in the model inputs. For this project two gridded networks of receptors were used along the roadway, as well as specific sensitive receptor locations. This chapter presents the results of the air dispersion modeling that was undertaken for each alternative.

The results from the No Build Alternative represent the predicted air quality conditions that will occur if no transportation improvements are undertaken in the corridor. Thus, all results have been presented in relation to this condition, such that the expected change in air quality (i.e. air pollutant concentrations) is apparent. Both worst case (maximum 24-hour) conditions and typical (annual average) conditions were evaluated.

For each pollutant and averaging time being evaluated, the magnitude of the maximum model predicted concentrations for each alternative and year are presented as percentages of the predicted concentrations for the No Build Alternative. Differences of less than $+/-10 \%$ (nominally $2-3 \mathrm{ug} / \mathrm{m}^{3}$ for $\mathrm{PM}_{2.5}$ and $15-30 \mathrm{ug} / \mathrm{m}^{3}$ for $\mathrm{NO}_{\mathrm{x}}$ ) were deemed to be within model tolerances and thus were considered to represent "no change" over No Build. Since the route alignments are Right of Way differed for many of the Alternatives, the results have been presented at defined distance intervals of $50 \mathrm{~m}, 100 \mathrm{~m}$ and 250 m from the ROW for comparative purposes. In many cases, this occurred at different model receptors for different Alternatives, since a receptor that was located 50 m from the ROW for one Alternative could have been within the ROW of another one.

In addition, where the concentrations exceed Federal or Provincial standards, objectives or guidelines, the change in the number of times the concentration was predicted to exceed (i.e. number of exceedances) was also reported, relative to the No Build Alternative. These measures were used to assess the potential impacts of any predicted changes to air quality.

Achievement of the Canada Wide Standard (CWS) is based on no more than 8 24 -hour periods with concentrations greater than $30 \mathrm{ug} / \mathrm{m}^{3}$. Thus, only results with greater than 8 exceedances were deemed to be in exceedance of the standard. In addition, the 8 day threshold was used to assess the significance of any changes in the number days predicted to be greater than $30 \mathrm{ug} / \mathrm{m} 3 \mathrm{in}$ comparison to No Build (i.e. if an Alternative had 9 exceedances less (or more)
than No Build, this difference was deemed to be significant, regardless of the total number of exceedance days). In addition, any exceedance of the annual criterion of $15 \mathrm{ug} / \mathrm{m}^{3}$ was deemed to be significant for the purpose of this assessment.

The results are presented separately for the Access Road alternatives, Customs/Inspection Plazas and Crossings.

## 4.1

## Access Road Alternatives

Tables 4.1 through 4.10 present the results of the air dispersion modelling for each of the connecting route alternatives. In order to compare microscale differences between the different alternatives, the results of each access road alternative will be presented and discussed in relation to specific areas along the route, starting east of the present Highway 401 terminus and ending at three potential river crossing locations. These are as follows:

- Highway 401/Highway 3 to Howard Avenue
- Howard Avenue to Cousineau Road
- Cousineau Road to Lennon Drain
- Lennon Drain to Pulford Street
- Pulford Street to Malden Road

All Access Road Alternatives commence at the existing Highway 401 terminus, and end at Malden Road south of EC Row Expressway.

The results are presented at increasing distances/offsets at $50 \mathrm{~m}, 100 \mathrm{~m}$ and 250 m from the ROW to provide an indication about how quickly the concentrations will decrease as you move away from the roadway.

The results presented below generally follow the expected trends based on the changes in the emission factors and increases in traffic volumes over time. The concentrations generally decrease as the distance from the roadway increases; the $\mathrm{PM}_{2.5}$ concentrations increase with time, as traffic volumes are predicted to increase from 2015 through 2035, and NOx concentrations decrease over time as the emission factors are going to be significantly reduced in the future, such that emissions are lower than 2015, regardless of predicted traffic growth.

It should be noted that the roadway and ramp alignments are essentially identical between Highway 401 and Howard Avenue for all non-tunnel alternatives. As a result, the maximum predicted concentrations and the changes in relation to No

Build are the same for these Alternatives, and thus any variations in the model predicted concentrations are likely due to slight differences in the forecasted traffic volumes for each alternative, in addition to some residual effect of emissions that occur in the previous segment. Therefore, the results will only be discussed in Section 4.1.1 for Alternative 1A. However, the results are applicable to Alternatives 1B, 2A and 2B for this area. The presence of the tunnel entrance and exit portals influences the concentrations in this area. The resulting differences will be discussed in Section 4.1.5 for Alternative 3 (tunneled access road).

As outlined previously, four separate tunnel ventilation options were examined. The results indicate that the location of the ventilation buildings does not have a significant affect; the locations of the entrance and exit portals have a higher impact on the results. The results of the "Jet Fans" tunnel ventilation option indicated that this option produced unacceptably high $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{\mathrm{x}}$ concentrations, and thus will not be discussed in detail in this report. Thus, the results will be discussed in the context of only one of the ventilation options (VB1A).

### 4.1.1



## Alternative 1A

As discussed previously, access road Alternative 1A is an at grade freeway with one-way service roads located on either side. The freeway is depressed where local arterial roads cross over it, such that these bridges are at-grade, rather than elevated.

The dispersion modeling results for Alternative 1A are presented in Tables 4.1 and 4.2 for $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{\mathrm{x}}$ respectively. The maximum predicted NOx concentrations are below the MOE AAQCs for both averaging periods (1-hour and 24-hour) at all locations along Alternative 1A.

Highway 401/Highway 3 to Howard Avenue

This segment represents the continuation of the existing Highway 401 alignment from Highway 3 to Howard Avenue. The emissions in segment are dominated by the freeway mainline and the on/off ramps for all alternatives. There is little difference between the non-tunnel Alternatives in this segment, as the ramp configurations do not change for any of the alternatives.

As can be see in the tables, the maximum predicted concentrations of $\mathrm{PM}_{2.5}$ are the same or less than the No Build option at 50 m away from the Right of Way, but are slightly higher than No Build at 100 and 250 m away. However, these differences are less than $10 \%$ and thus are considered to be the same as the No Build Alternative.

The annual average (rather than maximum) $\mathrm{PM}_{2.5}$ concentrations in this area are predicted to be the same or less than No Build at all distances from the ROW; but once again these differences are less than 10\%. There are slight reductions in the number of days predicted to exceed the CWS at 50 m away from the ROW in 2025 and 2035. However, these differences are less than 8 days, and thus are not considered to be different than the No Build Alternative. There is no change in the number of days predicted to be greater than the CWS at 100 m and 250 m away from the right of way in all horizon years.

As mentioned preyiously, all predicted NOx concentrations are below the relevant MOE AAQCs, and are below the No Build concentrations.

Overall, these results indicate a slight improvement in air quality in the area over the No Build Alternative (i.e. reduced NOx, lesser exceedances), but are generally very similar to No Build.

### 4.1.1.2

## Howard Avenue to Cousineau Road

This segment covers the area along the route between Cousineau Road and Howard Avenue. In this phase of the assessment, two separate alignment options (Option 1 and Option 2) were studied along the access road between Howard Avenue and St.Clair College. The first route alignment (Option 1), realigns the existing Talbot Road / Highway 3 corridor slightly to the northeast. This realignment begins at approximately at Howard Avenue and continues approximately to the entrance to St.Clair College.

The Option 2 alignment utilizes the existing Talbot Road / Highway 3 corridor as local access service roads without any realignment and aligns the freeway to the southeast.

The model results for each Option that was studied, for all horizon years are found in Tables 4.1 and 4.2, presented earlier.

The Tables illustrate that at 50 m from the ROW, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations are lower than the No Build Alternative in all horizon years, but are marginally higher at 100 and 250 m away. However, with the exception of

2035, the differences in these results are less than $10 \%$, and thus are not considered to be different than No-Build. Also, the number of days predicted to be greater than the CWS is less than No Build at 50 m away in 2035, and generally the same as No Build at 100 and 250 m away. Annual average (typical) concentrations are considered to be the same as No Build up to 250 m from the ROW.

The NOx results are somewhat different, as all predicted concentrations at all distances are less than No Build.

With respect to the differing alignments, there is no difference in the maximum predicted $\mathrm{PM}_{2.5}$ concentrations between Option 1 and Option 2 alignments. However, the Option 2 alignment results in an even greater reduction in the number of days predicted to exceed the CWS at 50 m from the ROW. Thus, the Option 2 alignment would be slightly preferred.

Overall, these results indicate a slight improvement in air quality in the area over the No Build Alternative (i.e. reduced NOx, lesser exceedances), but are generally very similar to No Build.

### 4.1.1.3

Cousineau Road to Lennon Drain

This segment represents the area between the Lennon Drain and Cousineau Road, and encompasses the St.Clair College area. For many Alternatives this includes freeway on and off ramps in addition to increased daytime/weekday traffic as staff and students enter the facility. As discussed earlier, two roadway alignment options were studied for part of this area.

For this Alternative, the results in this area are generally similar to those seen between Howard Avenue and Cousineau Road. With two exceptions, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations for this alternative are the same or marginally higher than the No Build Alternative at up to 250 m from the ROW. The concentrations are marginally higher than No-Build at 100 m from the ROW in 2015 and 2025.

Maximum predicted 1-hour NOx concentrations are less than the No Build Alternative at all horizon years at 50, 100 and 250 m from the ROW.

The differing Option 1 and Option 2 alignments result in similar maximum predicted $\mathrm{PM}_{2.5}$ concentrations, but the Option 2 alignment results in an even greater reduction in the number of days predicted to exceed the CWS at 50 m

### 4.1.1.4

### 4.1.1.5 Pulford Street to Malden Road

## Lennon Drain to Pulford Street

This area encompasses the access road from approximately Todd Lane/Cabana Road up to Pulford Street. The maximum predicted $\mathrm{PM}_{2.5}$ concentrations in this area are lower than the No Build Alternative at 50 m from the ROW, and are the same as No Build at 100 and 250 m away. There are no differences in the annual average concentrations at these distances in all years. However, at 50 m from the ROW there are significant reductions over No Build in the number of days predicted to exceed the CWS.

Maximum predicted NOx concentrations are below than the relevant MOE AAQCs, and are less than the No Build Alternative at all distances and horizon years studied.

These results indicate a noticeable improvement in overall air quality in the area through the implementation of Alternative 1A.

Due to its size and differences in the sources, this area was split into two separate sections for the AQ assessment. These are:

- Pulford St. - Labelle St. The air quality in this area is potentially affected by the new freeway and service roads.
- Labelle St. - Malden Rd. The air quality in this area is potentially affected by the presence of EC Row Expressway and the numerous on/off ramps to it and those between EC Row and the new freeway.

Thus, the results have been presented separately for each of the sections.


Also, in comparison to the Plaza B/C alignment, the Plaza A alignment results in fewer days that are predicted to exceed the CWS. Thus, in this area the Plaza A access road alignment is slightly preferred.

These results indicate a slight improvement in overall air quality in the area; however, in general there is very little difference between Alternative 1A and the No Build Alternative.

### 4.1.2 Alternative 1B

Access road Alternative 1B is very similar to Alternative 1A. One way service roads are located on either side of a below grade freeway. There are differences in the location of some of the on/off ramps between 1A and 1B, which results in some differing AQ effects in these areas.

The dispersion modeling results for Alternative 1B are presented in Tables 4.3 and 4.4 for $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{*}$ respectively. The maximum predicted NOx concentrations are below the MOE AAQCs for both averaging periods (1-hour and 24-hour) at all locations along Alternative 1B.
4.1.2.1

Howard Avenue to Cousineau Road

As mentioned previously, this segment covers the area along the route between Cousineau Road and Howard Avenue. The model results for each Option that was studied, for all horizon years are found in Tables 4.3 and 4.4, presented earlier.

The Tables illustrate that for the Option 1 alignment, at 50 m from the ROW the maximum predicted $\mathrm{PM}_{2.5}$ concentrations are lower than the No Build Alternative in all horizon years, but the same as No Build at 100 and 250 m away. However, these differences are only considered to be different than NoBuild in 2035 since these are the only changes greater than $10 \%$. Also, the number of days predicted to be greater than the CWS is less than No Build at 50 m away in all years, and the same as No Build at 100 and 250 m away. Annual average (typical) concentrations are considered to be the same as No Build up to 250 m from the ROW.

The maximum predicted 1-hour $\mathrm{NO}_{\mathrm{x}}$ concentrations presented in Table 4.4 are less than the No Build Alternative at all distances in all years, with one
exception. The change in concentration is less than $10 \%$ in 2015 at 250 m from the ROW, and thus it is considered to be the same as No Build. The maximum predicted 24-hour concentrations are less than No Build at up to 100 m from the ROW in all years studied.

In this area, there is a slight difference between the maximum predicted $\mathrm{PM}_{2.5}$ concentrations for the Option 1 and Option 2 alignments. The Option 2 alignment results in marginally lower concentrations and a greater reduction in the number of days predicted to exceed the CWS at 50 m from the ROW for all years studied. Also, the 1-hour NOx concentrations are also marginally lower with the Option 2 alignment in 2015 and 2025.
4.1.2.2

Cousineau Road to Lennon Drain

For this Alternative, the results presented in Tables 4.3 and 4.4 show that with one exception, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations are the same as the No Build Alternative at all distance intervals for all years. In the year 2035, the concentrations are different than No Build at 50 m from the ROW for both the Option 1 and Option 2 alignments. During this period, the Option 2 alignment results in a slight reduction in the number of days predicted to exceed the CWS.

Maximum predicted 1-hour NOx concentrations are less than the No Build Alternative at all horizon years at 50,100 and 250 m from the ROW for both alignment Options. The maximum predicted 24 -hour average concentrations with Option 1 are lower than No Build at 50 and 100 m away from the ROW in years 2025 and 2035, and additionally in 2015 with Option 2. Also, the maximum predicted 1-hour $\mathrm{NO}_{\mathrm{x}}$ concentrations are distinctly lower with the Option 2 alignment.

### 4.1.2.3 Lennon Drain to Pulford Street

This area encompasses the access road from approximately Todd Lane/Cabana Road up to Pulford Street. In 2015 and 2025, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations are lower than No Build at up to 100 m from the ROW. In 2035, the concentrations in this area are different than the No Build Alternative only at 50 m from the ROW and are the same as No Build at 100 and 250 m away. There are no differences in the annual average concentrations at these distances in all years. However, at 50 m from the ROW there are significant reductions over No Build in the number of days predicted to exceed the CWS.
4.1.2.4

## Pulford Street to Malden Road

As outlined earlier, this area was split into two separate sections for the AQ assessment. The results for each are discussed in the following sections.

### 4.1.2.4.1

4.1.2.4.2

## Pulford Street - Labelle Street

The Plaza A alignment results in no differences in the maximum and annual average predicted $\mathrm{PM}_{2.5}$ concentrations in this segment versus the No Build Alternative. All concentrations are within $10 \%$ of the No Build value and thus there is not considered to be any change. However, the Plaza B/C alignment shows reduced concentrations in comparison to No Build at 50 and 100 m away from the ROW, depending on the horizon year. Also, for this alignment, the number of days predicted to be greater than the CWS are much lower in 2035 at 50 m from the ROW than the No Build Alternative.

At 50 m from the ROW, the maximum predicted 1-hour and 24 -hour $\mathrm{NO}_{\mathrm{x}}$ concentrations for the Plaza A alignment are less than the No Build Alternative in all years. In 2025 and 2035, the maximum predicted 1-hour concentrations are less than No Build at 100 m from the ROW. However, the Plaza B alignment results in reduced 1-hour maximum NOx concentrations at all three distance intervals in all years studied, and maximum 24-hour concentrations that are less than No Build at 100 m from the ROW.

Labelle Street - Malden Rd.

Maximum predicted 1-hour NOx concentrations are below than the relevant MOE AAQCs, and are generally less than the No Build Alternative at all distances and horizon years studied. The maximum predicted 24-hour concentrations are less than No Build at distances up to 100 m from the ROW in 2015 and 2025, and up to 250 m away in 2035.

The results in this area show that, in general, there is no difference in the maximum predicted and annual average $\mathrm{PM}_{2.5}$ concentrations in comparison to the No Build Alternative. This is true for both the Plaza A and Plaza B/C alignments. However, in 2035 the Plaza B/C alignment results in a significant reduction in the number of days predicted to exceed the CWS at 50 m from the ROW.

For the Plaza A alignment, the predicted maximum 1-hour NOx concentrations are less than the No Build Alternative at all distances from the ROW, in all years. The 24-hour maximum concentrations are less than No Build up to 100 m from the ROW in 2025 and 2035. For the Plaza B/C alignment, the maximum predicted 1-hour NOx concentrations are less than No Build at 50, 100 and 250 m from the ROW in 2015, 2025 and 2035, respectively. The maximum 24-hour $\mathrm{NO}_{\mathrm{x}}$ concentrations for the Plaza B/C alignment are less than the No Build Alternative only at 50 m from the ROW in 2025 and 2035.

These results indicate a slight improvement in overall air quality in the area in terms of $\mathrm{NO}_{\mathrm{x}}$ concentrations; however, in general there is little difference between Alternative 1B and the No Build Alternative.

## Howard Avenue to Cousineau Road

The Tables illustrate that for the Option 1 alignment there is generally no difference between Alternative 2A and the No Build Alternative. This is because the differences in the maximum predicted $\mathrm{PM}_{2.5}$ concentrations are less than 10 \% between the Alternative and No Build. This is true at all distances, and all horizon years. However, at 50 m from the ROW the number of days predicted to be greater than the CWS is less than No Build in 2025 and 2035, and the same as No Build at 100 and 250 m away. Annual average (typical) concentrations are considered to be the same as No Build up to 250 m from the ROW.

The Option 2 alignment results in a reduction in the predicted maximum $\mathrm{PM}_{2.5}$ concentrations, and a further reduction in the number of days predicted to be greater than the CWS. These differences are significant at 50 m from the ROW in all years.

The maximum predicted 1-hour $\mathrm{NO}_{\mathrm{x}}$ concentrations are less than No Build at all distances, in all years, for both the Option 1 and Option 2 alignments. Predicted maximum 24-hour $\mathrm{NO}_{\mathrm{x}}$ concentrations are less than No Build at 100 m from the ROW for all horizon years. Also, the Option 2 alignment results in marginally lower predicted maximum 1-hour concentrations than the Option 1 alignment.
4.1.3.2

Cousineau Road to Lennon Drain

For this Alternative, the results presented in Tables 4.5 and 4.6 show that for the Option 1 alignment there is no difference in the maximum predicted $\mathrm{PM}_{2.5}$ concentrations between the Alternative and No Build at all distances and all horizon years. In addition, the number of days predicted to be above the CWS is only different than No Build at 50 m from the ROW in 2035. However, the Option 2 alignment results in marginally lower predicted maximum concentrations at 50 m from the ROW, and a reduction in the number of days predicted to exceed the CWS in 2025 and 2035.

Maximum predicted 1-hour NOx concentrations are less than the No Build Alternative for both alignment Options at all horizon years at up to 250 m from the ROW. Additionally, for the Option 2 alignment, the maximum predicted 24hour average concentrations are lower than No Build at 50 and 100 m away from the ROW in for all three years included in the study.

### 4.1.3.3

## Lennon Drain to Pulford Street

In this area, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations are lower than No Build at 50 m from the ROW in all horizon years. In addition, at 50 m from the ROW there are significant reductions in the number of days predicted to exceed the CWS. However, there are no differences in the annual average concentrations at these distances in all years.

Maximum predicted 1-hour NOx concentrations are below than the relevant MOE AAQCs, and are less than the No Build Alternative at distances up to 250 m from the ROW in all horizon years studied. Additionally, the maximum
predicted 24-hour concentrations are less than No Build up to 100 m from the ROW in all years.

### 4.1.3.4

4.1.3.4.1

### 4.1.3.4.2

Pulford Street - Labelle Street
The results in Tables 4.5 and 4.6 show that there are generally no differences between the Plaza A and Plaza B/C alignments. The predicted maximum $\mathrm{PM}_{2.5}$ concentrations are less than the No Build Alternative at 50 m from the ROW in all horizon years for both alignments. The differences in all other predicted maximum concentrations are within $10 \%$ of the No Build value and thus there is not considered to be any change. Also, both alignments show significant reductions in the number of days predicted to be greater than the CWS at 50 m from the ROW in all three years examined.

The Plaza A alignment results in lower maximum predicted 1-hour $\mathrm{NO}_{x}$ concentrations in comparison to No Build at 50 m from the ROW in 2015, and extending out to 100 m in 2025 and 2035. However, the Plaza B/C alignment results in further reductions in the 1 -hour predicted maximum $\mathrm{NO}_{x}$ concentrations at all three distance intervals studied, as well as reductions in the 24-hour maximum concentrations up to 100 m from the ROW.

## Labelle Street - Malden Rd.

The results in this area show that, at 50 m from the ROW the maximum predicted $\mathrm{PM}_{2.5}$ concentrations are less than the No Build Alternative for both the Plaza A and Plaza B/C alignments for all horizon years. However, there is generally no difference in the annual average $\mathrm{PM}_{2.5}$ concentrations in comparison to the No Build Alternative. Also, both alignments result in a significant reduction in the number of days predicted to exceed the CWS at 50 m from the ROW in all years, and up to 100 m from the ROW in 2035. The predicted concentrations and number of days greater than the CWS are similar for both alignments.

For the Plaza A alignment, the predicted maximum 1-hour $\mathrm{NO}_{\mathrm{x}}$ concentrations are less than the No Build Alternative at up to 250 m from the ROW in 2015,
and up to 100 m from the ROW in 2025 and 2035. However, with the exception of 2015 the Plaza B/C alignment shows reductions in the 1 -hour $\mathrm{NO}_{\mathrm{x}}$ concentrations at up to 250 m for all horizon years. For both alignments, the maximum predicted 24 -hour concentrations are similar to one another, and are less than No Build at 50 m from the ROW in all three years studied.

The alignment of Alternative 2B is almost identical to that of Alternative 2A. The primary difference is that 2 B is depressed (below grade) along the entire route from approximately Howard Avenue through to approximately Spring Garden Road.

The air dispersion modeling results for Alternative 2B are presented in Tables 4.7 and 4.8. The Tables show that all maximum predicted NOx concentrations are less than the relevant MOE criteria. Similar to the previous Alternatives, the results are discussed by geographical area in the following sections.

### 4.1.4.1

Howard Avenue to Cousineau Road

Tables 4.7 and 4.8 show that for both the Option 1 and Option 2 alignments, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations and the number of days predicted to exceed the CWS are less than the No Build Alternative at 50 m from the ROW in all horizon years. The annual average (typical) concentrations are considered to be the same as No Build up to 250 m from the ROW. Also, the results are generally similar for each Option.

The 1-hour predicted maximum $\mathrm{NO}_{\mathrm{x}}$ concentrations are less than No Build at all distances for all years, for both the Option 1 and Option 2 alignments. For both Options, the predicted maximum 24-hour $\mathrm{NO}_{\mathrm{x}}$ concentrations are less than No Build at up to 100 m from the ROW for all horizon years, and are generally similar. Also, the Option 2 alignment results in marginally lower predicted maximum 1-hour concentrations than the Option 1 alignment.

### 4.1.4.2 Cousineau Road to Lennon Drain

For this Alternative, the results presented in Tables 4.7 and 4.8 show that for the Option 1 alignment the maximum predicted $\mathrm{PM}_{2.5}$ concentrations are less than the No Build Alternative at 50 m from the ROW in all horizon years. This is also the
case for the Option 2 alignment in 2025 and 2035. In addition, the number of days predicted to be above the CWS is less than No Build at 50 m from the ROW in 2025 and 2035 for both Options. There are no differences in the annual average concentrations at these distances in all years.

Maximum predicted 1-hour NOx concentrations are less than the No Build Alternative for both alignment Options in all horizon years at 50, 100 and 250 m from the ROW. The maximum predicted 24 -hour concentrations are less than No Build at up to 100 m from the ROW in all years. The concentrations for both averaging periods are generally similar for each Option.

Lennon Drain to Pulford Street
In this area, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations are lower than No Build at 50 m from the ROW in all horizon years. In addition, at 50 m from the ROW there are significant reductions in the number of days predicted to exceed the CWS in 2015 and 2025; these reductions extend up to 100 m from the ROW in 2035. However, there are no differences in the annual average concentrations at these distances in all years.

The maximum predicted 1-hour $\mathrm{NO}_{\mathrm{x}}$ concentrations are less than the No Build Alternative at all distances and horizon years. The maximum predicted 24 -hour concentrations are less than No Build at distances up to 100 m from the ROW in 2015 and 2025, extending up to 250 m in 2035.
4.1.4.4
4.1.4.4.1

## Pulford Street - Labelle Street

The results in Tables 4.7 and 4.8 show that there are generally no differences between the Plaza A and Plaza B/C alignments. There are no differences in the predicted maximum and annual average $\mathrm{PM}_{2.5}$ concentrations between Alternative 2B and No Build in this area, as all predicted maximum concentrations are within $10 \%$ of the No Build value. However, both alignments show significant reductions in the number of days predicted to be greater than the CWS at 50 m from the ROW in all three years examined.

In 2015, the Plaza A alignment results in lower predicted maximum 1-hour NOx concentrations in comparison to No Build at 50 m from the ROW. This extends up to 100 m from the ROW in 2025 and 2035. The Plaza B/C alignment results in further reductions in the maximum predicted 1-hour $\mathrm{NO}_{x}$ concentrations in all years and at all distance intervals examined. Additionally, maximum predicted

24-hour NOx concentrations are less than the No Build Alternative at 100 m from the ROW in all years.
4.1.4.4.2

Labelle Street - Malden Rd.

The results in this area quite similar to the previous area (i.e. Pulford to Labelle) and show that in comparison to No Build there are generally no differences in the maximum predicted and annual average $\mathrm{PM}_{2.5}$ concentrations for Alternative 2B. Also, the results are similar for both the Plaza A and Plaza B/C alignments. However, both alignments show significant reductions in the number of days predicted to be greater than the CWS at 50 m from the ROW in 2015 and 2025, and up to 100 m from the ROW in 2035.

As seen in Table 4.8, with one exception the Plaza A alignment results in lower predicted maximum 1-hour $\mathrm{NO}_{\mathrm{x}}$ concentrations in comparison to the No Build Alternative in all years and at all distance intervals examined. In 2025 at 250 m from the ROW the change in concentration is less than $10 \%$ and thus is not considered to be different than No Build. This is similar to the results for the Plaza B/C alignment. For both alignments, the maximum predicted 24-hour concentrations are similar to one another, and are less than No Build at 50 m from the ROW in all three years studied.

### 4.1.5 Alternative 3

As discussed previously, access road Alternative 3 is a tunneled freeway with two-way service roads located at grade above the tunnel, along the approximate existing Huron Church Road / Highway 3 Right of Way. A number of entrance and egress portals are located at specific points along the access road to allow traffic to move from the service roads into the tunnel, and vice versa. Also, there are two main portals where the freeway sections enter/exit the ground. These are located approximately at Howard Avenue as well as at the E.C. Row Expressway.

The dispersion modeling results for Alternative 3 (based on ventilation option VB1A) are presented in Tables 4.9 and 4.10 for $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{\mathrm{x}}$ respectively. The maximum predicted NOx concentrations are below the MOE AAQCs for both averaging periods (1-hour and 24-hour) at all locations along Alternative 3.

### 4.1.5.1 $\mid$ Highway 401/Highway 3 to Howard Avenue

As discussed previously, this segment represents the continuation of the existing Highway 401 alignment from Highway 3 to Howard Avenue. The emissions in segment are generally dominated by the freeway mainline and the on/off ramps for all alternatives, and there is little difference in the results between the nontunnel Alternatives in this segment. All Alternative 3 options are somewhat different than the other Alternatives in that there are mainline entrance and egress points from the tunnel, which result in emissions from these portals. However, the effect of these portals is only significant for the Alternative 3 - Jet Fans option. With any of the Alternative 3-Fent Building options, the effect of these portals are seen only in very close proximity to the roadway, and do not affect areas outside of the ROW. As a result, this area is discussed separately for Alternative 3.

As can be see in the tables, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations are the same or greater than the No Build option at all distance intervals and all horizon years. However, the changes in concentration are less than $10 \%$ at 50 and 100 m away from the Right of Way, and thus are considered to be the same as No Build. The changes are greater than $10 \%$ at 250 m from the ROW, which shows the effect of the ventilation buildings which are located further to the west along the access road.

With one exception, the annual average (rather than maximum) $\mathrm{PM}_{2.5}$ concentrations in this area are predicted to be the same as No Build at all distances from the ROW. There are no differences in the number of days predicted to exceed the CWS.

As mentioned previously, all predicted NOx concentrations are below the relevant MOE AAQCs. In 2015, the maximum predicted 1-hour NOx concentrations are the same as No Build at 50 m away from the ROW, but are greater than No Build at 100 and 250 m away from the ROW. In 2025, the concentrations are less than No Build at up to 100 m from the ROW, which increases to 250 m from the ROW in 2035.

### 4.1.5.2 Howard Avenue to Cousineau Road

This segment covers the area along the route between Cousineau Road and Howard Avenue. In all previous Alternatives, two separate alignment options (Option 1 and Option 2) were studied along the access road between Howard

Avenue and St.Clair College. Only one alignment has been proposed for Alternative 3, and thus results for this area will not have Option 1 and Option 2.

The Tables illustrate that with one exception, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations are lower than the No Build Alternative at all distance intervals. The only exception is in 2035 at 250 m from the ROW, where the change in comparison to No Build is less than $10 \%$. Also, the number of days predicted to be greater than the CWS is less than No Build at 50 m from the ROW in all years. Annual average (typical) concentrations are also less than No Build at 50 m from the ROW in 2015, and up to 100 m from the ROW in 2025 and 2035. However, the change in annual average concentrations is less than $10 \%$ at 250 m from the ROW, and thus is considered to be the same as No Build. Thus, the effect is localized to within 100 m of the ROW.

The NOx results are somewhat different. With two exceptions, the maximum predicted 1-hour concentrations are less than No Build at all distances and in all years. The exceptions are at 250 m from the ROW in 2015 and 2025. In 2015, the predicted concentration is greater than No Build, and in 2025 the change is less than $10 \%$, and thus is not considered to be different. The maximum predicted 24-hour concentrations are less than No Build at up to 100 m from the ROW in all three years.
4.1.5.3

Cousineau Road to Lennon Drain

This segment represents the area between the Lennon Drain and Cousineau Road, and encompasses the St.Clair College area. For Alternative 3 this includes tunnel on and off ramps (entrance and exit portals) to allow freeway access for staff and students of the College.

Table 4.9 shows that in 2015 only the maximum predicted $\mathrm{PM}_{2.5}$ concentration at 50 m from the ROW is different (lower) than No Build. Maximum predicted concentrations at 100 and 250 m away, as well as annual average concentrations at all distance intervals are the same as No Build. In 2025 and 2035, the maximum predicted 24 -hour concentrations are less than No Build at all distance intervals, and the annual average concentrations are less than No Build at distances up to 100 m from the ROW. Also, the number of days predicted to be greater than the CWS are less than the No Build Alternative at 50 m from the ROW in 2025 and 2035.

Table 4.10 shows that with one exception, the maximum predicted 1-hour NOx concentrations are less than the No Build Alternative in all horizon years at all three distance intervals examined. The change is less than $10 \%$ at 50 m from the ROW in 2015, and thus it is considered to be the same as No Build. Maximum predicted 24-hour NOx concentrations are less than No Build at 100 m from the ROW in 2025 and 2035, but the same as No Build at 50 m and 250 m away. This is likely related to the effect of the tunnel ventilation buildings.

### 4.1.5.4

Lennon Drain to Pulford Street
The maximum predicted $\mathrm{PM}_{2.5}$ concentrations in this area are lower than the No Build Alternative at up to 100 m from the ROW, in all years included in the study. In 2025 and 2035, the effect extends out to 250 m . Also, at 50 m from the ROW there are significant reductions over No Build in the number of days predicted to exceed the CWS. This extends out to 100 m from the ROW in 2035. The annual average concentrations are less than the No Build Alternative at up to 100 m away from the ROW. There are no differences in the annual average concentrations at 250 m away from the ROW in all years.

Maximum predicted 1-hour NOx concentrations are below than the relevant MOE AAQCs, and are less than the No Build Alternative at all distances and horizon years studied. The maximum predicted 24-hour NOx concentrations are less than No Build at distances up to 100 m from the ROW in 2015 and 2025, and extends out to 250 m in 2035.

## Pulford Street to Malden Road

### 4.1.5.5.1

This area was split into two separate sections for the AQ assessment, with separate discussions for different Plaza alignments, as discussed previously.

## Pulford Street - Labelle Street

Table 4.9 shows that for the Plaza A alignment, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations in this area are different (less) than the No Build Alternative at all distance intervals and all years studied. The annual average $\mathrm{PM}_{2.5}$ concentrations for this alignment are also different than No Build at 50, 250 and 100 m from the ROW in 2105, 2025 and 2035, respectively. Also, the number of days predicted to be greater than the CWS are significantly less than the No Build Alternative in this area at 50 m from the ROW in all years, and extends out to 100 m from the ROW in 2035. However, there is no difference in the annual
average concentrations and number of days above the CWS at 250 m away. The results for Plaza B/C alignment are very similar to those of the Plaza A alignment.

Table 4.10 shows that with one exception the maximum predicted 1-hour $\mathrm{NO}_{\mathrm{x}}$ concentrations for this Alternative are less than the No Build Alternative in all years and at all distance intervals examined. The maximum predicted 24-hour concentrations are less than No Build at 50 m from the ROW. There is essentially no difference in the results between the Plaza A alignment and the Plaza B/C alignment.
4.1.5.5.2

Labelle Street - Malden Rd.

The model results in this area indicate that in general there are no differences in the maximum predicted and annual average $\mathrm{PM}_{2.5}$ concentrations in comparison to the No Build Alternative. In 2025 and 2035, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations are greater than No Build for the Plaza B/C alignment at 250 m and 100 m from the ROW respectively.

For the Plaza A alignment there is a significant reduction in the number of days predicted to exceed the CWS at 50 and 100 m from the ROW in 2025 and 2035 respectively. The Plaza B alignment results in a reduction in the number of these days at 50 m from the ROW in all years.

Predicted maximum 1 hour NOx concentrations are less than the No Build Alternative at up to 100 m from the ROW for the Plaza A alignment in all years, whereas this is the case for the Plaza B/C alignment only in 2025 and 2035. For both plaza alignments studied, the maximum predicted 24 -hour concentrations are less than No Build at 50 m from the ROW in all years included in the study.

### 4.2 Customs / Inspection Plaza Alternatives

As discussed previously, three separate alternatives were studied for Customs / Inspection Plaza alternatives. These are Plaza A, Plaza B / B1 and Plaza C. Tables 4.11 and 4.12 present the results of the air dispersion modelling $\left(\mathrm{PM}_{2.5}\right.$ and $\mathrm{NO}_{\mathrm{x}}$ ) for each of these Alternatives. In order to compare the location specific differences between the different alternatives, the results of each plaza alternative will be presented and discussed in relation to specific areas in the vicinity of each facility.

The plaza results show that the maximum predicted concentrations of $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{\mathrm{x}}$ are generally much higher in comparison to the access road alternatives. This is due to the longer idling time near the plazas as vehicles queue in line at the booths. Although the traffic data is similar for all Plaza alternatives, the footprints of the plaza properties, alignment of the plazas and proximity of nearby roads plays an important role in the maximum predicted concentrations, which is reflected in the differences in the resulting data.


The Plaza A Alternative is located adjacent to E.C. Row Expressway in the vicinity of Spring Garden Road / Armanda Street, and is the farthest from the Detroit River of any of the Alternatives under consideration. Plaza A provides potential access to all of the Crossing Alternatives (A, B or C) that are included in the study.

As can be seen in the Table 4.11, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations increase significantly within 100 m of the Plaza A boundary, in comparison to the No Build Alternative. The increase is a maximum of $250 \%$ at 50 m away from the property line of the facility, and is $136 \%$ at 250 m from the boundary in 2035. In addition, the number of days exceeding the CWS are also predicted to increase significantly at distances up to 100 m from the plaza boundary in 2035. The annual average concentrations also increase in comparison to No Build, but remain below the $15 \mathrm{ug} / \mathrm{m}^{3}$ criterion at 100 m away in 2035.

Similar to the $\mathrm{PM}_{2.5}$ results, the maximum predicted 1-hour NOx concentrations shown in Table 4.12 also increase significantly within 250 m of the plaza boundary. The maximum predicted concentrations exceed the MOE 1-hour $\mathrm{NO}_{\mathrm{x}}$ criterion on occasion in the immediate vicinity of Plaza A, and are more than 8 X higher than the predicted No Build concentrations at 50 m away in 2025 and 2035, and more than 4 X higher at 250 m away in 2035. However, the change in number of times that the MOE AAQC is predicted to be exceeded is not significant (i.e. $<8$ hours) beyond 50 m away.

Based on the results presented above, air quality is predicted to be generally poorer within approximately 100 m of the Plaza A boundary.

\section*{| 4.2.2 | Plaza B |
| :--- | :--- |}

The Plaza B alternatives are located in an industrial area immediately north of Broadway Street, west of Ojibway Parkway, near the Detroit River.

Plazas B and B1 are only slight variants of one another, and thus will be discussed in the same section. Due to the required elevation of the Crossing Alternatives and maximum grade allowances on the approach to the crossing, Plaza B could not provide access to Crossing B. Thus, the Plaza B1 variant was created to permit access to Crossing Alternative B.

### 4.2.2.1

## Plaza B1

Plaza B1 is located immediately to the west of Ojibway Parkway, and leads to Crossing Alternative B. The results shown in the Tables indicate a general worsening of air quality in the immediate vicinity of the Plaza. In addition, the nearby concentrations are affected by traffic on the E.C. Row interchange.

Within 250 m of the property boundary, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations increase significantly in comparison to the No Build Alternative. This increase ranges from 2.8 to 3.8 X higher relative to No Build at 50 m away, and approximately 1.3 to 1.4 X at 250 m away. In addition, the change in the number of days predicted to exceed the CWS is significant within 250 m of the plaza boundary in 2025 and 2035. Annual average $\mathrm{PM}_{2.5}$ concentrations are also higher compared to No Build, but are below the $15 \mathrm{ug} / \mathrm{m} 3$ criterion beyond 50 m away in 2015 and 2025, and beyond 100 m in 2035.

Table 4.12 presented the maximum predicted 1-hour NOx concentrations. The Table shows that the predicted concentrations are significantly greater than No Build within 250 m of the Plaza boundary. At worst, the maximum predicted concentrations are approximately 8 X higher than the predicted No Build concentrations at 50 m away in 2025, and approximately 3 X higher at 250 m away in 2035. Although the maximum predicted concentrations exceed the MOE 1-hour NOx criterion on occasion in the immediate vicinity of Plaza B1, the change in number of times is not significant (i.e. $<8$ hours) at any of the distance intervals and in any of the horizon years studied.

Based on the results presented above, a general worsening of air quality is expected within approximately 250 m of the Plaza B1 boundary. However, the highest impacts will likely occur within 50-100 m of the boundary.

### 4.2.2.2 $\quad$ Plaza B

Plaza B is located adjacent to Plaza B1, slightly farther to the west and closer to the Detroit River. Only Crossing Alternative C can be accessed from this Plaza Alternative.

Table 4.11 shows that the maximum predicted $\mathrm{PM}_{2.5}$ concentrations are significantly higher than the No Build Alternative within 250 m of the Plaza B property boundary. This increase is almost 3 X higher compared to No Build at 50 m away, and is approximately 1.5 X at 250 m away in 2035. Also, the number of days predicted to exceed the CWS increases significantly over the No Build Alternative within 250 m of the plaza boundary in 2035. In addition, annual average $\mathrm{PM}_{2.5}$ concentrations are higher compared to No Build, but are below the $15 \mathrm{ug} / \mathrm{m}^{3}$ criterion beyond 50 m from the Plaza B boundary in all three horizon years.

The maximum predicted 1-hour NOx concentrations shown in Table 4.12 are also significantly higher in comparison to the No Build Alternative within 250 m of the plaza boundary. This is true in all years that were examined. The maximum predicted concentrations exceed the MOE 1-hour NOx criterion on occasion at distances up to 100 m from the Plaza in all years, but the change in number of exceedances is only significant at 50 m away in 2025 and 2035. The change in the predicted maximum concentrations are approximately 10 X in comparison to No Build at 50 m away in 2025 and 2035, and approximately 3.5 X at 250 m away in 2035.

These results indicate that air quality is predicted to decrease within approximately 250 m from the Plaza B property boundary by 2035. The highest impacts will likely occur within 50-100 m of the boundary.

### 4.2.3

## Plaza C

The Plaza C Alternative is located in an industrial area in the vicinity of the Brighton Beach Generating Station, on the approximate footprint of the transformer station. Plaza C provides access to Crossing Alternative C only.

Similar to the $\mathrm{PM}_{2.5}$ results for the other Plaza alternatives, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations increase significantly over No Build at distances up to 250 m from the Plaza C boundary. This change relative to the No Build Alternative is a maximum of 2 X at 50 m away in 2015, and is approximately 2.1 -2.2 X at 250 m away in 2025 and 2035. Also, the change in the number of times that the CWS is predicted to be exceeded (relative to No Build) is
significant at distances up to 250 m away in 2035. The annual average $\mathrm{PM}_{2.5}$ concentration exceeds the $15 \mathrm{ug} / \mathrm{m} 3$ criterion at 50 m from the boundary in all horizon years.

The predicted maximum 1-hour $\mathrm{NO}_{\mathrm{x}}$ concentration at 50 m from the Plaza C property boundary ranges from $123 \%$ to $222 \%$. At 250 m away, this range is from $116 \%$ to $176 \%$. Although these increases are significant at all distances relative to No Build (i.e. > $10 \%$ change), the MOE AAQC is not exceeded at any distance interval, in any of the horizon years.

As can be seen above, the overall magnitude of the changes in maximum $\mathrm{NO}_{\mathrm{x}}$ and $\mathrm{PM}_{2.5}$ concentrations is generally less for the Plaza C Alternative than for any of the other Plaza Alternatives evaluated. This is due to the Plaza alignment and arrangement of roadways within the property. There is a larger buffer between the traveled portion of the roadways within Plaza C and the property boundary. As a result, the emissions have dispersed more by the time they reach the property boundary.

These results indicate a marginal worsening of air quality within approximately 250 m from the Plaza C property boundary. However, the most significant affects will likely occur within $50-100 \mathrm{~m}$ away.

## 4.3

Crossing Alternatives

As outlined earlier in the report, three separate bridge crossing alternatives were studied and evaluated as part of this project. These are:

- Crossing A
- Crossing B
- Crossing C

Also, there is a connecting roadway between the exit of each plaza and the entrance to the Crossings.

The air dispersion modeling results for all Crossing Alternatives are presented in Tables 4.13 through 4.14. In order to compare the location specific differences between the different alternatives, the results of each crossing alternative will be presented and discussed in relation to specific areas in the vicinity of each bridge and connecting roadway.

The results for the crossings indicate that the maximum predicted concentrations of $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{\mathrm{x}}$ are generally similar to those of the access road alternatives. However, for some Plaza / Crossing combinations there is some "spillover" of idle emissions from the Plaza, due to the proximity of the Plaza to the Crossing. This is the case for the Plaza B / Crossing B and Plaza C / Crossing C combinations.
4.3.1 $\quad$ Crossing A

Crossing Alternative A can be accessed from Plaza A only, and is located in the vicinity of Wright and Water Streets. It has the longest span of the three Alternatives studied, at 1.1 km .

As can be seen in the Table 4.13, the change in the maximum predicted $\mathrm{PM}_{2.5}$ concentrations at 50 m from the crossing / connecting roadway is $150 \%$ in comparison to the No Build Alternative in 2015. This drops to $127 \%$ at 250 m away. In 2035, the increases are $172 \%, 150 \%$ and $135 \%$ at 50,100 and 250 m , respectively. In addition, the number of days predicted to be in excess of the CWS increase significantly at distances up to 100 m from the ROW in all three horizon years.

The annual average concentrations are predicted to marginally increase in the vicinity of the crossing, and will exceed the criterion within 50 m in 2025 and 2035.

The changes in the maximum predicted 1-hour NOx concentrations shown in Table 4.14 are at maximum a two-fold increase over No Build, which occurs at 50 m away from the Crossing / connecting roadway in 2015. The increases are less than this at all other distances and all horizon years. Also, there are no exceedances of the MOE 1-hour $\mathrm{NO}_{\mathrm{x}}$ criterion in the vicinity of the crossing and connecting roadway.

Based on these results, a marginal decrease in air quality is predicted to occur at distance up to 100 m away from Crossing A and the associated connecting roadway.

### 4.3.2 <br> Crossing B

Crossing Alternative B can be accessed from Plaza A or Plaza B1. Crossing B is located adjacent to the Brighton Beach Power Station and has a span of approximately 800 m .

Table 4.13 shows that for the Plaza A / Crossing B combination, the change in the maximum predicted $\mathrm{PM}_{2.5}$ concentrations in comparison to the No Build Alternative ranges from $150-172 \%$ at 50 m away, 135 - $150 \%$ at 100 m and $127-135 \%$ at 250 m . Also, the number of days predicted to be in excess of the CWS is significantly higher than No Build at distances up to 100 m from the ROW in all horizon years.

The results for the Plaza B1 / Crossing B combination are somewhat different, due to the "spillover" effect mentioned previously. For this combination, the biggest change in the maximum predicted $\mathrm{PM}_{2.5}$ concentrations at 50, 100 and 250 m from the crossing / connecting roadway is $212 \%$, $188 \%$ and $148 \%$ respectively, in comparison to the No Build Alternative. This occurs in 2035. The number of days predicted to be in excess of the CWS is marginally higher for this combination, with significant increases at distances up to 100 m from the ROW in all years, and up to 250 m in 2035. These differences occur because Plaza B1 is located relatively close to Crossing B, and the emissions from the Plaza and the connecting roadway / crossing combine.

The annual average concentrations are the same as No Build at all distance intervals in 2015, but are higher than No Build at 50 m from the ROW in 2025 and 2035 for both Plaza / Crossing combinations.

The changes in the maximum predicted 1-hour NOx concentrations shown in Table 4.14 are greater than $10 \%$ at up to 250 m from the ROW in all horizon years. These changes are as much as $174 \%$ of the No Build at 50 m away, and up to $147 \%$ at 250 m for the Plaza A / Crossing B1 combination. However, there are no predicted exceedances of the MOE 1-hour $\mathrm{NO}_{\mathrm{x}}$ criterion in the vicinity of the crossing and connecting roadway. For the Plaza B1 / Crossing B combination, the increases range from $175 \%$ to $219 \%$ at 50 m away, with significant differences in the maximum predicted 1-hour NOx concentrations at distances up to 250 m from the Crossing / connecting roadway. Once again, there are no exceedances of the NOx criterion at any of the distance intervals studied, in any of the three horizon years.

Based on the above, air quality is predicted to decrease within 100 m of Crossing $B$ and or the associated connecting roadway.

\section*{| 4.3.3 | Crossing C |
| :--- | :--- |}

Crossing Alternative C can be accessed from Plaza A, Plaza B or Plaza C. It is located near Stirling Marine Fuels, and has the shortest span of the three Crossing Alternatives, at approximately 700 m .

Table 4.13 shows that for the Plaza A / Crossing C combination, the change in the maximum predicted $\mathrm{PM}_{2.5}$ concentrations is as much as $172 \%$ at 50 m away from the crossing / connecting roadway in comparison to the No Build Alternative in all horizon years. The increase in concentrations is a maximum of $135 \%$ at 250 m away. Also, the number of days predicted to be in excess of the CWS is significantly higher than No Build at distances up to 100 m from the ROW in all years studied.

The Plaza B / Crossing C combination results are similar to the previous combination. The highest increase in the maximum predicted $\mathrm{PM}_{2.5}$ concentrations is $193 \%$ at 50 m away, and $162 \%$ at 250 m away from the Crossing / connecting roadway. The number of days predicted to be in excess of the CWS is higher than No Build at distances up to 100 m from the ROW in all years.

For both of these combinations, the annual average concentrations are the same as No Build at all distance intervals and all years, with one exception. In 2035, the concentration is greater than the criterion of $15 \mathrm{ug} / \mathrm{m}^{3}$ at 50 m from the ROW.

The results for the Plaza C / Crossing C combination indicate that the biggest change in the maximum predicted $\mathrm{PM}_{2.5}$ concentrations at 50,100 and 250 m from the crossing / connecting roadway is approximately $167 \%, 161 \%$ and 133 \% respectively, in comparison to the No Build Alternative. These occur by 2035. Also, the number of days predicted to be greater than the CWS is significantly lower for this combination than the other two, with significant increases at distances up to 100 m from the ROW in 2025 and 2035. However, the annual average concentration is only marginally higher than No Build, and does not exceed the criterion at any of the distance intervals, in any of the horizon years.

The changes in the maximum predicted 1-hour NOx concentrations shown in Table 4.14 are greater than $10 \%$ at all distances and in all horizon years for the three combinations that are possible with Crossing C. At 50 m away, these changes range from a minimum of $119 \%$ to a maximum of $192 \%$. At 250 m away the changes range from $117 \%$ to $197 \%$. However, the MOE 1-hour $\mathrm{NO}_{\mathrm{x}}$
criterion is not exceeded in the vicinity of the crossing and connecting roadway for any combination.

Based on these results, a decrease in air quality is expected within 100 m of the connecting roadway of Crossing C with either Plaza A, Plaza B, or Plaza C.

|  | $\begin{aligned} & \text { Distance } \\ & \text { from ROW } \\ & (\mathrm{m}) \end{aligned}$ | Malden Ra to Labelle |  |  |  |  |  | le to Pultio |  |  |  |  |  | Pulford North of Lennon Drain |  |  | North of Lennon Drain to Cousineau Rd |  |  |  |  |  | Cousineau Rd to Howard Ave |  |  |  |  |  | Howard Ave to Highway 401 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Plaza A Alignment |  |  | Plaza BIC Aligmment |  |  | Plaza A Alignment |  |  | Plaza I C Aligment |  |  | Option 1 Alignment |  |  |  |  |  | Option 2 Alignment |  |  |  |  |  | Option 2 Alignment |  |  |    <br> 24 Hour   <br> Annual   <br> cWS   |  |  |
| Alternative 1A-2015 |  | $\begin{gathered} 24 \text { Hour } \\ 106 \% \end{gathered}$ | Annual |  | $\underset{\substack{24 \text { Hour } \\ \text { 94\% }}}{\text { ar }}$ | Annual | $\xrightarrow{>} \mathrm{Cws}$ |  |  |  |  |  |  |  |  |  |  | Annual | ${ }_{\text {c }}^{\text {cws }}$ | ${ }_{\substack{24 \mathrm{Hour} \\ 100 \%}}$ | Annual | ${ }_{\text {cws }}$ |  |  |  | ${ }_{\text {24 }}^{\text {240ur }}$ | Annual | ${ }^{\text {cws }}$ |  |  |  |
|  | 100 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 104\% | 100\% | 1 | 100\% | 100\% | 0 | 104\% | 100\% | 0 | 104\% | 922\% 91\% 910 | $\begin{aligned} & 0 \\ & 0 \end{aligned}$ |
|  | 250 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 104 | 91\% | 0 | 100\% | 110\% | 0 | 100 | 110\% | 0 | 109\% | 100 |  |
| Alternative 1A-2025 | 50 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 100\% | 100\% | -10 | 97\% | 100\% | -5 | 100\% | 100\% | -12 | 97\% | 1003 | -1 |
|  | 100 | 100\% | 100\%100\% | -100 | 103\%93\% | 100\%100\% | 0 |  |  |  |  |  |  | $\begin{aligned} & \text { 95\% } \\ & \text { 903\% } \\ & 1049 \end{aligned}$ | $\begin{aligned} & \text { 100\% } \\ & \text { 1008\% } \\ & \text { 1000\% } \end{aligned}$ |  | $\begin{aligned} & 933 \% \\ & 1036 \\ & 104 \% \end{aligned}$ | $\begin{aligned} & 100 \% \\ & \begin{array}{l} 100 \% \\ 929 \% \end{array} \\ & \hline 9 \end{aligned}$ | 10 | $\begin{gathered} 97 \% \\ 100 \% \end{gathered}$ | 100\%100\%100\% | $\begin{aligned} & -3 \\ & 0 \end{aligned}$ | $\begin{aligned} & \text { 103\% } \\ & \text { 107\% } \\ & \text { 104\% } \end{aligned}$ | $\begin{aligned} & 100 \% \\ & 100 \% \\ & 100 \% \end{aligned}$ | 0 <br> 7 | 107\%108\% | 100\% | $\begin{aligned} & 3 \\ & 0 \end{aligned}$ | $\begin{aligned} & \text { 103\% } \\ & 104 \% \end{aligned}$ | $\begin{aligned} & \text { 1000\% } \\ & \text { 1000\% } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & 1039 \\ & 1049 \\ & 104 \% \end{aligned}$ | $\begin{aligned} & 100 \% \\ & 100 \% \end{aligned}$ | 3 | 108\% | 1000 | $\bigcirc$ |
|  | 250 |  |  |  |  |  |  | 0 | 109\% | 100\% |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Alternative 1A-2035 | 50 |  | 100\% | -52 | 93\% | 94\% | ${ }^{31}$ | 95\% | 2100\% | 1 | 93\% | 94\% | -4 | 90\% | 100\% | -21 | 105\% | 100\% | 0 | 103\% | 93\% | -19 | 93\% | 100\% | -19 | 93\% | 100\% | -29 | 97\% | 92\% | -5 |  |  |  |
|  | 100 | 100\% | $\begin{aligned} & \text { 100\% } \\ & \text { 100\% } \end{aligned}$ | -13 <br> -3 | 93\% | 100\% | -7 | 103\% | 100\% | 7 | 103\% | 93\% | 7 | 100\% | 100\% | ${ }^{-3}$ | 113\% | 100\% | 15 | 109\% | 100\% | 6 | 100\% | 100\% | -1 | 110\% | 108\% | 4 | 104\% | 100\% | 0 |  |  |  |
|  | 250 | 97\% |  |  |  |  | - 3 | 104\% | 100\% | 0 | 104\% | 100\% | 0 | 108\% |  | 0 | 104\% | 100\% | 0 | 11\% | 100\% | 0 | 117\% | 100\% | 0 | 117\% | 100\% | 1 | 108\% | 100\% | 0 |  |  |  |

Table 4.2 Alternative 1A - Highest Maximum NOx Concentrations in Comparison to No Build

| Alternative 1A - 2015 | Distance from ROW <br> (m) | Madden Rd to Labelle |  |  |  | Labelle to Pulford |  |  |  | Pulford to North of Lennon <br> Drain |  | North of Lennon Drain to Cousineau Rd |  |  |  | Cousineau Rd to Howard Ave |  |  |  | Howard Ave to Highway 4c |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\frac{\text { Plaza A Alignment }}{}$ |  | Plaza BIC Alignment |  | Plaza A Alignme |  | Plaza Bic Alignment |  | ${ }^{1}$ Hour ${ }^{\text {a }}$ |  | $\frac{\text { Option } 1 \text { Alignment }}{1 \text { Hour }}$ |  | Option 2 Alignment |  | Option 1 Alignment |  | Option 2 Alignment |  |  |  |
|  | 50 | $\underbrace{1 \text { Hour }}_{\text {cheor }}$ | ${ }_{\text {220\% }}^{24 \mathrm{Hour}}$ | ${ }_{884 \%}^{1 \text { Hour }}$ | ${ }_{89 \%}^{24 \text { Hour }}$ | ${ }_{\substack{1 \\ 810 \\ \text { Hour }}}$ | ${ }_{\text {86\% }}^{24 \text { Hour }}$ | ${ }_{749 \%}^{1 \text { Hour }}$ | $\underset{89 \%}{24 \text { Hour }}$ |  |  | ${ }_{\text {1 }}^{\text {1 Hour }}$ 89\% | ${ }_{\text {2 }}^{\text {24 Hour }}$ | ${ }_{\substack{1 \text { Hour } \\ 58 \%}}$ | ${ }^{24 \mathrm{H} \text { gour }}$ | ${ }_{\substack{1 \text { Hour } \\ 75 \%}}$ | ${ }_{\text {290\% }}^{\text {24 Hour }}$ | ${ }_{\text {1 }}^{\text {1 }}$ 60\% | ${ }_{\text {2 }}^{\text {24 Hour }}$ |  |  |
| Alternative 1A - 2015 | 100 | 73\% | 98\% | 90\% | 101\% | 96\% | 91\% | 81\% | 92\% | ${ }^{66 \%}$ | ${ }^{85 \%}$ | 79\% | 87\% | 55\% | 87\% | 72\% | 88\% | 69\% | 88\% | ${ }^{80 \%}$ | 98\% |
|  | 250 | ${ }^{89 \%}$ | ${ }^{86 \%}$ | 95\% | ${ }^{93 \%}$ | 112\% | 99\% | 83\% | 95\% | 70\% | 95\% | 84\% | 98\% | 56\% | 94\% | 91\% | 98\% | 78\% | 96\% | $94 \%$ | 99\% |
|  | 50 | 77\% | 87\% | 82\% | 90\% | 71\% | 85\% | 72\% | 89\% | 43\% | 76\% | 78\% | 91\% | ${ }^{61 \%}$ | 91\% | 63\% | 84\% | 58\% | ${ }^{84 \%}$ | 67\% | 848 |
| Atternative 1A-2025 | 100 | ${ }^{83 \%}$ | 90\% | 89\% | 94\% | 79\% | 92\% | 76\% | 93\% | 54\% | 81\% | ${ }^{73 \%}$ | 83\% | 59\% | 83\% | ${ }^{61 \%}$ | 86\% | 58\% | 85\% | ${ }^{68 \%}$ | ${ }^{91 \%}$ |
|  | 250 | ${ }^{89 \%}$ | 94\% | 92\% | 95\% | 93\% | 96\% | 83\% | 95\% | 69\% | 94\% | 77\% | 94\% | 61\% | 92\% | 83\% | 95\% | 76\% | 94\% | 83\% | 97\% |
| Alternative 1A-2035 | 50 | 59\% | 85\% | 69\% | 95\% | ${ }^{84 \%}$ | 84\% | 69\% | 91\% | 30\% | 70\% | 65\% | 87\% | 55\% | 80\% | 58\% | 81\% | 53\% | 80\% | 64\% | 83\% |
|  | 100 | 69\% | ${ }^{89 \%}$ | 78\% | 93\% | ${ }^{85 \%}$ | 93\% | 69\% | 90\% | 42\% | 76\% | 66\% | 80\% | 54\% | ${ }^{80 \%}$ | 57\% | 84\% | 55\% | 83\% | $64 \%$ | 90\% |
|  | 250 | 82\% | 91\% | 86\% | 94\% | 107\% | 99\% | 81\% | 95\% | 57\% | 90\% | 66\% | 91\% | 58\% | 91\% | 77\% | 95\% | 73\% | 94\% | 80\% | 96\% |


| Aternativ 18-2015 | $\left.\begin{array}{\|c} \text { Distanceot fom } \\ \text { Row } \\ (\mathrm{m}) \end{array}\right)$ | n R to to |  |  |  |  |  | ${ }^{\text {ale }}$ to $P$ |  |  |  |  |  | Pulford North of Lennon Drain |  |  | North of Lemnon Drain to Cousinau Rd |  |  |  |  |  | Cousineau Rd to Howard Ave |  |  |  |  |  | Howard Ave to Highway 401 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | za Aligmment |  |  | Plaza $\mathrm{B} / \mathrm{CAlignment}$ |  |  | Plaza A Alignment |  |  | Plaza B/C Alignment |  |  |  |  |  | Oprion 1 Aligment |  |  | Option 2 Aligmment |  |  | Option 1 Aligmment |  |  | Ooption 2 Aligment |  |  | 24Hour <br> $100 \%$ <br> 10 |  |  |
|  |  | $\begin{gathered} 94 \% \\ 100 \% \\ 100 \% \end{gathered}$ | Annual ${ }^{\text {cws }}$ |  | ${ }^{24} \mathbf{2 4}$ Hour ${ }^{\text {and }}$ |  |  | ${ }_{\substack{24 \text { Hour } \\ 946}}$ | Annual | ${ }_{\text {cows }}$ | $\underset{\substack{24 \text { Hour } \\ 896}}{ }$ | ${ }_{\text {Annual }}^{\text {and }}$ | ${ }_{\text {c }}^{\text {cws }}$ | ${ }_{\text {chem }}^{\substack{24 \text { Hour }}}$ | Annual $92 \%$ | $\xrightarrow{\text { cws }}$ | ${ }_{\substack{24 \mathrm{Hour} \\ 919}}$ | Annual | ${ }_{\text {cws }}$ | $\underset{\substack{24 \mathrm{Hour} \\ 919}}{ }$ | Annual | ${ }_{\text {cws }}$ | ${ }_{\substack{24 \mathrm{Hour} \\ 910}}$ | ${ }^{\text {Annual }}$ 100\% | $\stackrel{\text { cws }}{4}$ | ${ }^{24 \mathrm{Hour}}$ | ${ }^{\text {Anuaal }}$ | $\stackrel{\text { cws }}{-7}$ |  |  |  |
|  | 50 |  | 92\% | ${ }^{-1}$ | 103\% | 100\% | ${ }^{-1}$ | 94\% | 100\% | ${ }^{-3}$ | 87\% | 100\% | ${ }^{-3}$ | 87\% | 100\% | ${ }_{-1}$ | 104\% | 100\% | 0 | 100\% | 100\% | 0 | 100\% | 100\% | 0 | 96\% | 100\% | 0 | 104\% | 91\% | 0 |
|  | 100 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 100\% | 0 |
|  | 250 |  | 109\% 100\% | ${ }_{7}$ | ${ }^{\text {100\% }}$ 97\% | 100\% 100\% | - | - $104 \%$ | 100\% 100\% | $\stackrel{-7}{ }$ | 92\%\% | 100\% 93\% | ${ }_{-7}{ }_{-7}$ | ${ }_{\text {100\% }}^{\text {79\% }}$ | 100\% 93\% | ${ }^{33}$ | 96\% 91\% | 100\% 100\% | ${ }^{-12}$ | 100\% ${ }_{\text {91\% }}^{\text {a }}$ | 91\% $92 \%$ | ${ }^{-15}$ | 100\% | 110\%\% | $\stackrel{0}{-9}$ | ${ }^{\text {100\% }}$ 86\% | 110\% | ${ }^{-18}$ | -109\% ${ }_{\text {97\% }}$ | 100\% | ${ }_{-1}^{0}$ |
| Aternative 18-2025 | ${ }^{50}$ | 103\% ${ }^{\text {103\% }}$ | 108\% | 7 | 103\% | 100\% | 2 | 97\% | 108\% | 6 | 91\% | 92\% | -4 | 87\% | 92\% | -5 | 100\% | 100\% | 2 | 97\% | 92\% | - | 100\% | 100\% | - | 97\% | 100\% | 。 | 108\% | 100\% | $\bigcirc$ |
|  | 250 | ${ }^{96 \%}$ | 100\% | - | 100\% | 100\% | 0 | 108\% | 100\% | 0 | 100\% | ${ }^{92 \%}$ | 0 | ${ }_{96 \%}$ | 100\% | 0 | ${ }^{96 \%}$ | 100\% | 2 | 104\% | 100\% | - | 100\% | 100\% | 0 | 100\% | 100\% | - | ${ }^{113}$ | 100\% | 0 |
| Aternative 18-2035 | 50 | 102\% | 100\% | - 4 | 98\% | 94\% | ${ }^{23}$ | 95\% | -100\% | 2 | 84\% | ${ }^{88 \%}$ | ${ }^{22}$ | ${ }^{80 \%}$ | 93\% | -42 | ${ }_{89} 89$ | 107\% | ${ }^{27}$ | ${ }^{89 \%}$ | ${ }^{93 \%}$ | ${ }^{33}$ | 85\% | 100\% | -39 | ${ }^{83 \%}$ | 100\% | -41 | 97\% | ${ }^{92 \%}$ | -5 |
|  | 100 | 103\% | 100\% | ${ }^{6}$ | 100\% | 100\% | -5 | 94\% | 100\% | 10 | ${ }^{89 \%}$ | 93\% | -6 | 94\% | 100\% | -8 | 97\% | 100\% | 2 | 100\% | 100\% | $\bigcirc$ | 100\% | 100\% | ${ }^{-1}$ | 100\% | 108\% | ${ }^{-1}$ | 104\% | 100\% | 0 |
|  | 250 | 100\% | 108\% | -1 | 97\% | 100\% | -3 | 107\% | 100\% | 0 | 96\% | 92\% | 0 | 100\% | 100\% | 0 | 100\% | 100\% | 0 | 104\% | 100\% | $\bigcirc$ | 113\% | 100\% | 0 | 113\% | 100\% | 0 | 108\% | 100\% | 0 |

Table 4.4 Alternative 1B - Highest Maximum NOx Concentrations in Comparison to No Build

|  | Distance fromROW$(\mathrm{m})$ | Malden Rd to Labelle |  |  |  | Labelle to Pulford |  |  |  | Pulford to North of LennonDrain |  | North of Lennon Drain to Cousineau Rd |  |  |  | Cousineau Rd to Howard Ave |  |  |  | Howard Ave to Highway 401 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Plaza A Alignment |  | Plaza BIC Alignment |  | Plaza A Alignment |  | Plaza B/C Alignment |  |  |  | Option 1 Alignment |  | Option 2 Alignment |  | Option 1 Alignment |  | Opfion 2 Alignment |  |  |  |
|  |  | 1 Hour | ${ }^{24}$ Hour | 1 Hour | ${ }^{24} \mathrm{Hour}$ | 1 Hour | ${ }^{24}$ Hour | 1 Hour | ${ }^{24}$ Hour |  |  | 1 Hour | ${ }^{24}$ Hour | 1 Hour | ${ }^{24} \mathrm{Hour}$ | 1 Hour | ${ }^{24}$ Hour | 1 Hour | ${ }^{24}$ Hour |  |  |
| Alternative 18-2015 | 50 | 78\% | 88\% | 88\% | ${ }^{91 \%}$ | 80\% | 85\% | 73\% | 85\% | 50\% | 77\% | 83\% | ${ }^{91 \%}$ | 55\% | ${ }^{88 \%}$ | 73\% | 86\% | 63\% | 84\% | 79\% | 92\% |
|  | 100 | ${ }^{82 \%}$ | 102\% | 94\% | 102\% | 93\% | 90\% | 70\% | 87\% | 61\% | 83\% | 77\% | ${ }^{84 \%}$ | 53\% | 84\% | 68\% | 87\% | 63\% | ${ }^{87 \%}$ | 77\% | 989 |
|  | 250 | ${ }^{89 \%}$ | 93\% | 95\% | 93\% | 106\% | 96\% | 84\% | 94\% | 70\% | 95\% | 81\% | 96\% | 56\% | 93\% | 91\% | 96\% | 77\% | 95\% | 92\% | 100\% |
| Alternative 18-2025 | 50 | 69\% | 87\% | 76\% | 90\% | 86\% | 89\% | 70\% | 86\% | 39\% | 75\% | 74\% | 88\% | 59\% | 88\% | 61\% | ${ }^{82 \%}$ | 56\% | ${ }^{82 \%}$ | 66\% | ${ }^{84 \%}$ |
|  | 100 | 79\% | 90\% | ${ }^{83 \%}$ | 93\% | 90\% | 94\% | 69\% | 89\% | 53\% | 80\% | 69\% | ${ }^{82 \%}$ | 57\% | ${ }^{82 \%}$ | 59\% | 85\% | 58\% | 85\% | 68\% | 91\% |
|  | 250 | 86\% | 3\% | 91\% | 95\% | 103\% | 99\% | 86\% | 95\% | 67\% | 94\% | 75\% | 929, | 61\% | ${ }^{929}$ | ${ }^{81 \%}$ | 95\% | ${ }^{76 \%}$ | 95\% | ${ }^{829}$ | 96\% |
| Alternative 18-2035 | 50 | 64\% | 85\% | 70\% | 88\% | 83\% | 84\% | 68\% | 81\% | 34\% | 69\% | 63\% | ${ }^{86 \%}$ | 53\% | 80\% | 57\% | 80\% | 52\% | ${ }^{79 \%}$ | ${ }^{63 \%}$ | 83\% |
|  | 100 | 75\% | 90\% | 79\% | 93\% | 84\% | 93\% | 64\% | 88\% | 41\% | 76\% | 62\% | 80\% | 53\% | 80\% | 56\% | 83\% | 55\% | ${ }^{82 \%}$ | 64\% | 90\% |
|  | 250 | 83\% | 93\% | 84\% | 94\% | 106\% | 99\% | 81\% | 94\% | 55\% | 90\% | 64\% | 91\% | 57\% | 91\% | 75\% | 94\% | 73\% | 94\% | 80\% | 96\% |


|  | Distance fromROW$(\mathrm{m})$ | Malden Rd to Labelle |  |  |  |  |  | Labelle to Pultord |  |  |  |  |  | Pulford North of Lennon Drain |  |  | North of Lennon Drain to Cousineau Rd |  |  |  |  |  | Cousineau Rd to Howard Ave |  |  |  |  |  | Howard Ave to Highway 401 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | ${ }^{248 \mathrm{Hour}}$ | ara Aligment |  |  |  |  | Plaza A Aligment |  |  | Plaza $/$ / Alignment |  |  | Howr Anual |  |  | $\underline{1} 1$ Alignm |  |  | Option 2 Aligmment |  |  | n 1 Align |  |  | Option 2 Aligmment |  |  | ${ }^{24 \mathrm{Hou}}$ |  |  |
| Atterative 2A - 2015 | 50 |  | ${ }^{\text {Annual }} 92$ | ${ }_{\text {che }}{ }^{\text {17 }}$ | ${ }_{\text {2 }}^{\text {2 }}$ 88\%\% | ${ }^{\text {Annual }}$ | ${ }_{\text {cows }}$ | ${ }_{\substack{24 \text { Hour } \\ 89 \%}}^{\text {ar }}$ | ${ }_{\text {annual }}^{\text {a3\% }}$ | ${ }_{\text {c }}{ }_{\text {cws }}$ | ${ }_{\text {2 }}^{29 \text { Hour }}$ | ${ }^{\text {Annual }}$ 93\% | ${ }_{-14}$ | ${ }^{224 \text { Hour }}$ | ${ }^{\text {Annual }} 92$ | ${ }_{\text {cher }}$ | ${ }^{247400} 9$ | Annual | ${ }_{-4}^{\text {cws }}$ | ${ }_{\text {2 }}^{\text {24 } 8 \text { \%our }}$ | ${ }^{\text {Annual }}$ 100\% | $\xrightarrow{\text { cws }}$ | ${ }_{919}^{24 \mathrm{Hour}}$ | ${ }^{\text {Annual }}$ 100\% | ${ }_{\text {> }}^{\substack{\text { cws } \\-6}}$ | $\underbrace{\text { 220 }}_{\text {24 Hour }}$ | ${ }^{\text {Annual }}$ | ${ }_{\text {c }}^{\text {cws }}$ |  | ${ }^{\text {Annual }} 100 \%$ | $\stackrel{\text { cws }}{0}$ |
|  | 100 | ${ }_{94 \%}$ | ${ }^{88 \%}$ | ${ }^{-3}$ | 95\% | ${ }^{88 \%}$ | -3 | 94\% | 108\% | ${ }^{-3}$ | 97\% | 100\% | -2 | 83\% | 92\% | ${ }_{-1}$ | 104\% | 100\% | 0 | 96\% | 100\% | 0 | 96\% | 100\% | 0 | 93\% | 100\% | 0 | 100\% | 100\% | 0 |
|  | 250 | 95\% | 99\% | 0 | 95\% | 99\% | 0 | 96\% | 100\% | 0 | 96\% | 100\% | 0 | 96\% | 100\% | 0 | 100\% | 100\% | 0 | 96\% | ${ }_{91}$ | 0 | ${ }^{966}$ | 110 | 0 | 96\% | ${ }^{110 \%}$ | 0 | 109\% | 1008 | 0 |
| Aterenative 2A - 2025 | 50 | 87\% | ${ }^{91 \%}$ | ${ }^{23}$ | 87\% | 92\% | $-26$ | 85\% | 93\% | -22 | ${ }^{85 \%}$ | 93\% | -25 | 82\% | 93\% | ${ }^{29}$ | 100\% | 100\% | -7 | 91\% | ${ }^{92 \%}$ | ${ }^{-14}$ | ${ }^{928}$ | 1004 | -12 | 86\% | 100\% | ${ }^{-18}$ | 97\% | 1008 | -1 |
|  | 100 | 99\% | ${ }^{93 \%}$ | -4 | 101\% | 94\% | 4 | 100\% | 108\% | 10 | 100\% | 100\% | 1 | 94\% | 92\% | -5 | 103\% | 100\% | 0 | 97\% | ${ }^{92 \%}$ | 0 | 100\% | 100\% | 0 | ${ }^{93 \%}$ | 100\% | 0 | ${ }^{108 \%}$ | 1003 | 0 |
|  | 250 | 95\% | 93\% | 0 | 95\% | ${ }^{94 \%}$ | 0 | 104\% | 92\% | 0 | 100\% | 100\% | 0 | 104\% | 100\% | 0 | 104\% | 100\% | 0 | 100\% | 100\% | 0 | 100\% | 100\% | 0 | 100\% | 100\% | 0 | 109\% | 100\% | 0 |
| Atermative 2A - 2035 | 50 | 85\% | 91\% | ${ }^{42}$ | 85\% | ${ }^{91 \%}$ | -44 | ${ }^{86 \%}$ | 94\% | ${ }^{31}$ | 86\% | 94\% | ${ }^{33}$ | 88\% | 93\% | ${ }^{37}$ | 100\% | 100\% | -15 | ${ }^{89 \%}$ | 93\% | ${ }^{.33}$ | 93\% | 100\% | ${ }^{21}$ | 80\% | 93\% | -44 | 100\% | 92\% | -5 |
|  | 100 | 100\% | ${ }_{91 \%}$ | ${ }^{17}$ | 101\% | ${ }^{91 \%}$ | ${ }^{-17}$ | 94\% | 107\% | -10 | 100\% | 100\% | 1 | 100\% | 100\% | -5 | 103\% | 108\% | 1 | 94\% | 100\% | ${ }^{-3}$ | 103\% | 100\% | -2 | 97\% | 100\% | ${ }^{-3}$ | 104\% | 92\% | 0 |
|  | 250 | 95\% | 98\% | ${ }^{-3}$ | 96\% | 98\% | 3 | 100\% | 1008 | 0 | 104\% | 100\% | 0 | 104\% | 109\% | 0 | 107\% | 109\% | 0 | 1008 | 100\% | 0 | 113 | 100\% | 0 | 113 | 100\% | 0 | 113\% | 100\% | 0 |

Table 4.6 Alternative 2A - Highest Maximum NOx Concentrations in Comparison to No Build

|  | Distance from ROW <br> (m) | Malden Rd to Labelle |  |  |  | Labelle to Pulford |  |  |  | Pulford to North of Lennon Drain |  | North of Lennon Drain to Cousineau Rd |  |  |  | Cousineau Rd to Howard Ave |  |  |  | Howard Ave to Highway 40 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Plaza A Alignment |  | Plaza BIC Alignment |  | Plaza A Alignment |  | Plaza BIC Alignment |  |  |  | Option 1 Aligmment |  | Option 2 Alignment |  | Option 1 Alignment |  | Option 2 Alignment |  | ${ }^{1}{ }^{\text {Hour }}$ ( ${ }^{24}{ }^{24 \text { Hour }}$ |  |
| Alterative 2A - 2015 | 50 | $\begin{aligned} & \text { 17 } 65 \% \\ & 65 \% \end{aligned}$ | 24 Hour | $\begin{gathered} 1 \text { Hour } \\ 68 \% \end{gathered}$ | ${ }_{\substack{24 \mathrm{Hour} \\ 81 \%}}$ | 1 Hour | ${ }^{24 \mathrm{H} \text { \%our }}$ | $\begin{gathered} 1 \text { Hour } \\ 74 \% \end{gathered}$ | 24 Hour $85 \%$ | $\begin{gathered} 1 \text { Hour } \\ 45 \% \end{gathered}$ | $24 \text { Hour }$ | ${ }_{\substack{1 \text { Hour } \\ 59 \%}}$ | ${ }_{922 \%}^{24 \text { Hour }}$ | ${ }_{\text {1 }}^{\text {1 Hour }}$ 53\% | 24 Hour $88 \%$ | ${ }_{7}^{1 \text { Hour }}$ |  | ${ }_{1}^{1 \text { Hour }}$ 64\% | ${ }_{\substack{24 \mathrm{Hour} \\ 84 \%}}$ |  |  |
|  | 100 | 82\% | 100\% | 75\% | 97\% | 98\% | 96\% | 70\% | 88\% | 53\% | 82\% | 54\% | 86\% | 52\% | 84\% | 74\% | 88\% | 69\% | 87\% | 77\% | 98\% |
|  | 250 | ${ }^{89 \%}$ | 93\% | 93\% | 92\% | 100\% | 95\% | 87\% | 94\% | 65\% | $94 \%$ | 60\% | 95\% | 58\% | 94\% | 85\% | 98\% | 85\% | 96\% | ${ }^{89 \%}$ | 99\% |
| Alternative 2A - 2025 | 50 | 68\% | ${ }^{87 \%}$ | 66\% | ${ }^{87 \%}$ | 81\% | 85\% | 70\% | 80\% | 38\% | 74\% | ${ }^{62 \%}$ | 91\% | 58\% | 88\% | 65\% | 86\% | 60\% | 82\% | 65\% | 85\% |
|  | 100 | 83\% | 93\% | 76\% | 93\% | 87\% | 95\% | 70\% | 90\% | 50\% | 80\% | 60\% | 83\% | 56\% | 82\% | 69\% | ${ }^{89 \%}$ | 58\% | 85\% | 68\% | 91\% |
|  | 250 | 97\% | 96\% | 90\% | 95\% | 98\% | 96\% | ${ }^{86 \%}$ | 95\% | 63\% | 92\% | 63\% | 94\% | 62\% | 92\% | 88\% | 97\% | 78\% | 94\% | 82\% | 97\% |
| Alternative 2A - 2035 | 50 | 65\% | 85\% | 63\% | 85\% | 76\% | 80\% | 66\% | 81\% | 33\% | 69\% | 54\% | 87\% | 51\% | 85\% | 59\% | 82\% | 56\% | 80\% | 66\% | 80\% |
|  | 100 | 78\% | 92\% | 71\% | 92\% | 79\% | ${ }^{89 \%}$ | 64\% | ${ }^{89 \%}$ | 40\% | 76\% | 53\% | 80\% | 52\% | 78\% | 62\% | 85\% | 55\% | 83\% | 65\% | 90\% |
|  | 250 | 93\% | 95\% | 86\% | 94\% | $99 \%$ | 96\% | 83\% | 95\% | 53\% | 90\% | 58\% | 92\% | 57\% | 91\% | 82\% | 96\% | 75\% | 94\% | 80\% | 96\% |


|  |  |  |  |  |  |  |  |  |  |  |  | Highest | B. Conce | 析 | , | fw | ualm |  |  |  |  |  |  |  |  |  |  |  | Howard Ave to tighway 401 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & \text { Distance from } \\ & \text { ROW } \\ & \text { (m) } \end{aligned}$ | Malden Rd to Label |  |  |  |  |  | Labelle to Pulford |  |  |  |  |  | Pulford North of Lennon Drain |  |  | North of Lennon Drain to Cousineau Rd |  |  |  |  |  | Cousineau Rd to Howard Ave |  |  |  |  |  |  |  |  |
|  |  | $z \mathrm{Alilgm}$ |  |  |  |  |  | Plaza Aligmment |  |  | Plaza IC Alignment |  |  |  |  |  | Option 1 Alignment |  |  | Option 2 Aligment |  |  | Oprion 1 Alignment |  |  | Oprion 2 Alignment |  |  | ${ }_{\substack{24 \mathrm{Hog} \\ 1048}}^{\text {20 }}$ | Hour |  |
| Alterative 28-2015 | 50 | $\underbrace{2}_{\substack{24 \mathrm{Hou} \\ 960 \%}}$ | ${ }^{\text {Annual }}{ }^{\text {Pws }}$ | ${ }_{\text {cows }}$ | $\stackrel{\text { Plazab } / \text { C Alignment }}{\text { 24 Hour }}$ |  |  | $\underset{\substack{24 \mathrm{Hour} \\ 970 \%}}{ }$ | Annual | ${ }_{-13}$ | ${ }_{\substack{24 \mathrm{Hour} \\ 9 \text { 90\% }}}$ | ${ }_{\substack{\text { Annual } \\ 9380}}$ | $\underset{-13}{\substack{c \mathrm{cws}}}$ | ${ }_{\text {240ur }}^{\text {740\% }}$ | ${ }^{\text {Annual }}$ | $\stackrel{\text { cws }}{\text {-17 }}$ | 24 Hour $88 \%$ | Annual | ${ }_{\text {c }}^{\text {cws }}$ | ${ }_{\text {24 }}^{\text {240ur }}$ | ${ }^{\text {Annual }}$ 100\% | ${ }_{-}^{\text {cws }}$ | ${ }_{\text {24 }}^{85 \%}$ | ${ }^{\text {Annual }}$ | ${ }_{-8}{ }_{-}^{\text {cws }}$ | ${ }_{\text {24 }}^{\text {85\%ur }}$ | ${ }^{\text {Annual }}$ 100\% | $\stackrel{>}{-8}$ |  |  |  |
|  |  | ${ }_{89 \%}$ | 86\% | ${ }^{-3}$ | 90\% | 87\% | ${ }^{-3}$ | 94\% | 100\% | ${ }^{-3}$ | 90\% | 100\% | ${ }^{-3}$ | 83\% | 92\% | ${ }_{-1}$ | 96\% | 100\% | 0 | 96\% | 100\% | 0 | 93\% | 100\% | 0 | 96\% | 100\% | - | 104\% | 100\% | 0 |
|  | 100 | ${ }^{94 \%}$ | 98\% | 0 | 95\% | 98\% | 0 | 96\% | 100\% | 0 | 96\% | 100\% | 0 | 96\% | 100\% | - | 96\% | 100\% | 0 | 100\% | 91\% | 0 | 96\% | 110\% | 0 | 100\% | 110\% | 0 | 114\% | 100\% | 0 |
|  | 250 | 97\% | 93\% | -20 | 97\% | 93\% | ${ }^{-22}$ | 95\% | 93\% | -19 | 95\% | 93\% | ${ }^{21}$ | 79\% | 93\% | ${ }^{-33}$ | 86\% | 100\% | -12 | 86\% | 92\% | -15 | 89\% | 100\% | - 16 | 80\% | 100\% | -17 | 97\% | 100\% | ${ }_{-1}$ |
| Atermative 2 B - 2025 | 50 100 | ${ }_{93 \%}^{99 \%}$ | 992\% | -20 | 94\% | ${ }_{92 \%}$ | 2 | 100\% | 100\% | - | 94\% | 100\% |  | 90\% | 92\% | 5 | 90\% | 100\% | ${ }_{-1}$ | 93\% | 92\% | 0 | 93\% | 100\% | ${ }_{-1}$ | 93\% | 100\% | 0 | 108\% | 100\% | 。 |
|  | 250 | 94\% | ${ }^{93 \%}$ | 0 | 94\% | ${ }_{93 \%}$ | 0 | 104\% | 92\% | 0 | 104\% | 100\% | 0 | 104\% | 100\% | 0 | 96\% | 100\% | 0 | 100\% | 100\% | 0 | 100\% | 100\% | 0 | 100\% | 100\% | 0 | 109\% | 100\% | - |
| Alternative 28-2035 | 50 | 97\% | ${ }_{93 \%}$ | ${ }^{36}$ | 97\% | ${ }_{93 \%}$ | ${ }^{-38}$ | 98\% | 94\% | ${ }^{25}$ | 98\% | 94\% | ${ }^{27}$ | 73\% | 87\% | -49 | 87\% | 93\% | ${ }^{-31}$ | 84\% | ${ }_{93 \%}$ | ${ }^{35}$ | ${ }_{85 \%}$ | 100\% | -39 | ${ }^{80 \%}$ | 93\% | -44 | 100\% | 92\% | . |
|  | 100 | 92\% | 88\% | $-21$ | 94\% | ${ }^{89 \%}$ | -19 | 97\% | 100\% | 6 | 91\% | 93\% | -6 | 91\% | ${ }_{92 \%}$ | -10 | 94\% | 100\% | -4 | 91\% | 100\% | -4 | 97\% | 100\% | -3 | ${ }_{97 \%}$ | 100\% | ${ }^{-3}$ | 104\% | ${ }^{92 \%}$ | 0 |
|  | 250 | 93\% | 97\% | ${ }^{3}$ | 94\% | 98\% | ${ }^{-}$ | 104\% | 92\% | 0 | 104\% | 100\% | 0 | 100\% | 100\% | 0 | 100\% | 100\% | $\bigcirc$ | 100\% | 100\% | $\bigcirc$ | 113\% | 100\% | 0 | 113\% | 100\% | 0 | 113\% | 100\% | 0 |

Table 4.8 Alternative 2B - Highest Maximum NOx Concentrations in Comparison to No Build

|  | $\begin{aligned} & \text { Distance from } \\ & \text { ROW } \end{aligned}$ | Malden Rd to Labelle |  |  |  | elle to Pulford |  |  |  | Pulford to North of LennonDrain |  | North of Lennon Drain to Cousineau R |  |  |  | Cousineau Rd to Howard Ave |  |  |  | oward Ave to Highway 401 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Plaza A Alignment |  | abic Alig |  | Plaza A |  | aza Blc |  |  |  | Option 1 Alighment |  | Option 2 Alignment |  | Option 1 Aligment |  | Option 2 Alignment |  | ${ }^{1}{ }^{1 \text { Hour }}$ |  |
| Aternative 2B-2015 | 50 | 1 Hour | 24 Hour $83 \%$ | $1 \text { Hour }$ | 24 Hour $81 \%$ | $\begin{gathered} 1 \text { Hour } \\ 86 \% \end{gathered}$ | 24 Hour $84 \%$ | $\begin{aligned} & 1 \text { Hour } \\ & 74 \% \% \end{aligned}$ | 24 Hour $84 \%$ | 1 Hour <br> $430 \%$ | 24 Hour $75 \%$ | $1 \text { Hour }$ | ${ }^{\text {24 }}$ 88\%\% | ${ }_{\substack{1 \\ \text { 54\%\% }}}^{\text {Hour }}$ | ${ }_{888}^{24 \text { Hour }}$ | ${ }_{\text {1 }}^{\text {1 }}$ 65\%\% | ${ }_{\text {85\% }}^{24 \mathrm{Hour}}$ | $\underbrace{1 \text { Hour }}_{63 \%}$ | ${ }_{\text {23\% }}^{24} \mathbf{H}$ |  |  |
|  | 100 | 82\% | 100\% | 75\% | 97\% | 98\% | $87 \%$ | 70\% | 87\% | 52\% | 82\% | 51\% | 838 | 53\% | 84\% | 69\% | 88\% | 63\% | 859\% | 77\% | 999, |
|  | 250 | 88\% | 93\% | ${ }^{91 \%}$ | 92\% | 101\% | 95\% | 87\% | 94\% | 59\% | 93\% | 58\% | 94\% | 58\% | 94\% | 85\% | 98\% | 77\% | 95\% | 90\% | 99\% |
| Alternative 2B-2025 | 50 | 67\% | ${ }^{87 \%}$ | 66\% | 87\% | 81\% | 85\% | 70\% | 85\% | 37\% | 75\% | 58\% | 88\% | 5\%\% | 86\% | 63\% | 84\% | 58\% | 82\% | 64\% | 85\% |
|  | 100 | 78\% | 90\% | 74\% | 90\% | 86\% | 92\% | 70\% | 90\% | 47\% | 80\% | 56\% | 81\% | 54\% | 78\% | 66\% | 86\% | 57\% | 85\% | 67\% | 91\% |
|  | 250 | 91\% | 93\% | 88\% | 94\% | 100\% | 96\% | 85\% | 95\% | 62\% | 92\% | 61\% | ${ }^{92 \%}$ | 59\% | 88\% | 86\% | 94\% | 77\% | 94\% | ${ }^{81 \%}$ | 960 |
| Alternative 28-2035 | 50 | 64\% | 85\% | 62\% | 85\% | 73\% | 80\% | 65\% | 81\% | 33\% | 71\% | 53\% | ${ }^{85 \%}$ | 51\% | 85\% | 59\% | ${ }^{82 \%}$ | 54\% | 799 | 66\% | 84\% |
|  | 100 | 75\% | 90\% | 71\% | 90\% | 77\% | ${ }^{89 \%}$ | 64\% | 89\% | 41\% | 76\% | 52\% | ${ }^{78 \%}$ | 52\% | 78\% | ${ }^{52 \%}$ | 85\% | 54\% | ${ }^{82 \%}$ | 64\% | 90\% |
|  | 250 | 88\% | 94\% | 85\% | 94\% | 97\% | 95\% | 83\% | 95\% | 53\% | 0\% | 57\% | 91\% | 57\% | 91\% | 83\% | 97\% | 75\% | 94\% | 79\% | 96\% |

Table 4.9 Alternative 3 - Highest Maximum PM2.5 Concentrations in Comparison to No Build

|  | Distance fromROW (m) | Malden Rd to Labelle |  |  |  |  |  | Labelle to Pufford |  |  |  |  |  | Pulford North of Lennon Drain |  |  | North of Lennon Drain to CousineauRd |  |  | Cousineau Rd to Howard Ave |  |  | Howard Ave to Highway 401 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Plaza A Alignment |  |  |  |  |  |  |  |  |  |  |  | ${ }^{24}$ Hour ${ }^{\text {Annual }}$ - ${ }^{\text {cWS }}$ |  |  | ${ }^{24}$ Hour | Annual | > cws | ${ }^{24}$ Hour | Annual |  | 24 Hour | $\begin{gathered} \text { Annual } \\ \text { 100\% } \end{gathered}$ | $>{ }^{\text {cws }}$ |
|  | 50 | $\underset{\substack{24 \mathrm{Hour}}}{\substack{24 \%}}$ | Annual | >cws | Plaza |  |  | $\stackrel{\text { Plaza } \mathrm{A} \text { Alignment }}{\text { 24 }}$ |  |  | $\xrightarrow{\text { Plaza B/C Alignment }}$ |  |  |  |  |  | $\underset{-8}{>\mathrm{Cws}}$ |  |  |  |  |  |  |  |
|  | 100 | 100\% | 92\% | -2 | 103\% | 92\% | ${ }^{-1}$ | 81\% | 92\% | ${ }^{-3}$ | 84\% | 92\% | ${ }^{-3}$ | 73\% | 83\% | ${ }^{-1}$ |  | 93\% | 91\% | 0 | 79\% | 91\% | 0 | 104\% | 100\% | 0 |
|  | 250 | 104\% | 100\% | 0 | 104\% | 100\% | 0 | 88\% | 91\% | 0 | 92\% | 91\% | 0 | 92\% | 91\% | 0 | 92\% | 91\% | 0 | 88\% | 100\% | 0 | 114\% | 110\% | 0 |
| Alterative 3-2025 | 50 | 100\% | 930, | -12 | 97\% | 93\% | ${ }^{-13}$ | 70\% | 80\% | -40 | 73\% | 87\% | -40 | 56\% | 71\% | ${ }^{-38}$ | 80\% | ${ }^{85 \%}$ | -17 | 64\% | 77\% | ${ }^{20}$ | 100\% | 100\% | 0 |
|  | 100 | 103\% | 100\% | -1 | 109\% | 100\% | 5 | ${ }^{81 \%}$ | 85\% | -4 | 81\% | 92\% | -4 | 71\% | 77\% | -5 | 87\% | 83\% | ${ }^{-1}$ | 73\% | 83\% | ${ }^{-1}$ | 108\% | 100\% | 0 |
|  | 250 | 104\% | 100\% | 0 | 111\% | 100\% | 1 | 88\% | 83\% | 0 | 92\% | ${ }^{92 \%}$ | 0 | 88\% | 91\% | 0 | ${ }^{88 \%}$ | 91\% | 0 | 85\% | 91\% | 0 | 113\% | 100\% | 0 |
| Alterative 3-2035 | 50 | 95\% | 94\% | -44 | 100\% | 94\% | -25 | 67\% | 75\% | -74 | 70\% | 81\% | -72 | 56\% | 67\% | -58 | 79\% | ${ }^{79 \%}$ | -39 | 58\% | 71\% | -50 | 100\% | 100\% | 1 |
|  | 100 | 103\% | 93\% | -16 | 115\% | 100\% | -3 | 77\% | 79\% | -15 | 77\% | 86\% | -15 | ${ }^{72 \%}$ | 77\% | 10 | 84\% | 83\% | -5 | 74\% | 83\% | -4 | 107\% | 100\% | 1 |
|  | 250 | 100\% | 100\% | -2 | 107\% | 100\% | -1 | 89\% | 92\% | 0 | 93\% | 92\% | 0 | 85\% | 91\% | 0 | 89\% | 91\% | 0 | 92\% | 91\% | 0 | 113\% | 100\% | 0 |

Table 4.10 Alternative 3-Highest Maximum NOx Concentrations in Comparison to No Build

| Highest NOx Concentration at Intervals from Right of Way (ug/ $\mathrm{m}^{3}$ ) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Distance from <br> ROW <br> (m) | Malden Rd to Labelle |  |  |  | Labelle to Pulford |  |  |  | Pulford to North of LennonDrain |  | North of Lennon Drain to Cousineau Rd |  | Cousineau Rd to Howard Ave |  | Howard Ave to Highway 401 |  |
|  |  | Plaza A Alignment |  | Plaza B/C Alignment |  | Plaza A Alignment |  | Plaza B/C Alignment |  | $\begin{gathered} 1 \text { Hour } \\ 49 \% \end{gathered}$ | 24 Hour | $\begin{aligned} & 1 \text { Hour } \\ & 92 \% \end{aligned}$ | $24 \text { Hour }$ | $1 \text { Hour }$ | $\begin{aligned} & 24 \text { Hour } \\ & 80 \% \end{aligned}$ | $\begin{aligned} & \hline \text { 1 Hour } \\ & \text { 101\% } \end{aligned}$ | $\begin{aligned} & 24 \text { Hour } \\ & 97 \% \end{aligned}$ |
| Alternative 3-2015 | 50 | $\begin{gathered} \hline 1 \text { Hour } \\ 86 \% \end{gathered}$ | $\begin{gathered} \hline 24 \text { Hour } \\ 84 \% \end{gathered}$ | $\begin{gathered} \hline 1 \text { Hour } \\ 91 \% \end{gathered}$ | $24 \text { Hour }$ | $\begin{gathered} \hline 1 \text { Hour } \\ 87 \% \end{gathered}$ | $\begin{aligned} & 24 \text { Hour } \\ & 90 \% \end{aligned}$ | $\begin{gathered} \hline 1 \text { Hour } \\ 87 \% \end{gathered}$ | $\begin{gathered} 24 \text { Hour } \\ 90 \% \end{gathered}$ |  |  |  |  |  |  |  |  |
|  | 100 | 88\% | 100\% | 92\% | 101\% | 88\% | 95\% | 88\% | 95\% | 64\% | 79\% | 86\% | 98\% | 90\% | 85\% | 111\% | 102\% |
|  | 250 | 98\% | 92\% | 99\% | 94\% | 94\% | 98\% | 94\% | 98\% | 82\% | 93\% | 85\% | 95\% | 115\% | 94\% | 122\% | 103\% |
| Alternative 3-2025 | 50 | 84\% | 89\% | 83\% | 90\% | 74\% | 87\% | 74\% | 87\% | 40\% | 73\% | 77\% | 96\% | 64\% | 80\% | 75\% | 88\% |
|  | 100 | 90\% | 94\% | 87\% | 94\% | 75\% | 93\% | 75\% | 93\% | 54\% | 78\% | 73\% | 88\% | 68\% | 84\% | 78\% | 95\% |
|  | 250 | 97\% | 96\% | 96\% | 95\% | 88\% | 96\% | 88\% | 96\% | 69\% | 92\% | 78\% | 95\% | 93\% | 94\% | 95\% | 99\% |
| Alternative 3-2035 | 50 | 72\% | 88\% | 70\% | 89\% | 70\% | 83\% | 70\% | 83\% | 35\% | 68\% | 66\% | 93\% | 59\% | 78\% | 71\% | 86\% |
|  | 100 | 76\% | 93\% | 73\% | 93\% | 68\% | 90\% | 68\% | 90\% | 43\% | 74\% | 66\% | 84\% | 64\% | 82\% | 73\% | 93\% |
|  | 250 | 89\% | 95\% | 85\% | 95\% | 87\% | 96\% | 87\% | 96\% | 57\% | 90\% | 69\% | 92\% | 87\% | 94\% | 89\% | 97\% |

Table 4.11 - PLAZA ALTERNATIVES - Highest Maximum PM $_{2.5}$ Concentrations in Comparison to No Build

| Year | Distance from Property Boundary (m) | Plaza A |  |  | Plaza B |  |  | Plaza B1 |  |  | Plaza C |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 24 Hour | Annual | >CWS | 24 Hour | Annual | >CWS | 24 Hour | Annual | >CWS | 24 Hour | Annual | >CWS |
| 2015 | 50 | 217\% | 200\% | 134 | 250\% | 164\% | 127 | 283\% | 200\% | 148 | 200\% | 136\% | 84 |
|  | 100 | 167\% | 140\% | 15 | 190\% | 120\% | 20 | 209\% | 140\% | 54 | 195\% | 140\% | 39 |
|  | 250 | 129\% | 120\% | 0 | 138\% | 110\% | 0 | 130\% | 110\% | 0 | 162\% | 120\% | 2 |
| 2025 | 50 | 236\% | 200\% | 156 | 284\% | 182\% | 167 | 374\% | 220\% | 177 | 212\% | 155\% | 110 |
|  | 100 | 173\% | 150\% | 36 | 209\% | 130\% | 35 | 239\% | 150\% | 77 | 205\% | 140\% | 59 |
|  | 250 | 136\% | 130\% | 1 | 141\% | 110\% | 3 | 143\% | 110\% | 8 | 173\% | 120\% | 6 |
| 2035 | 50 | 250\% | 209\% | 168 | 288\% | 200\% | 175 | 383\% | 240\% | 193 | 223\% | 164\% | 129 |
|  | 100 | 186\% | 150\% | 56 | 218\% | 140\% | 48 | 250\% | 160\% | 87 | 223\% | 150\% | 77 |
|  | 250 | 136\% | 130\% | 3 | 150\% | 110\% | 8 | 140\% | 120\% | 11 | 182\% | 130\% | 17 |

Table 4.12 - PLAZA ALTERNATIVES - Highest Maximum NO $\mathrm{N}_{\mathrm{x}}$ Concentrations in Comparison to No Build

| Year | Distance from Property Boundary$\qquad$ | Plaza A |  | Plaza B |  | Plaza B1 |  | Plaza C |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1-Hour | Exceedances | 1-Hour | Exceedances | 1-Hour | Exceedances | 1-Hour | Exceedances |
| 2015 | 50 | 344\% | 8 | 621\% | 6 | 522\% | 2 | 123\% | 0 |
|  | 100 | 194\% | 0 | 463\% | 2 | 368\% | 1 | 128\% | 0 |
|  | 250 | 107\% | 0 | 252\% | 0 | 223\% | 0 | 116\% | 0 |
| 2025 | 50 | 805\% | 14 | 929\% | 18 | 790\% | 7 | 213\% | 0 |
|  | 100 | 458\% | 1 | 705\% | 7 | 590\% | 1 | 208\% | 0 |
|  | 250 | 393\% | 0 | 287\% | 0 | 310\% | 0 | 173\% | 0 |
| 2035 | 50 | 886\% | 16 | 1096\% | 17 | 691\% | 6 | 222\% | 0 |
|  | 100 | 533\% | 1 | 758\% | 6 | 655\% | 3 | 216\% | 0 |
|  | 250 | 448\% | 0 | 356\% | 0 | 306\% | 0 | 176\% | 0 |

Table 4.13-CROSSING ALTERNATIVES - Highest Maximum PM ${ }_{2.5}$ Concentrations in Comparison to No Build

| Year | Distance from ROW <br> (m) | Crossing A |  |  | Crossing B |  |  | Crossing B |  |  | Crossing C |  |  | Crossing C |  |  | Crossing C |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | From Plaza A |  |  | From Plaza A |  |  | From Plaza B1 |  |  | From Plaza A |  |  | From Plaza B |  |  | From Plaza C |  |  |
|  |  | 24 Hour | Annual | Exceedances | 24 Hour | Annual | Exceedances | 24 Hour | Annual | Exceedances | 24 Hour | Annual | Exceedances | 24 Hour | Annual | Exceedances | 24 Hour | Annual | Exceedances |
| 2015 |  | 150\% | 136\% | 49 | 150\% | 136\% | 49 | 188\% | 136\% | 60 | 150\% | 136\% | 49 | 161\% | 127\% | 50 | 142\% | 100\% | 15 |
|  | 100 | 135\% | 130\% | 14 | 135\% | 130\% | 14 | 170\% | 130\% | 20 | 135\% | 130\% | 14 | 136\% | 120\% | 20 | 139\% | 109\% | 1 |
|  | 250 | 127\% | 120\% | , | 127\% | 120\% | 0 | 136\% | 110\% | 0 | 127\% | 120\% | 0 | 129\% | 110\% | 0 | 119\% | 100\% | 0 |
| 2025 | 50 | 171\% | 145\% | 83 | 171\% | 145\% | 83 | 204\% | 145\% | 87 | 171\% | 145\% | 83 | 182\% | 136\% | 78 | 158\% | 118\% | 35 |
|  | 100 | 148\% | 130\% | 33 | 148\% | 130\% | 33 | 183\% | 130\% | 42 | 148\% | 130\% | 33 | 182\% | 130\% | 43 | 152\% | 130\% | 17 |
|  | 250 | 130\% | 120\% |  | 130\% | 120\% | 4 | 139\% | 120\% |  | 130\% | 120\% | 4 | 157\% | 120\% | 2 | 124\% | 120\% | , |
| 2035 | 50 | 172\% | 155\% | 89 | 172\% | 155\% | 89 | 212\% | 155\% | 103 | 172\% | 155\% | 89 | 193\% | 145\% | ${ }^{86}$ | 167\% | 127\% | 44 |
|  | 100 | 150\% | 127\% | 44 | 150\% | 127\% | 44 | 188\% | 140\% | 57 | 150\% | 127\% | 44 | 195\% | 140\% | 55 | 161\% | 130\% | 21 |
|  | 250 | 135\% | 120\% | 5 | 135\% | 120\% | 5 | 148\% | 120\% | 8 | 135\% | 120\% | 5 | 162\% | 120\% | 6 | 133\% | 120\% | 0 |

Table 4.14-CROSSING ALTERNATIVES - Highest Maximum NOx Concentrations in Comparison to No Build

| Year | Distance from ROW (m) | Crossing A |  | Crossing B |  | Crossing B |  | Crossing C |  | Crossing C |  | Crossing C |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | From Plaza A |  | From Plaza A |  | From Plaza B1 |  | From Plaza A |  | From Plaza B |  | From Plaza C |  |
|  |  | 1-Hour | Exceedances | 1-Hour | Exceedances | 1-Hour | Exceedances | 1-Hour | Exceedances | 1-Hour | Exceedances | 1-Hour | Exceedances |
| 2015 | 50 | 197\% | 0 | 161\% | 0 | 174\% | 0 | 176\% | 0 | 148\% | 0 | 119\% | 0 |
|  | 100 | 141\% | 0 | 141\% | 0 | 145\% | 0 | 161\% | 0 | 139\% | 0 | 149\% | 0 |
|  | 250 | 128\% | 0 | 128\% | 0 | 114\% | 0 | 140\% | 0 | 117\% | 0 | 121\% | 0 |
| 2025 | 50 | 149\% | 0 | 145\% | 0 | 177\% | 0 | 175\% | 0 | 165\% | 0 | 125\% | 0 |
|  | 100 | 130\% | 0 | 138\% | 0 | 167\% | 0 | 165\% | 0 | 161\% | 0 | 133\% | 0 |
|  | 250 | 124\% | 0 | 130\% | 0 | 158\% | 0 | 158\% | 0 | 124\% | 0 | 117\% | 0 |
| 2035 | 50 | 170\% | 0 | 174\% | 0 | 219\% | 0 | 192\% | 0 | 155\% | 0 | 151\% | 0 |
|  | 100 | 157\% | 0 | 168\% | 0 | 178\% | 0 | 192\% | 0 | 146\% | 0 | 156\% | 0 |
|  | 250 | 147\% | 0 | 147\% | 0 | 153\% | 0 | 197\% | 0 | 142\% | 0 | 139\% | 0 |



The previous chapter presented the air dispersion modeling results for each Access Road, Plaza, and Crossing Alternative studied, and examined the potential changes to air quality in comparison to the No Build Alternative (i.e. doing nothing at all). This section of the report presents the comparative evaluation of the different options and discusses the potential benefits and effects in comparison to one another. Once again, this is done separately for the Access Road, Plaza and Crossing Alternatives.
5.1 Access Road Alternatives

In order to evaluate the potential benefit and effects of each Access Road Alternative and compare to one another, the maximum predicted $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{x}$ concentrations at each distance interval were averaged along the entire route between Grand Marais Road and Howard Avenue. In this manner, the average change in the concentrations compared to No Build could be assessed. These results are presented in Table 5.1 for both $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{\mathrm{x}}$.

The key finding is that implementation of almost any of the Alternatives results in improved air quality on average in comparison to the No-Build option. Some Alternatives and alignments result in more dramatic improvements than others. In general, depressed Alternatives (1B \& 2B) result in lower concentrations and fewer exceedances of criteria on average than at-grade Alternatives ( $1 \mathrm{~A} \& 2 \mathrm{~A}$ ). A tunneled Alternative with a vent building (Alternative 3) results in the greatest reduction in $\mathrm{PM}_{2.5}$ concentrations, but the lowest reduction in $\mathrm{NO}_{\mathrm{x}}$ concentrations. As mentioned previously, the Jet Fans tunnel ventilation option typically resulted in unacceptable concentrations of $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{\mathrm{x}}$, and frequently exceeded the relevant criteria by a significant amount, and thus was not considered further in this assessment.

Table 5.1 shows that all alternatives result in lower concentrations and number of exceedances on average in comparison with the No Build scenario. The depressed options consistently result in slightly lower $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{\mathrm{x}}$ concentrations. Also, the reduction in number of exceedances of the $\mathrm{PM}_{2.5}$ criterion is greater for the depressed options than for the at-grade alternatives. These results are discussed in further detail in the following sections.
5.1.1 Comparison of At Grade, Below Grade \& Cut \& Cover Tunnel Alternatives

As discussed earlier, this segment encompasses various different types of roadway links, and thus was split into two separate sub-segments. Overall, this segment has the highest concentrations of $\mathrm{NO}_{\mathrm{x}}$ and $\mathrm{PM}_{2.5}$ in the future, and thus has the poorest air quality.
5.1.1.1

At Grade versus Below Grade Alternatives

The effect of depressing the roadway is discussed and examined in this section, through the comparison of Alternative 1A to 1B, and of 2A to 2B. As can be seen in Table 5.1, comparing the relative $\mathrm{PM}_{2.5}$ concentrations between 1 A and 1 B , Alternative 1 B (depressed) results in marginally lower concentrations (relative to No Build) at 50 m from the roadway. The difference is greater than $10 \%$ in 2035, and thus is considered to be significant. Similarly, Alternative 1B results in a greater reduction in the number of days predicted to be greater than the CWS. However, this effect is limited to approximately 50 m of the ROW. At 100 m from the ROW there is no difference between Alternative 1 A and 1 B , and no difference between implementation of either Alternative 1A or 1B and No Build.

A similar trend is seen in the comparison of Alternative 2A versus 2B. The $\mathrm{PM}_{2.5}$ concentrations at 50 m away are marginally lower in 2015 and 2025, and significantly lower in 2035 for the depressed roadway. Also, there is a greater reduction in the number of days predicted to exceed the CWS at 50 m away from the roadway. This reduction is significant in 2035, but is once again limited to between 50 and 100 m from the ROW.

The annual average concentrations do not exceed the criterion on average for any of the alternatives examined, in any of the horizon years.

In terms of NOx concentrations, there are no predicted exceedances of the MOE 1-hour NOx criterion for any of Alternatives $1 \mathrm{~A}, 1 \mathrm{~B}, 2 \mathrm{~A}$, or 2 B at any of the distance intervals studied. As mentioned previously, implementation of any of these alternatives result in significant (i.e. > 10\%) decreases in the maximum predicted concentrations, relative to No build. The depressed options (Alternative 1 B and 2B) show marginal decreases in the relative $\mathrm{NO}_{\mathrm{x}}$ concentrations in comparison to the at-grade options. However, these decreases are not significant (i.e. $<10 \%$ ) between Alternative 1A and 1B or 2A and 2B.

### 5.1.1.2 At Grade versus Tunnel Alternatives

In this section of the report, the effect of tunneling the roadway is examined in comparison to an at grade roadway. This will be done yia a comparison of the results along the route between Alternative 1A and 3, as well as 2A to 3 .

Comparing the results presented in Table 5.1 for Alternatives 1 A and 3 show that a tunneled alternative would result in significant reductions in the maximum $\mathrm{PM}_{2.5}$ concentrations at 50 m from the ROW in all horizon years examined. This is true for comparisons of Alternative 3 to both Alternative 1A and 2A. Also, in comparison to Alternative 1 A there is a significant reduction (i.e. $>8$ ) in the number of days predicted to exceed the CWS at 50 m away for a tunneled access road in comparison to an at-grade roadway in 2025. This difference becomes significant at 100 m away in 2035 . However, these differences are less when Alternative 3 is compared to Alternative 2A, and is only significant at 50 m away in 2025 and 2035.

The annual average concentrations do not exceed the criterion on average for any of the alternatives examined, in any of the horizon years.

With respect to the maximum predicted 1-hour NOx concentrations, there are no predicted exceedances of the MOE 1-hour NOx criterion for any of the at-grade or tunneled Alternatives examined. Comparing the relative magnitude of the maximum predicted concentrations between 1 A and 3 shows that there is no difference at any of the distance intervals, in any of the horizon years. However, a comparison between Alternative 2A and 3 indicates that a tunneled alternative increases the maximum predicted concentrations over an at-grade access road with 2-way service roads at 50 m from the ROW. However, this difference is significant only in the year 2015. The increases in all other years are marginal.

Based on these results, the effect of tunneling the roadway (either positive or negative) does not extend beyond a maximum of 100 m away in comparison to at grade Alternatives.

### 5.1.1.3

Below Grade versus Tunnel Alternatives

This evaluation examines differences between below grade (depressed) alternatives and the tunneled alternative. This will be done through a comparison of Alternative 1B to 3 and Alternative 2B to 3 .

The results presented in Table 5.1 show that there are significant differences (i.e. $>10 \%$ ) in the relative maximum $\mathrm{PM}_{2.5}$ concentrations between the depressed alternatives (1B and 2B) in comparison to the tunneled alternative (3). Also, when compared to both Alternatives 1 B and 2B, a tunneled alternative would result in reductions in the number of days predicted to exceed the CWS. However, the reductions are only significant (i.e. >8) at 50 m from the ROW in 2035. The differences in all previous years and at other distance intervals are marginal.

The $\mathrm{NO}_{\mathrm{x}}$ results are similar to what was found when the at-grade alternatives were compared to a tunneled alternative. There are no predicted exceedances of the MOE 1-hour NOx criterion for any of the depressed or tunneled Alternatives. In comparing the relative magnitude of the maximum predicted $\mathrm{NO}_{x}$ concentrations between 1 B and 3 shows that there is no difference at any of the distance intervals, in any of the horizon years ${ }^{1}$. However, a comparison between Alternative 2 B and 3indicates that a tunneled alternative increases the maximum predicted concentrations over an at-grade access road with 2-way service roads at 50 m from the ROW. However, this difference is significant only in the year 2015. The increases in all other years are marginal.

Based on the above comparisons, the effect of tunneling the roadway (either positive or negative) is limited to within $50-100 \mathrm{~m}$ from the roadway in comparison to below grade (depressed) Alternatives.

### 5.1.2 Service Road Configurations

As part of the assessment, two separate configurations of freeway service roads were studied. These included one-way service roads on either side of the freeway, and two way service roads located approximately on the existing Highway 3 / Huron Church Road alignment. The differences between these configurations will be evaluated through comparisons between Alternatives 1A and 2 A , as well as 1 B and 2 B .

Comparison of the $\mathrm{PM}_{2.5}$ data between the at-grade alternatives shows that the two way service road configurations (Alternative 2) results in marginally lower maximum $\mathrm{PM}_{2.5}$ concentrations, in comparison to the one-way service road configurations (Alternative 1). The difference is significant (> 10\%) for the

[^0]Option 2 alignment in 2015. Also, the two-way service road alignments result in reductions in the number of days predicted to be greater than the CWS. These differences are significant at 50 m from the ROW in 2035.

The comparison between the $\mathrm{PM}_{2.5}$ results for the below grade options shows a slightly different trend. For these options, although there is a marginal reduction in the maximum $\mathrm{PM}_{2.5}$ concentrations between the one and two-way service roads, and slight reductions in the number of days predicted to exceed the CWS, none of these differences are significant. This is true at all distance intervals and all horizon years studied. However, the two-way service road configuration does result in a reduction in the maximum predicted 1-hour NOx concentrations at 50 m away from the roadway in 2015 for both the at grade and below grade options. The reductions are marginal in 2025 and 2035.

Thus, the results indicate that the two-way service road configurations result in lower maximum $\mathrm{PM}_{2.5}$ and $\mathrm{NO}_{x}$ concentrations, and fewer days that are predicted to exceed the CWS. However, this effect is limited to less than 100 m away from the ROW, and is more pronounced for at-grade alternatives than below grade alternatives.

### 5.1.3

Route Alignments Between St.Clair College \& Howard Avenue

As outlined previously, two separate route alignment options were studied in the area between St.Clair College and Howard Avenue. The first route alignment (Option 1), realigns the existing Talbot Road / Highway 3 corridor slightly to the northeast. This realignment begins at approximately at Howard Avenue and continues approximately to the entrance to St.Clair College.

The Option 2 alignment utilizes the existing Talbot Road / Highway 3 corridor as local access service roads without any realignment and aligns the freeway to the southeast.

In order to evaluate whether there are any differences between the two alignments, the Option 1 and Option 2 results will be compared to one another for each alternative. This will be done separately for the at-grade and below grade alternatives.

### 5.1.3.1 $\mid$ At Grade Alternatives

The $\mathrm{PM}_{2.5}$ results from Alternative 1 A show that the maximum predicted concentrations are similar for both Option 1 and Option 2 at 50 m away in all horizon years. The number of days predicted to exceed the CWS is reduced for the Option 2 alignment at 50 m away by 2025. However, this difference is not significant until 2035. This same trend is seen in the results for Alternative 2A.

The Option 2 alignment also results in reduced maximum predicted 1-hour NOx concentrations. However, these reductions are only significant for Alternative 1A at 50 m from the ROW in 2015 .
5.1.3.2

Below Grade Alternatives

Similar to the results for the at-grade alternatives, the Option 2 alignment results in slightly lower maximum $\mathrm{PM}_{2.5}$ concentrations than the Option 1 alignment for the below grade alternatives. However, none of these differences are significant at any distance interval, in any of the years examined. This is also true for the change in the number of predicted CWS exceedance days.

The $\mathrm{NO}_{\mathrm{y}}$ results for the below grade alternatives exhibit the identical trend seen for the at-grade alternatives. The Option 2 alignment results in reduced maximum predicted 1-hour NOx concentrations. However, these reductions are only significant for Alternative 1B at 50 m from the ROW in 2015.

The results outlined above indicate that in comparison to the Option 1 route alignment, the Option 2 alignment results in reduced $\mathrm{PM}_{2.5}$ concentrations for atgrade alternatives, and reduced $\mathrm{NO}_{\mathrm{x}}$ concentrations for both at-grade or depressed options. However the differences between the Options are limited to distances less than 100 m away from the ROW.

### 5.2 Evaluation of Plaza Alternatives

The dispersion model results presented previously for each of the four plaza alternatives were used to complete a comparative evaluation of the different plaza options. This evaluation is presented in Table 5.2.

### 5.2.1

## PM2.5 Concentrations

As can be seen in Table 5.2, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations at 50 m away from the property boundary increase by a factor of 2 X to more than 3 X the No Build concentrations in each of the horizon years for all four plaza options. The changes at all distance intervals from the boundary were shown earlier in Table 4.11, and are significant at 250 m for all Plaza Alternatives and all horizon years. Similarly, all of the Plaza Alternatives result in a significant increase the number of days predicted to exceed the CWS at 100 m away, in comparison to No Build.

All of the plaza alternatives have a significant impact on the air quality in the immediate vicinity of the properties. In 2015 and 2025, this distance is approximately 100 m from the property boundary, but increases to 250 m from the boundary in 2035 for Plaza B, B1 and C.

The largest difference of any alternatives (i.e. highest increase) is seen in the vicinity of Plaza B1 in 2035. Plaza B1 also has the largest increase in number of days predicted to exceed the CWS within 100 m of the Plaza boundary.

The lowest concentrations and lowest change in the number of days predicted to exceed the CWS is seen in the vicinity of Plaza C. This is likely due to the arrangement of the roadways within the property footprint, which results in an additional buffer between the emission sources (i.e. cars and trucks) and the property boundary.

### 5.2.2 $\quad \mathrm{NO}_{x}$ Concentrations

All of the plaza alternatives have a significant impact on the air quality in the immediate vicinity of the property boundaries. The maximum predicted 1-hour $\mathrm{NO}_{\mathrm{X}}$ concentrations at 50 m away from the property boundary increase by as much as a factor of 6X in 2015, 9X in 2025 and almost 11X in 2035, in comparison to the No Build concentrations for all four plaza options. The increases in concentration are significant at distances up to 250 m from the property boundary, for all Plaza Alternatives, and all horizon years.

Plaza A and B also show significant increases in the number of hours when the MOE AAQC for NOx is predicted to be exceeded at 50 m away in 2025. The NOx criterion is not exceeded at Plaza B1 or Plaza C at any of the distance intervals in any of the horizon years.

Plaza B results in the highest increase in maximum predicted concentrations and the largest increase in the number of exceedances of the $\mathrm{NO}_{\mathrm{x}}$ criterion at distances up to 100 m from the property boundary in 2025 and 2035.

The lowest concentrations and lowest change in the number of days predicted to exceed the $\mathrm{NO}_{\mathrm{x}}$ criterion is seen in the vicinity of Plaza C . As mentioned previously, this is likely due to an additional buffer between the vehicles and the property boundary, because of the facility layout.

## Evaluation of Crossing Alternatives

The potential effects of the Crossing Alternatives are quite different from those of the Plaza Alternatives, and thus have been presented separately. The evaluation is also presented in Table 5.2. Values that pertain to the referenced crossing, rather than the plaza, are shown in brackets below the plaza value.

The results from the Crossing Alternatives / Connecting Roadways are similar to those seen for the Access Road Alternatives. However, the Crossings and connecting roads are elevated, and thus the emissions behave a little differently than ground level sources.

### 5.3.1 $\quad \mathrm{PM}_{2.5}$ Concentrations

As can be seen in Table 5.2, the maximum predicted $\mathrm{PM}_{2.5}$ concentrations at 50 m away from the ROW of the Crossings and connecting roadways increase by a factor of 1.5-2X the No Build concentrations in each of the horizon years for the three crossing options. Table 4.13 presented the changes at all distance intervals from the boundary, and are significant at 250 m for all Crossing Alternatives and all horizon years. All of the Crossing Alternatives result in a significant increase the number of days predicted to exceed the CWS at 100 m away, in comparison to No Build.

Crossing B combined with Plaza B1 results in the largest difference of any of the alternatives (i.e. highest increase) in 2035. This is likely due to the "spillover effect" of emissions from Plaza B1, as this effect is not seen when Crossing B is
combined with Plaza A. This crossing alternative also has the largest increase in number of days predicted to exceed the CWS within 50-100 m of the ROW.

The lowest concentrations and lowest change in the number of days predicted to exceed the CWS is seen in the vicinity of Crossing C, when combined with Plaza C. The difference in the maximum predicted concentrations, and number of days predicted to be in excess of the CWS is significantly lower than all other Plaza/Crossing combinations.

Based on the above evaluation, the impact of the Crossings / Connecting Roadway is limited to approximately 100 m from the ROW.

NOx Concentrations
The maximum predicted 1 -hour $\mathrm{NO}_{\mathrm{x}}$ concentrations increase by more than a factor of 2 X in comparison to the No Build concentrations at 50 m away from the ROW by 2035. The increases in concentration are significant (>10\%) at distances up to 250 m away, for all Alternatives, and all horizon years. Similar to the results of the Access Road Alternatives, there are no exceedances of the MOE 1 -hour NOx criterion in the vicinity of the Crossings / Connecting Roadway.

The largest increase in the maximum predicted concentrations is seen in the vicinity of Crossing B from Plaza B1, at all distance intervals studied in all years. Conversely, the lowest increases in concentrations are seen in the vicinity of Crossing C from Plaza C.

Table 5.1 Access Road Evaluation Table


Table 5.2 Plaza and Crossings Evaluation Table



### 7.0 ReFERENCES

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Table A-1 Hourly Traffic profiles used in Modelling

| Period Starting | Profile 1 |  | Profile 2 |  | Profile 3 |  | Profile 4 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Outbound | Inbound | Outbound | Inbound | Outbound | Inbound | Outbound | Inbound |
| 12:00 AM | 47 | 27 | 22 | 22 | 8 | 9 | 29 | 20 |
| 01:00 AM | 33 | 21 | 17 | 16 | 4 | 4 | 26 | 15 |
| 02:00 AM | 33 | 19 | 14 | 15 | 3 | 3 | 24 | 14 |
| 03:00 AM | 32 | 19 | 14 | 12 | 2 | 2 | 26 | 11 |
| 04:00 AM | 41 | 18 | 18 | 12 | 2 | 3 | 34 | 12 |
| 05:00 AM | 65 | 19 | 37 | 16 | 8 | 8 | 54 | 14 |
| 06:00 AM | 135 | 28 | 92 | 29 | 29 | 21 | 114 | 24 |
| 07:00 AM | 157 | 30 | 124 | 46 | 50 | 43 | 152 | 34 |
| 08:00 AM | 175 | 38 | 149 | 53 | 81 | 88 | 139 | 53 |
| 09:00 AM | 141 | 43 | 103 | 44 | 57 | 68 | 102 | 52 |
| 10:00 AM | 114 | 48 | 82 | 46 | 67 | 68 | 100 | 56 |
| 11:00 AM | 111 | 57 | 85 | 56 | 81 | 80 | 99 | 63 |
| 12:00 PM | 112 | 58 | 87 | 58 | 81 | 79 | 100 | 64 |
| 01:00 PM | 114 | 61 | 85 | 59 | 82 | 74 | 96 | 65 |
| 02:00 PM | 117 | 69 | 95 | 68 | 89 | 84 | 102 | 74 |
| 03:00 PM | 108 | 88 | 104 | 94 | 102 | 95 | 111 | 89 |
| 04:00 PM | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| 05:00 PM | 113 | 96 | 99 | 100 | 99 | 102 | 94 | 110 |
| 06:00 PM | 116 | 82 | 92 | 75 | 91 | 96 | 98 | 98 |
| 07:00 PM | 86 | 65 | 71 | 58 | 73 | 71 | 78 | 79 |
| 08:00 PM | 94 | 67 | 63 | 59 | 61 | 58 | 74 | 65 |
| 09:00 PM | 84 | 57 | 53 | 50 | 50 | 40 | 64 | 51 |
| 10:00 PM | 75 | 48 | 50 | 42 | 27 | 29 | 52 | 44 |
| 11:00 PM | 62 | 38 | 37 | 36 | 18 | 18 | 39 | 35 |

Profile 1: Huron Church North of E.C. Row in Base Cases
Profile 2: Huron Church South of E.C. Row in Base Cases and Freeway in Alternatives 1-3
Profile 3: All other Streets in Base Cases and Alternatives 1-3
Profile 4: Huron Church North of E.C. Row and E.C. Row in Alternatives 1-3

Profiles have been standardized to modelled p.m. peak hour 4:00 to 5:00 p.m.

Table A-1 Contd.

| Profile 1 | Profile 2 | Profile 3 | Profile 4 |
| :---: | :---: | :---: | :---: |
| 0.021 | 0.015 | 0.007 | 0.016 |
| 0.016 | 0.011 | 0.003 | 0.013 |
| 0.015 | 0.010 | 0.002 | 0.012 |
| 0.015 | 0.009 | 0.002 | 0.012 |
| 0.017 | 0.010 | 0.002 | 0.014 |
| 0.024 | 0.018 | 0.006 | 0.022 |
| 0.047 | 0.042 | 0.020 | 0.044 |
| 0.054 | 0.060 | 0.037 | 0.059 |
| 0.061 | 0.071 | 0.068 | 0.061 |
| 0.053 | 0.052 | 0.050 | 0.049 |
| 0.047 | 0.045 | 0.054 | 0.050 |
| 0.048 | 0.049 | 0.064 | 0.051 |
| 0.049 | 0.051 | 0.064 | 0.052 |
| 0.050 | 0.050 | 0.062 | 0.051 |
| 0.054 | 0.057 | 0.069 | 0.056 |
| 0.057 | 0.070 | 0.078 | 0.064 |
| 0.058 | 0.070 | 0.080 | 0.063 |
| 0.061 | 0.069 | 0.080 | 0.065 |
| 0.057 | 0.058 | 0.074 | 0.062 |
| 0.044 | 0.045 | 0.057 | 0.050 |
| 0.047 | 0.043 | 0.047 | 0.044 |
| 0.041 | 0.036 | 0.036 | 0.037 |
| 0.036 | 0.032 | 0.022 | 0.030 |
| 0.029 | 0.025 | 0.014 | 0.023 |

Table A-2 24-Hour Annual Average Daily Traffic (AADT) for Alternative 1A - Year 2015




*     - For consistency, Huron Church Rd/Talbot Rd runs North-South and all crossing roads run East-West


| LOCATION | SECTION |  | $\begin{gathered} \hline \text { Alternative 1A } \\ \hline 2035 \end{gathered}$ |  |  |  | 24 Hour AADT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Local Cars |  | Local Trucks |  | International Cars |  | International Trucks |  |
|  |  |  | AM PEAK HOUR | PM PEAK HOUR |  |  |  |  |  |
|  | FROM | T0 |  |  | NB | SB | NB | SB | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB/EB |
| HC Road | Riverside | University |  |  |  |  |  |  | 6681 | 5506 | 199 | 95 | 3 | 1 | 0 | 81 |
|  | University | Wyandotte |  |  |  |  | 2848 | 3663 | 92 | 125 | 69 | 237 | 62 | 81 |
|  | Wyandotte | AMB Off Ramp |  |  |  |  | 2062 | 3072 | 0 | 0 | 47 | 168 | 0 | 0 |
|  | AMB Off Ramp | College |  |  |  |  | 7913 | 6352 | 238 | 107 | 7713 | 1 | 3781 | 0 |
|  | College St | Girardot St | 1846 | 887 | 1546 | 1845 | 18212 | 16345 | 580 | 514 | 7061 | 5277 | 306 | 3756 |
|  | Girardot St | Tecumseh Rd | 1743 | 853 | 1449 | 1695 | 18263 | 17708 | 712 | 614 | 6376 | 4624 | 287 | 3524 |
|  | Tecumseh Rd | Dorchester St | 1782 | 1143 | 1806 | 1846 | 21568 | 21530 | 867 | 768 | 6028 | 3842 | 287 | 3124 |
|  | Dorchester St | Prince Rd/Totten St | 1904 | 1184 | 1737 | 1903 | 22344 | 23275 | 763 | 691 | 5483 | 3517 | 244 | 2844 |
|  | Prince Rd/Totten St | Malden Rd | 2068 | 1458 | 1974 | 2175 | 25100 | 27702 | 860 | 830 | 5462 | 3268 | 278 | 2696 |
|  | Malden Rd | Industrial Rd | 1728 | 1194 | 1509 | 1841 | 19140 | 23044 | 648 | 684 | 5492 | 3421 | 10 | 2762 |
|  | Industrial Rd | EC Row N. Ramp Terminal | 1915 | 1277 | 1704 | 2060 | 22486 | 25804 | 740 | 759 | 5337 | 3536 | 0 | 2728 |
|  | EC Row N. Ramp Termina | EC Row S. Ramp Terminal | 1450 | 1725 | 1225 | 2448 | 16155 | 34765 | 488 | 890 | 4690 | 3428 | 0 | 2489 |
|  | S. of EC Row S. Ramp Terminal |  | 2046 | 1344 | 1886 | 1883 | 24303 | 26445 | 722 | 568 | 4907 | 3037 | 0 | 2010 |
| S Service Rd | N. of Lambton St |  | n/a | 830 | n/a | 826 | n/a | 13352 | $\mathrm{n} / \mathrm{a}$ | 204 | n/a | 867 | n/a | 0 |
|  | Lambton St | Todd Ln/Cabana Rd | n/a | 788 | n/a | 739 | n/a | 12677 | n/a | 155 | n/a | 530 | n/a | 0 |
|  | Todd Ln/Cabana Rd | St Clair College | n/a | 1006 | n/a | 1082 | n/a | 18273 | n/a | 98 | n/a | 0 | n/a | 0 |
|  | St Clair College | Cousineau Dr | n/a | 418 | n/a | 846 | n/a | 8674 | n/a | 91 | n/a | 1745 | n/a | 143 |
|  | Cousineau Dr | Howard Ave | n/a | 359 | n/a | 394 | n/a | 5944 | n/a | 77 | n/a | 426 | n/a | 136 |
|  | E. of Howard Ave |  | n/a | 922 | n/a | 793 | n/a | 13828 | n/a | 278 | n/a | 0 | n/a | 0 |
| N Service Rd | N. of Labelle St |  | 2046 | n/a | 1886 | n/a | 25012 | n/a | 618 | n/a | 4799 | n/a | 0 | n/a |
|  | Labelle St | Grand Marais Rd Ramp | 1864 | n/a | 1877 | n/a | 27023 | n/a | 408 | n/a | 2258 | n/a | 0 | n/a |
|  | Grand Marais Rd Ramp | Pulford St | 1058 | n/a | 813 | n/a | 13112 | n/a | 201 | n/a | 1256 | n/a | 0 | n/a |
|  | Pulford St | Todd Ln/Cabana Rd | 1076 | n/a | 868 | n/a | 13747 | n/a | 223 | n/a | 1227 | n/a | 0 | n/a |
|  | Todd Ln/Cabana Rd | St Clair College | 322 | n/a | 564 | n/a | 7428 | n/a | 0 | n/a | 0 | n/a | 0 | n/a |
|  | St Clair College | Cousineau Dr | 1085 | n/a | 941 | n/a | 11970 | n/a | 87 | n/a | 3222 | n/a | 0 | n/a |
|  | Cousineau Dr | Howard Ave | 368 | n/a | 614 | n/a | 7331 | n/a | 246 | n/a | 434 | n/a | 0 | n/2 |
|  | E. of Howard Ave |  | 705 | n/a | 1144 | n/a | 15128 | n/a | 344 | n/a | 0 | n/a | 0 | n/a |
| Ojibway Pwy | EC Row Expressway | GN Booth Dr | 735 | 470 | 680 | 860 | 11387 | 10979 | 144 | 129 | 25 | 19 | 140 | 648 |
|  | GN Booth Dr | Sandwich St | 720 | 473 | 686 | 833 | 11322 | 10778 | 143 | 127 | 26 | 19 | 141 | 652 |
|  | Sandwich St | Prospect Ave | 679 | 425 | 633 | 793 | 10666 | 10473 | 74 | 72 | 50 | 44 | 0 | 0 |
|  | N. of Prospect Ave |  | 672 | 415 | 632 | 780 | 10603 | 10274 | 74 | 71 | 49 | 43 | 0 | 0 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| CROSSING ROADS |  |  | WB | EB | WB | EB |  |  |  |  |  |  |  |  |
| Wyandotte | W of HuronChurch |  |  |  |  |  | 4604 | 4465 | 0 | 0 | 381 | 446 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2776 | 4327 | 17 | 154 | 771 | 942 | 58 | 0 |
| University | W of HuronChurch |  |  |  |  |  | 1547 | 1310 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2225 | 2085 | 125 | 92 | 68 | 22 | 81 | 62 |
| Riverside | W of HuronChurch |  |  |  |  |  | 3717 | 4039 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 7009 | 5930 | 0 | 0 | 169 | 47 | 0 | 0 |
| AMB Off Ramp | E of HuronChurch |  |  |  |  |  | 0 | 931 | 0 | 43 | 0 | 7710 | 0 | 3781 |
| AMB On Ramp | E of HuronChurch |  |  |  |  |  | 265 | 0 | 7 | 0 | 6598 | 0 | 273 | 0 |
| Patricia | AMB | Wyandotte |  |  |  |  | 469 | 928 | 14 | 41 | 3573 | 3465 | 233 | 394 |
| College St | E. of HC Road |  | 319 | 351 | 490 | 406 | 6584 | 5577 | 171 | 131 | 3 | 582 | 0 | 296 |
|  | W. of HC Road |  | 90 | 52 | 187 | 79 | 1746 | 1039 | 0 | 0 | 529 | 39 | 0 | 0 |
| Girardot St | E. of HC Road |  | 59 | 83 | 84 | 48 | 973 | 1167 | 0 | 0 | 182 | 0 | 0 | 0 |
|  | W. of HC Road |  | 85 | 146 | 180 | 120 | 2213 | 2122 | 36 | 22 | 42 | 27 | 0 | 0 |
| Tecumseh Rd | E. of HC Road |  | 332 | 329 | 420 | 509 | 5868 | 6355 | 135 | 146 | 206 | 450 | 0 | 342 |
|  | W. of HC Road |  | 252 | 505 | 527 | 395 | 6315 | 7254 | 0 | 0 | 362 | 99 | 0 | 0 |
| Dorchester St | E. of HC Road |  | 78 | 86 | 135 | 96 | 1466 | 1600 | 0 | 0 | 274 | 0 | 0 | 0 |
|  | W. of HC Road |  | 76 | 46 | 86 | 52 | 1375 | 790 | 22 | 9 | 22 | 9 | 0 | 0 |
| Prince Rd/Totten St | E. of HC Road |  | 154 | 126 | 129 | 207 | 2228 | 2764 | 0 | 0 | 83 | 124 | 0 | 0 |
|  | W. of HC Road |  | 241 | 329 | 338 | 365 | 4992 | 5628 | 0 | 0 | 62 | 83 | 0 | 0 |
| Malden Rd | E. of HC Road |  | 117 | 71 | 113 | 111 | 1546 | 1009 | 0 | 0 | 313 | 503 | 0 | 0 |
|  | W. of HC Road |  | 469 | 573 | 506 | 527 | 7608 | 7977 | 399 | 391 | 380 | 44 | 234 | 890 |
| Industrial Rd | E. of HC Road |  | 305 | 161 | 225 | 285 | 3605 | 3587 | 46 | 57 | 702 | 191 | 8 | 27 |
|  | W. of HC Road |  | 307 | 114 | 198 | 303 | 4307 | 3085 | 183 | 210 | 0 | 0 | 0 | 308 |
| EC Row N. Ramp Terminal | E. of HC Road (W-N/S Off | Ramp \& N-W On Ramp) | 1060 | 113 | 1020 | 140 | 14587 | 1944 | 376 | 2 | 1534 | 267 | 0 | 0 |
|  | W. of HC Road (S-W On R | amp) | 34 | n/a | 33 | n/a | 387 | n/a | 11 | n/a | 45 | n/a | 182 | n/a |
| EC Row S. Ramp Terminal | E. of HC Road (N-E On Ra | mp) | n/a | 700 | n/a | 756 | n/a | 11821 | n/a | 230 | n/a | 0 | n/a | 0 |
|  | W. of HC Road (S-E On Ra | mp \& E-N/S Off Ramp) | 446 | 169 | 698 | 228 | 8584 | 2930 | 312 | 86 | 541 | 301 | 585 | 0 |
| Labelle St | E. of N. Service Rd |  | 229 | 47 | 53 | 44 | 1900 | 803 | 0 | 0 | 212 | 0 | 0 | 0 |
| Grand Marais Rd Ramp | E. of N. Service Rd |  | 498 | 96 | 246 | 168 | 5666 | 2026 | 0 | 0 | 245 | 228 | 0 | 0 |
| Fazio Dr | W. of S. Service Rd |  | 216 | 178 | 324 | 242 | 4549 | 3403 | 74 | 39 | 79 | 39 | 0 | 0 |
| Pulford St | E. of N. Service Rd |  | 47 | 65 | 17 | 72 | 401 | 1205 | 0 | 0 | 76 | 0 | 0 | 0 |
| Todd Ln/Cabana Rd | E. of N. Service Rd |  | 617 | 406 | 696 | 540 | 10749 | 8206 | 0 | 0 | 86 | 65 | 0 | 0 |
|  | between N. and S. Service R |  | 552 | 1125 | 816 | 964 | 11958 | 15840 | 0 | 146 | 0 | 777 | 0 | 0 |
|  | between S. Service Rd and | Huron Church Line | 930 | 1238 | 1167 | 992 | 17887 | 17167 | 103 | 140 | 333 | 626 | 0 | 0 |
|  | W. of Huron Church Line |  | 506 | 534 | 557 | 628 | 9340 | 9633 | 0 | 0 | 7 | 5 | 0 | 0 |
| St Clair College | E. of N. Service Rd |  | 142 | 844 | 265 | 170 | 3287 | 8894 | 0 | 0 | 101 | 276 | 0 | 0 |
|  | between N. and S. Service R |  | 96 | 306 | 216 | 74 | 2706 | 2761 | 0 | 0 | 0 | 201 | 0 | 0 |
| Cousineau Dr | E. of N. Service Rd |  | 287 | 242 | 372 | 257 | 4052 | 3386 | 0 | 0 | 1456 | 993 | 0 | 0 |
|  | between N. and S. Service R |  | 361 | 458 | 740 | 499 | 6330 | 7763 | 7 | 160 | 2824 | 0 | 0 | 0 |
|  | W. of S. Service Rd |  | 364 | 539 | 1051 | 438 | 10109 | 7212 | 25 | 153 | 1974 | 738 | 0 | 0 |
| Howard Ave | E. of N. Service Rd |  | 493 | 482 | 559 | 768 | 9092 | 10245 | 151 | 202 | 3 | 6 | 0 | 0 |
|  | between N. and S. Service R |  | 794 | 451 | 863 | 543 | 13254 | 8541 | 226 | 182 | 150 | 2 | 0 | 0 |
|  | W. of S. Service Rd |  | 733 | 953 | 900 | 980 | 13154 | 16611 | 281 | 354 | 82 | 49 | 0 | 0 |
| EC Row Expressway | E. of Ojibway Pwy |  | 1335 | 480 | 1300 | 640 | 20974 | 9441 | 499 | 250 | 191 | 84 | 140 | 18 |
|  | W. of Ojibway Pwy |  | 1585 | 835 | 1910 | 780 | 24225 | 13029 | 311 | 285 | 6091 | 0 | 150 | 0 |
|  | E. of Huron Church Rd |  | 3237 | 2884 | 2812 | 3676 | 43009 | 48627 | 905 | 1071 | 3558 | 5208 | 688 | 2127 |
|  | At Malden Rd |  | 2324 | 1907 | 1965 | 2450 | 30637 | 31372 | 566 | 596 | 2197 | 4367 | 917 | 1336 |
|  | W. of Matchette |  | 1676 | 480 | 1433 | 640 | 21519 | 9441 | 512 | 250 | 1921 | 84 | 832 | 18 |
|  <br> GN Booth Dr <br> Sandwich St <br> Prospect Ave | W. of Ojibway Pwy |  | 27 | 10 | 13 | 44 | 347 | 450 | 6 | 6 | 4 | 4 | 0 | 0 |
|  | W. of Ojibway Pwy |  | 82 | 89 | 121 | 107 | 1594 | 1496 | 157 | 102 | 25 | 33 | 0 | 0 |
|  | W. of Ojibway Pwy | $\overline{\mathrm{ds} r}$ | 29 | 33 | 9 | 21 | 333 | 427 | 5 | 4 | 3 | 6 | 0 | 0 |

[^1]


| LOCATION | SECTION |  | 2015 |  |  |  | 24 Hour AADT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Local Cars |  | Local Trucks |  | International Cars |  | International Trucks |  |
|  |  |  | AM PEAK HOUR | PM PEAK HOUR |  |  |  |  |  |
|  | FROM | T0 |  |  | NB | SB | NB | SB | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB/EB |
| HC Road | Riverside | University |  |  |  |  |  |  | 6735 | 5369 | 173 | 84 | 3 | 1 | 0 | 3 |
|  | University | Wyandotte |  |  |  |  | 3090 | 3626 | 91 | 118 | 58 | 242 | 20 | 3 |
|  | Wyandotte | AMB Off Ramp |  |  |  |  | 2285 | 3005 | 0 | 0 | 37 | 172 | 0 | 0 |
|  | AMB Off Ramp | College |  |  |  |  | 8617 | 6228 | 229 | 94 | 6211 | 1 | 2391 | 0 |
|  | College St | Girardot St | 1791 | 824 | 1486 | 1710 | 18255 | 16675 | 543 | 487 | 6361 | 4419 | 205 | 2512 |
|  | Girardot St | Tecumseh Rd | 1719 | 789 | 1325 | 1545 | 17763 | 17139 | 636 | 571 | 5615 | 3744 | 182 | 2347 |
|  | Tecumseh Rd | Dorchester St | 1788 | 1072 | 1668 | 1709 | 21118 | 21024 | 778 | 716 | 5388 | 3342 | 186 | 2185 |
|  | Dorchester St | Prince Rd/Totten St | 1891 | 1128 | 1608 | 1770 | 21714 | 22815 | 693 | 656 | 4945 | 3065 | 159 | 1997 |
|  | Prince Rd/Totten St | Malden Rd | 2033 | 1334 | 1827 | 1951 | 24278 | 26074 | 777 | 757 | 4852 | 3020 | 180 | 1940 |
|  | Malden Rd | Industrial Rd | 1668 | 1133 | 1473 | 1575 | 19251 | 21200 | 577 | 580 | 4904 | 3155 | 7 | 1837 |
|  | Industrial Rd | EC Row N. Ramp Termina | 1821 | 1192 | 1607 | 1726 | 21772 | 23501 | 662 | 652 | 4735 | 2918 | 0 | 1845 |
|  | EC Row N. Ramp Termin | EC Row S. Ramp Termina | 1342 | 1491 | 1172 | 2010 | 15712 | 30308 | 416 | 646 | 4099 | 2573 | 0 | 1516 |
|  | S. of EC Row S. Ramp Terminal |  | 1715 | 1187 | 1531 | 1500 | 20499 | 23057 | 528 | 389 | 3776 | 2217 | 0 | 1283 |
| S Service Rd | N. of Bethlehem Ave |  | n/a | 1187 | n/a | 1501 | n/a | 19685 | n/a | 343 | n/a | 2215 | n/a | 1249 |
|  | Bethlehem Ave | Lambton St | n/a | 294 | n/a | 304 | n/a | 4783 | n/a | 124 | n/a | 313 | n/a | 0 |
|  | Lambton St | Pulford St | n/a | 350 | n/a | 224 | n/a | 4812 | n/a | 29 | n/a | 222 | n/a | 0 |
|  | Pulford St | Todd Ln/Cabana Rd | n/a | 662 | n/a | 684 | n/a | 11759 | n/a | 94 | n/a | 0 | n/a | 0 |
|  | Todd Ln/Cabana Rd | Huron Church Line | n/a | 643 | n/a | 792 | n/a | 12007 | n/a | 112 | n/a | 385 | n/a | 0 |
|  | Huron Church Line | St Clair College | n/a | 876 | n/a | 1006 | n/a | 16439 | n/a | 97 | n/a | 0 | n/a | 0 |
|  | St Clair College | Cousineau Dr | n/a | 349 | n/a | 753 | n/a | 7717 | n/a | 104 | n/a | 1359 | n/a | 154 |
|  | Cousineau Dr | Howard Ave | n/a | 349 | n/a | 369 | n/a | 5460 | n/a | 105 | n/a | 534 | n/a | 184 |
|  | E. of Howard Ave |  | n/a | 767 | n/a | 697 | n/a | 11823 | n/a | 237 | n/a | 0 | n/a | 0 |
| N Service Rd | N. of Labelle St |  | 1715 | n/a | 1531 | n/a | 21702 | n/a | 443 | n/a | 3356 | n/a | 0 | n/a |
|  | Labelle St | Grand Marais Rd Ramp | 1449 | n/a | 1436 | n/a | 21787 | n/a | 235 | n/a | 1276 | n/a | 0 | n/a |
|  | Grand Marais Rd Ramp | Pulford St | 255 | n/a | 407 | n/a | 5040 | n/a | 8 | n/a | 376 | n/a | 0 | n/a |
|  | Pulford St | Todd Ln/Cabana Rd | 233 | n/a | 422 | n/a | 5484 | n/a | 12 | n/a | 0 | n/a | 0 | n/a |
|  | Todd Ln/Cabana Rd | Huron Church Line | 727 | n/a | 678 | n/a | 10740 | n/a | 124 | n/a | 449 | n/a |  | n/a |
|  | Huron Church Line | St Clair College | 292 | n/a | 568 | n/a | 7197 | n/a | 32 | n/a | 0 | n/a | 0 | n/a |
|  | St Clair College | Cousineau Dr | 1030 | n/a | 872 | n/a | 12769 | n/a | 123 | n/a | 1825 | n/a | 0 | n/a |
|  | Cousineau Dr | Howard Ave | 339 | n/a | 574 | n/a | 6898 | n/a | 111 | n/a | 451 | n/a | 0 | n/a |
|  | E. of Howard Ave |  | 632 | n/a | 998 | n/a | 13346 | n/a | 283 | n/a | 0 | n/a | 0 | n/a |
| Ojibway Pwy | EC Row Expressway | GN Booth Dr | 665 | 409 | 570 | 790 | 9926 | 9978 | 137 | 137 | 27 | 14 | 91 | 427 |
|  | GN Booth Dr | Sandwich St | 650 | 420 | 576 | 763 | 9861 | 9845 | 136 | 135 | 27 | 15 | 91 | 438 |
|  | Sandwich St | Prospect Ave | 615 | 390 | 540 | 729 | 9354 | 9613 | 75 | 78 | 50 | 37 | 0 | 0 |
|  | N. of Prospect Ave |  | 608 | 379 | 539 | 715 | 9292 | 9397 | 75 | 77 | 49 | 36 | , | 0 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| CROSSING ROADS |  |  | WB | EB | WB | EB | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB/EB |
| Wyandotte | W of HuronChurch |  |  |  |  |  | 4808 | 4435 | 0 | 0 | 359 | 435 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2813 | 4048 | 21 | 135 | 722 | 937 | 18 | 0 |
| University | W of HuronChurch |  |  |  |  |  | 1254 | 1192 | 0 | 0 | 0 | , | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 1947 | 1986 | 118 | 91 | 70 | 21 | 3 | 20 |
| Riverside | W of HuronChurch |  |  |  |  |  | 3390 | 3487 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 6598 | 5633 | 0 | 0 | 173 | 37 | 0 | 0 |
| AMB Off Ramp | E of HuronChurch |  |  |  |  |  | 0 | 931 | 0 | 43 | 0 | 7710 | 0 | 3781 |
| AMB On Ramp | E of HuronChurch |  |  |  |  |  | 309 | , | 11 | 0 | 5792 | 0 | 174 | 0 |
| Patricia | AMB | Wyandotte |  |  |  |  | 552 | 1458 | 21 | 57 | 3367 | 3412 | 171 | 267 |
| College St | E. of HC Road |  | 300 | 349 | 479 | 384 | 6343 | 5558 | 168 | 124 |  | 535 | 0 | 144 |
|  | W. of HC Road |  | 79 | 38 | 141 | 62 | 1670 | 752 | 0 | 0 | 197 | 54 | 0 | 0 |
| Girardot St | E. of HC Road |  | 51 | 86 | 87 | 45 | 1017 | 1029 | 0 | 0 | 116 | 130 | 0 | 0 |
|  | W. of HC Road |  | 81 | 153 | 191 | 126 | 2258 | 2216 | 41 | 25 | 48 | 33 | 0 | 0 |
| Tecumseh Rd | E. of HC Road |  | 312 | 324 | 394 | 462 | 5489 | 6174 | 139 | 148 | 201 | 359 | 0 | 156 |
|  | W. of HC Road |  | 242 | 468 | 524 | 390 | 6420 | 6866 | 0 | 0 | 184 | 127 | 0 | 0 |
| Dorchester St | E. of HC Road |  | 75 | 84 | 131 | 93 | 1520 | 1350 | 0 | 0 | 173 | 183 | 0 | 0 |
|  | W. of HC Road |  | 76 | 46 | 86 | 52 | 1370 | 786 | 26 | 10 | 24 | 11 | 0 | 0 |
| Prince Rd/Totten St | E. of HC Road |  | 139 | 130 | 115 | 205 | 1998 | 2777 | 0 | 0 | 77 | 130 | 0 | 0 |
|  | W. of HC Road |  | 233 | 288 | 315 | 340 | 4701 | 5101 | 0 | 0 | 81 | 76 | 0 | 0 |
| Malden Rd | E. of HC Road |  | 84 | 55 | 85 | 76 | 1172 | 923 | 0 | 0 | 205 | 203 | 0 | 0 |
|  | W. of HC Road |  | 429 | 534 | 464 | 470 | 6798 | 7406 | 386 | 398 | 553 | 38 | 154 | 576 |
| Industrial Rd | E. of HC Road |  | 248 | 145 | 197 | 251 | 3425 | 3181 | 49 | 57 | 139 | 172 | 6 | 16 |
|  | W. of HC Road |  | 290 | 93 | 167 | 275 | 3914 | 2791 | 158 | 192 | 0 | , | 0 | 183 |
| EC Row N. Ramp Terminal | E. of HC Road (E-N/S Off | Ramp \& S-W On Ramp) | 924 | 110 | 872 | 123 | 13014 | 1881 | 270 | 0 | 1050 | 162 | 0 | 0 |
|  | W. of HC Road (N-W On | Ramp) | 36 | n/a | 30 | n/a | 420 | n/a | 14 | n/a | 51 | n/a | 122 | n/a |
| EC Row S. Ramp Terminal | E. of HC Road (S-E On Ra | amp) | n/a | 450 | n/a | 447 | n/a | 7341 | n/a | 66 | n/a | 0 | n/a | 0 |
|  | W. of HC Road (N-E On R | amp \& W-N/S Off Ramp) | 364 | 137 | 630 | 208 | 7642 | 2447 | 263 | 81 | 451 | 376 | 280 | 0 |
| Labelle St/Bethlehem Ave | E. of N. Service Rd |  | 248 | 132 | 125 | 110 | 2670 | 1934 | 0 | 0 | 234 | 182 | 0 | 0 |
|  | between N. and S. Service |  | 70 | 220 | 90 | 170 | 1403 | 3105 | 0 | 0 | 0 | 97 | 0 | 0 |
|  | W. of S. Service Rd |  | 100 | 250 | 105 | 150 | 1803 | 3251 | 0 | 0 | 2 | , | 0 | 0 |
| Grand Marais Rd/Lambton Rd | E. of N. Service Rd |  | 295 | 200 | 225 | 175 | 3919 | 3026 | 0 | 0 | 264 | 244 | 0 | 0 |
|  | between N. and S. Service |  | 99 | 200 | 180 | 160 | 2426 | 2777 | 3 | 24 | 0 | 140 | 0 | 0 |
|  | W. of S. Service Rd |  | 45 | 170 | 155 | 75 | 1647 | 1912 | 29 | 17 | 38 | 32 | 0 | 0 |
| Pulford St | E. of N. Service Rd |  | 143 | 154 | 50 | 65 | 1306 | 1733 | 0 | 0 | 159 | 216 | 0 | 0 |
|  | between N. and S. Service |  | 115 | 120 | 40 | 40 | 1393 | 1073 | 0 | 0 | 0 | 161 | 0 | 0 |
|  | W. of S. Service Rd |  | 110 | 80 | 20 | 30 | 1143 | 855 | 24 | 7 | 11 | 14 | 0 | 0 |
| Todd Ln/Cabana Rd | E. of N. Service Rd |  | 552 | 351 | 541 | 484 | 8177 | 6567 | 0 | 0 | 544 | 610 | 0 | 0 |
|  | between N. and S. Service |  | 463 | 427 | 604 | 535 | 9351 | 7180 | 0 | 0 | 0 | 894 | 0 | 0 |
|  | W. of S. Service Rd |  | 466 | 411 | 670 | 775 | 9920 | 9948 | 0 | 0 | 13 | 10 | 0 | 0 |
| Huron Church Line | between N. and S. Service |  | 50 | 475 | 90 | 200 | 1178 | 4863 | 41 | 77 | 0 | 303 | 0 | 0 |
|  | W. of S. Service Rd |  | 250 | 525 | 620 | 340 | 7017 | 6398 | 93 | 98 | 341 | 345 | 0 | 0 |
| St Clair College | E. of N. Service Rd |  | 132 | 798 | 226 | 159 | 2874 | 8398 | 0 | 0 | 94 | 257 | 0 | 0 |
|  | between N. and S. Service |  | 80 | 257 | 180 | 72 | 2255 | 2418 | 0 | 0 | 0 | 169 | 0 | 0 |
| Cousineau Dr | E. of N. Service Rd |  | 283 | 217 | 362 | 289 | 4440 | 3623 | 0 | 0 | 956 | 793 | 0 | 0 |
|  | between N. and S. Service |  | 343 | 408 | 558 | 404 | 6207 | 6705 | 0 | 0 | 1400 | 0 | 0 | 0 |
|  | W. of S. Service Rd |  | 334 | 460 | 797 | 313 | 9797 | 6318 | 0 | 0 | 0 | , | 0 | 0 |
| Howard Ave | E. of N. Service Rd |  | 390 | 418 | 490 | 599 | 7585 | 8325 | 133 | 155 | 0 | 3 | 0 | 0 |
|  | betweem N . and S. Service |  | 651 | 384 | 755 | 437 | 11382 | 7054 | 211 | 160 | 44 | 2 | 0 | 0 |
|  | W. of S. Service Rd |  | 591 | 743 | 778 | 789 | 11110 | 13141 | 250 | 308 | 21 | 29 | 0 | 0 |
| EC Row Expressway | E. of Ojibway Pwy |  | 720 | 412 | 880 | 520 | 12916 | 7790 | 311 | 383 | 18 | , | 60 | 0 |
|  | W. of Ojibway Pwy |  | 919 | 737 | 1430 | 620 | 16318 | 10907 | 214 | 248 | 3903 | 0 | 103 | 0 |
|  | E. of Huron Church Rd |  | 2138 | 2124 | 2209 | 2834 | 31755 | 37865 | 679 | 775 | 2215 | 3256 | 393 | 1170 |
|  | At Malden Rd |  | 1360 | 1447 | 1490 | 1965 | 20864 | 24508 | 447 | 539 | 1305 | 3374 | 545 | 1040 |
|  | W. of Matchette |  | 940 | 412 | 980 | 520 | 15444 | 7790 | 364 | 383 | 20 | 0 | 66 | 0 |
|  <br> GN Booth Dr <br> Sandwich St <br> Prospect Ave | W. of Ojibway Pwy |  | 27 | 10 | 13 | 44 | 346 | 448 | 7 | . | 4 | 5 | 0 | 0 |
|  | W. of Ojibway Pwy |  | 74 | 69 | 101 | 97 | 1361 | 1261 | 148 | 91 | 24 | 35 | 0 | 0 |
|  | W. of Ojibway Pwy |  | 29 | 33 | 9 | 21 | 331 | 426 | 7 | 4 | 4 | 7 | 0 | 0 |



| LOcation | SECTION |  | 2025 |  |  |  | 24 Hour AADT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Local Cars |  | Local Trucks |  | International Cars |  | International Trucks |  |
|  |  |  | AM PEAK HOUR | PM PEAK HOUR |  |  |  |  |  |
|  | FROM | T0 |  |  | NB | SB | NB | SB | NB/WB | SB/EB | NB/WB | SB / EB | NB/WB | SB / EB | NB/WB | SB/EB |
| HC Road | Riverside | University |  |  |  |  |  |  | 6800 | 5480 | 184 | 89 | 3 | 1 | 0 | 41 |
|  | University | Wyandotte |  |  |  |  | 3052 | 3697 | 91 | 121 | 67 | 233 | 23 | 41 |
|  | Wyandotte | AMB Off Ramp |  |  |  |  | 2234 | 3061 | 0 | 0 | 46 | 163 | 0 | 0 |
|  | AMB Off Ramp | College |  |  |  |  | 8494 | 6347 | 234 | 100 | 6673 | 1 | 2947 | 0 |
|  | College St | Girardot St | 1809 | 849 | 1561 | 1778 | 18669 | 16795 | 572 | 497 | 6714 | 4695 | 263 | 2978 |
|  | Girardot St | Tecumseh Rd | 1722 | 825 | 1407 | 1627 | 18255 | 17710 | 683 | 597 | 5954 | 4048 | 237 | 2803 |
|  | Tecumseh Rd | Dorchester St | 1758 | 1093 | 1767 | 1786 | 21600 | 21412 | 834 | 741 | 5636 | 3507 | 240 | 2539 |
|  | Dorchester St | Prince Rd/Totten St | 1843 | 1151 | 1716 | 1845 | 22280 | 23208 | 743 | 678 | 5128 | 3231 | 207 | 2353 |
|  | Prince Rd/Totten St | Malden Rd | 2011 | 1397 | 1932 | 2050 | 24902 | 27007 | 831 | 798 | 5032 | 3187 | 233 | 2165 |
|  | Malden Rd | Industrial Rd | 1664 | 1154 | 1479 | 1716 | 19169 | 22115 | 608 | 631 | 4976 | 3389 | 8 | 2120 |
|  | Industrial Rd | EC Row N. Ramp Terminal | 1844 | 1183 | 1677 | 1885 | 22450 | 24431 | 704 | 687 | 4868 | 3056 | 0 | 2066 |
|  | EC Row N. Ramp Terminal | EC Row S. Ramp Terminal | 1354 | 1579 | 1196 | 2235 | 15806 | 32888 | 438 | 713 | 4242 | 2742 | 0 | 1746 |
|  | S. of EC Row S. Ramp Terminal |  | 1931 | 1217 | 1720 | 1700 | 22807 | 24956 | 615 | 428 | 4386 | 2287 | 0 | 1444 |
| S Service Rd | N. of Bethlehem Ave |  | n/a | 1217 | n/a | 1665 | n/a | 21124 | n/a | 380 | n/a | 2245 | n/a | 1374 |
|  | Bethlehem Ave | Lambton St | n/a | 323 | n/a | 319 | n/a | 5139 | n/a | 137 | n/a | 333 | n/a | 0 |
|  | Lambton St | Pulford St | n/a | 385 | n/a | 254 | n/a | 5418 | n/a | 18 | n/a | 206 | n/a | 0 |
|  | Pulford St | Todd Ln/Cabana Rd | n/a | 818 | n/a | 715 | n/a | 13482 | n/a | 64 | n/a | 0 | n/a | 0 |
|  | Todd Ln/Cabana Rd | Huron Church Line | n/a | 707 | n/a | 901 | n/a | 13405 | n/a | 121 | n/a | 468 | n/a | 0 |
|  | Huron Church Line | St Clair College | n/a | 944 | n/a | 1069 | n/a | 17578 | n/a | 115 | n/a | 0 | n/a | 0 |
|  | St Clair College | Cousineau Dr | n/a | 390 | n/a | 775 | n/a | 8111 | n/a | 108 | n/a | 1463 | n/a | 221 |
|  | Cousineau Dr | Howard Ave | n/a | 352 | n/a | 369 | n/a | 5422 | n/a | 99 | n/a | 568 | n/a | 233 |
|  | E. of Howard Ave |  | n/a | 857 | n/a | 727 | n/a | 12772 | n/a | 252 | n/a | 0 | n/a | 0 |
| N Service Rd | N. of Labelle St |  | 1901 | n/a | 1720 | n/a | 23981 | n/a | 514 | n/a | 3821 | n/a | 0 | n/a |
|  | Labelle St | Grand Marais Rd Ramp | 1616 | n/a | 1616 | n/a | 24437 | n/a | 282 | n/a | 1408 | n/a | , | n/a |
|  | Grand Marais Rd Ramp | Pulford St | 280 | n/a | 424 | n/a | 5372 | n/a | 8 | n/a | 374 | n/a | 0 | n/a |
|  | Pulford St | Todd Ln/Cabana Rd | 255 | n/a | 472 | n/a | 6082 | n/a | 12 | n/a | 11 | n/a | 0 | n/a |
|  | Todd Ln/Cabana Rd | Huron Church Line | 780 | n/a | 720 | n/a | 11396 | n/a | 139 | n/a | 515 | n/a | 0 | n/a |
|  | Huron Church Line | St Clair College | 321 | n/a | 605 | n/a | 7722 | n/a | 56 | n/a | 0 | n/a |  | n/a |
|  | St Clair College | Cousineau Dr | 1058 | n/a | 889 | n/a | 12982 | n/a | 136 | n/a | 1934 | n/a | 0 | n/a |
|  | Cousineau Dr | Howard Ave | 338 | n/a | 594 | n/a | 6978 | n/a | 112 | n/a | 513 | n/a | 0 | n/a |
|  | E. of Howard Ave |  | 646 | n/a | 1112 | n/a | 14432 | n/a | 300 | n/a | 0 | n/a | 0 | n/a |
| Ojibway Pwy | EC Row Expressway | GN Booth Dr | 700 | 440 | 620 | 820 | 10615 | 10438 | 140 | 134 | 26 | 21 | 114 | 534 |
|  | GN Booth Dr | Sandwich St | 685 | 443 | 626 | 793 | 10549 | 10237 | 139 | 132 | 27 | 21 | 115 | 538 |
|  | Sandwich St | Prospect Ave | 646 | 405 | 582 | 753 | 9965 | 9953 | 74 | 74 | 48 | 42 | 0 | 0 |
|  | N. of Prospect Ave |  | 639 | 394 | 581 | 740 | 9903 | 9744 | 74 | 73 | 48 | 41 | 0 | 0 |
| CROSSING ROADS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  | WB | EB | WB | EB | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB/EB |
| Wyandotte | W of HuronChurch |  |  |  |  |  | 4729 | 4420 | 0 | 0 | 370 | 438 | , | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2772 | 4133 | 18 | 142 | 750 | 926 | 22 | 0 |
| University | W of HuronChurch |  |  |  |  |  | 1365 | 1272 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2079 | 2079 | 121 | 91 | 70 | 21 | 41 | 23 |
| Riverside | W of HuronChurch |  |  |  |  |  | 3552 | 3655 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 6817 | 5737 | 0 | 0 | 164 | 46 | , | 0 |
| AMB Off Ramp | E of HuronChurch |  |  |  |  |  | , | 931 | 0 | 43 | 0 | 7710 | 0 | 3781 |
| AMB On Ramp | E of HuronChurch |  |  |  |  |  | 246 | \#REF! | 6 | \#REF! | 6082 | \#REF! | 223 | \#REF! |
| Patricia | AMB Wyandotte |  |  |  |  |  | 435 | 1328 | 13 | 54 | 3485 | 3362 | 216 | 305 |
| College St |  |  | 307 | 353 | 483 | 399 | 6437 | 5640 | 163 | 127 | 3 | 568 | 0 | 191 |
|  | W. of HC Road |  | 84 | 52 | 165 | 66 | 1677 | 904 | 0 | 0 | 391 | 51 | 0 | 0 |
| Girardot St | E. of HC Road |  | 54 | 84 | 87 | 46 | 1032 | 1014 | 0 | 0 | 122 | 135 | 0 | 0 |
|  | W. of HC Road |  | 82 | 149 | 184 | 121 | 2208 | 2148 | 42 | 25 | 47 | 30 | 0 | 0 |
| Tecumseh Rd | E. of HC Road |  | 317 | 319 | 385 | 468 | 5448 | 6099 | 137 | 145 | 201 | 389 | 0 | 212 |
|  | W. of HC Road |  | 245 | 487 | 508 | 387 | 6271 | 6983 | 0 | 0 | 217 | 132 | 0 | 0 |
| Dorchester St | E. of HC Road |  | 76 | 85 | 134 | 95 | 1544 | 1361 | 0 | 0 | 182 | 196 | 0 | 0 |
|  | W. of HC Road |  | 76 | 46 | 86 | 52 | 1369 | 787 | 26 | 10 | 24 | 10 | 0 | 0 |
| Prince Rd/Totten St | E. of HC Road |  | 148 | 126 | 123 | 143 | 2133 | 2247 | 0 | 0 | 80 | 106 | 0 | 0 |
|  | W. of HC Road |  | 237 | 314 | 398 | 359 | 5414 | 5466 | 0 | 0 | 101 | 79 | 0 | 0 |
| Malden Rd | E. of HC Road |  | 102 | 63 | 99 | 96 | 1355 | 1081 | 0 | 0 | 276 | 277 | 0 | 0 |
|  | W. of HC Road |  | 442 | 557 | 488 | 496 | 7049 | 7669 | 401 | 393 | 570 | 46 | 203 | 766 |
| Industrial Rd | E. of HC Road |  | 270 | 152 | 213 | 272 | 3722 | 3440 | 48 | 57 | 148 | 156 | 8 | 21 |
|  | W. of HC Road |  | 296 | 104 | 183 | 298 | 4094 | 3044 | 168 | 204 | 0 | 0 | 0 | 211 |
| EC Row N. Ramp Terminal | E. of HC Road (E-N/S Off Ramp \& S-W On Ramp) |  | 1029 | 111 | 987 | 125 | 14906 | 1894 | 311 | 0 | 1028 | 174 | 0 | 0 |
|  | W. of HC Road (N-W On Ramp) |  | 32 | n/a | 31 | n/a | 380 | n/a | 11 | n/a | 46 | n/a | 146 | n/a |
| EC Row S. Ramp Terminal | E. of HC Road (S-E On Ramp) |  | n/a | 668 | n/a | 612 | n/a | 10444 | n/a | 103 | n/a | 0 | n/a | 0 |
|  | W. of HC Road (N-E On Ramp \& W-N/S Off Ramp) |  | 425 | 154 | 661 | 214 | 8237 | 2636 | 296 | 85 | 550 | 362 | 383 | 0 |
| Labelle St/Bethlehem Ave | E. of N. Service Rd |  | 262 | 142 | 137 | 126 | 2867 | 2132 | 0 | 0 | 246 | 207 | 0 | 0 |
|  | between N . and S. Service Rd |  | 44 | 242 | 98 | 192 | 1232 | 3459 | 0 | 0 | 0 | 106 | 0 | 0 |
|  | W. of S. Service Rd |  | 77 | 375 | 115 | 165 | 1676 | 4354 | 0 | 0 | 2 | 6 | 0 | 0 |
| Grand Marais Rd/Lambton Rd | E. of N. Service Rd |  | 324 | 220 | 247 | 192 | 4348 | 3339 | 0 | , | 245 | 256 | 0 | 0 |
|  | between N. and S. Service Rd |  | 108 | 220 | 198 | 175 | 2661 | 3047 | , | 29 | 0 | 153 | 0 | 0 |
|  | W. of S. Service Rd |  | 49 | 192 | 170 | 82 | 1801 | 2141 | 34 | 20 | 42 | 33 | 0 | 0 |
| Pulford St |  |  | 156 | 169 | 55 | 71 | 1422 | 1890 | 0 | 0 | 178 | 245 | 0 | 0 |
|  | $\begin{array}{\|l\|} \hline \text { E. of N. Service Rd } \\ \hline \text { between N. and S. Service Rd } \\ \hline \end{array}$ |  | 126 | 132 | 44 | 44 | 1528 | 1170 | 0 | 0 | 0 | 185 | 0 | 0 |
|  | $\begin{array}{\|l\|} \hline \text { between N. and S. Service Rd } \\ \hline \text { W. of S. Service Rd } \\ \hline \end{array}$ |  | 121 | 88 | 22 | 32 | 1258 | 934 | 26 | 8 | 11 | 14 | 0 | 0 |
| Todd Ln/Cabana Rd | E. of N. Service Rd |  | 566 | 355 | 604 | 518 | 8855 | 6797 | 0 | 0 | 535 | 685 | 0 | 0 |
|  | between N. and S. Service Rd |  | 465 | 468 | 665 | 695 | 9884 | 8599 | 0 | 0 | 0 | 1229 | 0 | 0 |
|  | W. of S. Service Rd |  | 569 | 457 | 788 | 887 | 11859 | 11282 | 0 | 0 | 15 | 11 | 0 | 0 |
| Huron Church Line | between N. and S. Service Rd |  | 55 | 522 | 105 | 220 | 1352 | 5299 | 39 | 89 | 0 | 357 | 0 | 0 |
|  |  |  | 275 | 577 | 691 | 368 | 7790 | 6936 | 103 | 114 | 379 | 400 | 0 | 0 |
|  |  |  | 137 | 821 | 248 | 165 | 3101 | 8647 | 0 | 0 | 98 | 268 | 0 | 0 |
|  | between N. and S. Service Rd |  | 88 | 282 | 198 | 73 | 2480 | 2599 | 0 | 0 | 0 | 191 | 0 | 0 |
| St Clair College | E. of N. Service Rd |  | 281 | 228 | 348 | 259 | 4303 | 3409 | 0 | 0 | 951 | 860 | 0 | 0 |
| Cousineau Dr | between N. and S. Service Rd |  | 298 | 402 | ${ }^{651}$ | 453 | 6334 | 7082 | 0 | 11 | 1590 | 0 | 0 | 0 |
|  | W. of S. Service Rd |  | 300 | 472 | 939 | 383 | 10685 | 7010 | 0 | 11 | 0 | 0 | 0 | 0 |
| Howard Ave |  |  | 449 | 447 | 528 | 723 | 8436 | 9599 | 144 | 187 | 0 | 4 | 0 | 0 |
|  | E. of N . Service Rd <br> betweem N. and S. Service Rd |  | 726 | 421 | 812 | 489 | 12433 | 7815 | 221 | 177 | 57 | 1 | 0 | 0 |
|  | betweem N. and S. Service Rd W. of S. Service Rd |  | 666 | 866 | 843 | 878 | 12241 | 14977 | 262 | 342 | 27 | 37 | 0 | 0 |
| EC Row Expressway | E. of Ojibway Pwy |  | 1030 | 450 | 1101 | 580 | 16912 | 8373 | 407 | 400 | 227 | 160 | 152 | 69 |
|  | W. of Ojibway Pwy |  | 1255 | 795 | 1676 | 700 | 19953 | 12035 | 255 | 270 | 5399 | 0 | 134 | 0 |
|  | E. of Huron Church Rd |  | 2722 | 2564 | 2647 | 3299 | 38796 | 44381 | 827 | 929 | 2908 | 4048 | 571 | 1646 |
|  | At Malden Rd |  | 1836 | 1655 | 1816 | 2240 | 25905 | 27562 | 552 | 614 | 2055 | 4070 | 785 | 1404 |
|  | W. of Matchette |  | 1300 | 450 | 1218 | 580 | 19940 | 8373 | 472 | 400 | 252 | 160 | 168 | 69 |
| GN Booth Dr | W. of Ojibway Pwy |  | 27 | 10 | 13 | 44 | 346 | 448 | , | 8 | 4 | 5 | 0 | 0 |
| Sandwich St | W. of Ojibway Pwy |  | 79 | 79 | 107 | 103 | 1455 | 1387 | 151 | 97 | 24 | 34 | 0 | 0 |
| Prospect Ave | W. of Ojibway Pwy |  | 29 | 33 | 9 | 21 | 331 | 426 | 7 | 5 | 4 | 6 | 0 | 0 |



Table A-7 $\quad$ 24-Hour Annual Average Daily Traffic (AADT) for Alternative 1B - Year 2035

| LOCATION | SECTION |  | 2035 |  |  |  | 24 Hour AADT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Local Cars |  | Local Trucks |  | International Cars |  | International Trucks |  |
|  |  |  | AM PEAK HOUR | PM PEAK HOUR |  |  |  |  |  |
|  | FROM | TO |  |  | NB | SB | NB | SB | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB/EB |
| HC Road | Riverside | University |  |  |  |  |  |  | 6718 | 5664 | 203 | 94 | 3 | 1 | 0 | 81 |
|  | University | Wyandotte |  |  |  |  | 2886 | 3812 | 92 | 124 | 68 | 237 | 62 | 81 |
|  | Wyandotte | AMB Off Ramp |  |  |  |  | 2085 | 3201 | 0 | 0 | 46 | 169 | 0 | 0 |
|  | AMB Off Ramp | College |  |  |  |  | 7999 | 6549 | 244 | 106 | 7545 | 1 | 3755 | 0 |
|  | College St | Girardot St | 1846 | 887 | 1546 | 1845 | 18469 | 16494 | 574 | 523 | 6880 | 5152 | 303 | 3764 |
|  | Girardot St | Tecumseh Rd | 1743 | 853 | 1449 | 1695 | 18403 | 17853 | 712 | 625 | 6238 | 4516 | 284 | 3530 |
|  | Tecumseh Rd | Dorchester St | 1782 | 1143 | 1806 | 1846 | 21703 | 21695 | 867 | 781 | 5937 | 3720 | 289 | 3103 |
|  | Dorchester St | Prince Rd/Totten St | 1904 | 1184 | 1737 | 1903 | 22442 | 23399 | 768 | 705 | 5418 | 3417 | 247 | 2837 |
|  | Prince Rd/Totten St | Malden Rd | 2068 | 1458 | 1974 | 2175 | 25203 | 27845 | 865 | 847 | 5387 | 3091 | 281 | 2646 |
|  | Malden Rd | Industrial Rd | 1728 | 1194 | 1509 | 1841 | 19460 | 23370 | 645 | 683 | 5267 | 3219 | 10 | 2695 |
|  | Industrial Rd | EC Row N. Ramp Terminal | 1915 | 1277 | 1704 | 2060 | 22816 | 26119 | 734 | 752 | 5107 | 3366 | 0 | 2631 |
|  | EC Row N . Ramp Terminal | EC Row S. Ramp Terminal | 1450 | 1725 | 1225 | 2448 | 16270 | 35653 | 459 | 783 | 4568 | 3017 | 0 | 2217 |
|  | S. of EC Row S. Ramp Terminal |  | 2046 | 1344 | 1886 | 1883 | 24465 | 27343 | 684 | 474 | 4856 | 2621 | 0 | 1717 |
| S Service Rd | N. of Bethlehem Ave |  | n/a | 1344 | n/a | 1837 | n/a | 23078 | n/a | 420 | n/a | 2547 | n/a | 1629 |
|  | Bethlehem Ave | Lambton St | $\mathrm{n} / \mathrm{a}$ | 352 | n/a | 355 | n/a | 5636 | $\mathrm{n} / \mathrm{a}$ | 161 | n/a | 377 | n/a | 0 |
|  | Lambton St | Pulford St | n/a | 420 | n/a | 311 | n/a | 6187 | n/a | 23 | n/a | 230 | n/a | 0 |
|  | Pulford St | Todd Ln/Cabana Rd | n/a | 908 | n/a | 875 | n/a | 15648 | n/a | 75 | n/a | 0 | n/a | 0 |
|  | Todd Ln/Cabana Rd | Huron Church Line | n/a | 771 | n/a | 1048 | n/a | 15091 | n/a | 134 | n/a | 575 | n/a | 0 |
|  | Huron Church Line | St Clair College | n/a | 1006 | n/a | 1132 | n/a | 18666 | $\mathrm{n} / \mathrm{a}$ | 127 | n/a | 0 | n/a | 0 |
|  | St Clair College | Cousineau Dr | n/a | 418 | n/a | 846 | n/a | 8636 | n/a | 109 | n/a | 1699 | n/a | 292 |
|  | Cousineau Dr | Howard Ave | n/a | 359 | n/a | 394 | n/a | 5540 | n/a | 95 | n/a | 668 | n/a | 293 |
|  | E. of Howard Ave |  | n/a | 922 | n/a | 793 | n/a | 13825 | $\mathrm{n} / \mathrm{a}$ | 281 | n/a | 0 | n/a | 0 |
| N Service Rd | N. of Labelle St |  | 2046 | n/a | 1876 | n/a | 25730 | n/a | 569 | n/a | 4241 | n/a | 0 | n/a |
|  | Labelle St | Grand Marais Rd Ramp | 1744 | n/a | 1762 | n/a | 26402 | n/a | 326 | n/a | 1568 | n/a | 0 | n/a |
|  | Grand Marais Rd Ramp | Pulford St | 306 | n/a | 442 | n/a | 5700 | n/a | 8 | n/a | 391 | n/a | 0 | n/a |
|  | Pulford St | Todd Ln/Cabana Rd | 278 | n/a | 529 | n/a | 6766 | n/a | 13 | n/a | 0 | n/a | 0 | n/a |
|  | Todd Ln/Cabana Rd | Huron Church Line | 844 | n/a | 744 | n/a | 11954 | n/a | 151 | $\mathrm{n} / \mathrm{a}$ | 597 | n/a | 0 | n/a |
|  | Huron Church Line | St Clair College | 350 | n/a | 628 | n/a | 8105 | n/a | 100 | n/a | 0 | n/a | 0 | n/a |
|  | St Clair College | Cousineau Dr | 1085 | n/a | 941 | n/a | 13535 | n/a | 131 | $\mathrm{n} / \mathrm{a}$ | 2005 | n/a | 0 | n/a |
|  | Cousineau Dr | Howard Ave | 368 | n/a | 614 | n/a | 7336 | n/a | 102 | $\mathrm{n} / \mathrm{a}$ | 545 | n/a | 0 | n/a |
|  | E. of Howard Ave |  | 705 | n/a | 1144 | n/a | 15156 | n/a | 316 | n/a | 0 | n/a | 0 | n/a |
| Ojibway Pwy | EC Row Expressway | GN Booth Dr | 735 | 470 | 680 | 860 | 11383 | 10973 | 146 | 131 | 26 | 19 | 142 | 654 |
|  | GN Booth Dr | Sandwich St | 720 | 473 | 686 | 833 | 11317 | 10772 | 146 | 129 | 26 | 19 | 143 | 658 |
|  | Sandwich St | Prospect Ave | 679 | 425 | 633 | 793 | 10661 | 10469 | 76 | 73 | 52 | 47 | 0 | 0 |
|  | N. of Prospect Ave |  | 672 | 415 | 632 | 780 | 10599 | 10270 | 75 | 71 | 51 | 46 | 0 | 0 |
| CROSSING ROADS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  | WB | EB | WB | EB | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB/EB |
| Wyandotte | W of HuronChurch E of HuronChurch |  |  |  |  |  | 4627 | 4439 | 0 | 0 | 381 | 446 | 0 | 0 |
|  |  |  |  |  |  |  | 2803 | 4299 | 17 | 157 | 770 | 942 | 58 | 0 |
| University | W of HuronChurch |  |  |  |  |  | 1511 | 1306 | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  |  |  |  |  |  | 2207 | 2097 | 124 | 92 | 68 | 22 | 81 | 62 |
| Riverside | W of HuronChurch E of HuronChurch |  |  |  |  |  | 3642 | 3993 | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  |  |  |  |  |  | 7055 | 5911 | 0 | 0 | 170 | 46 | 0 |  |
| AMB Off Ramp | E of HuronChurch E of HuronChurch |  |  |  |  |  | 0 | 931 | 0 | 43 | 0 | 7710 | 0 | 3781 |
| AMB On Ramp | E of HuronChurch |  |  |  |  |  | 222 | 0 | 6 | 0 | 6416 | 0 | 273 | 0 |
| Patricia | AMB | Wyandotte |  |  |  |  | 389 | 969 | 12 | 42 | 3571 | 3469 | 234 | 394 |
| College St |  |  | 319 | 351 | 490 | 406 | 6583 | 5598 | 172 | 130 | 4 | 579 | 0 | 273 |
|  | E. of HC Road W. of HC Road |  | 90 | 52 | 187 | 79 | 1730 | 1027 | 0 | 0 | 542 | 48 | 0 | 0 |
| Girardot St | E. of HC Road |  | 59 | 83 | 84 | 48 | 1037 | 1025 | 0 | 0 | 125 | 130 | 0 | 0 |
|  | W. of HC Road |  | 85 | 146 | 180 | 120 | 2202 | 2109 | 42 | 26 | 47 | 33 | 0 | 0 |
| Tecumseh Rd | E. of HC Road |  | 332 | 329 | 420 | 509 | 5868 | 6315 | 140 | 146 | 202 | 468 | 0 | 366 |
|  | W. of HC Road |  | 252 | 505 | 527 | 395 | 6321 | 7251 | 0 | 0 | 357 | 104 | 0 | 0 |
| Dorchester St | E. of HC Road |  | 78 | 86 | 135 | 96 | 1561 | 1382 | 0 | 0 | 187 | 191 | 0 | 0 |
|  | W. of HC Road |  | 76 | 46 | 86 | 52 | 1368 | 785 | 26 | 11 | 24 | 11 | 0 | 0 |
| Prince Rd/Totten St | E. of HC Road |  | 154 | 126 | 129 | 207 | 2228 | 2764 | 0 | 0 | 83 | 125 | 0 | 0 |
|  | W. of HC Road |  | 241 | 329 | 338 | 365 | 4985 | 5626 | 0 | 0 | 68 | 85 | 0 | 0 |
| Malden Rd | E. of HC Road |  | 117 | 71 | 113 | 111 | 1545 | 990 | 0 | 0 | 313 | 519 | 0 | 0 |
|  | W. of HC Road |  | 469 | 573 | 506 | 527 | 7378 | 7922 | 405 | 408 | 599 | 52 | 251 | 932 |
| Industrial Rd | E. of HC Road |  | 305 | 161 | 225 | 285 | 3613 | 3596 | 45 | 56 | 697 | 185 | 7 | 27 |
|  | W. of HC Road |  | 307 | 114 | 198 | 303 | 4310 | 3115 | 179 | 210 | 0 | 0 | 0 | 269 |
| EC Row N. Ramp Terminal | E. of HC Road (E-N/S Off Ramp \& S-W On Ramp) |  | 1060 | 113 | 1020 | 140 | 15527 | 1966 | 327 | 6 | 998 | 242 | 0 | 0 |
|  | W. of HC Road (N-W On Ramp) |  | 34 | n/a | 33 | n/a | 389 | n/a | 11 | $\mathrm{n} / \mathrm{a}$ | 48 | n/a | 176 | n/a |
| EC Row S. Ramp Terminal | E. of HC Road (S-E On Ramp) |  | n/a | 700 | n/a | 756 | n/a | 11928 | n/a | 124 | n/a | 0 | n/a | 0 |
|  | W. of HC Road (N-E On Ramp \& W-N/S Off Ramp) |  | 446 | 169 | 698 | 228 | 8650 | 2901 | 325 | 91 | 465 | 322 | 607 | 0 |
| Labelle St/Bethlehem Ave | E. of N. Service Rd |  | 277 | 167 | 150 | 132 | 3077 | 2391 | 0 | 0 | 259 | 226 | 0 | 0 |
|  | between N. and S. Service Rd |  | 72 | 276 | 108 | 204 | 1573 | 3822 | 0 | 0 | 0 | 112 | 0 | 0 |
|  | W. of S. Service Rd |  | 108 | 220 | 126 | 180 | 2053 | 3279 | 0 | 0 | 2 | 4 | 0 | 0 |
| Grand Marais Rd/Lambton |  |  | 354 | 240 | 270 | 210 | 4753 | 3657 | 0 | 0 | 257 | 272 | 0 | 0 |
|  | E. of N. Service Rdbetween N. and S. Service Rd |  | 118 | 240 | 216 | 192 | 2905 | 3323 | 2 | 34 | 0 | 170 | 0 | 0 |
|  | W. of S. Service Rd |  | 54 | 214 | 186 | 90 | 1973 | 2362 | 37 | 25 | 46 | 41 | 0 | 0 |
| Pulford St | E. of N. Service Rd |  | 171 | 184 | 60 | 78 | 1550 | 2076 | 0 | 0 | 199 | 254 | 0 | 0 |
|  | between N. and S. Service Rd |  | 138 | 144 | 48 | 48 | 1672 | 1282 | 0 | 0 | 0 | 201 | 0 | 0 |
|  | W. of S. Service Rd |  | 132 | 96 | 24 | 36 | 1372 | 1023 | 28 | 10 | 13 | 18 | 0 | 0 |
| Todd Ln/Cabana Rd | E. of N . Service Rd |  | 617 | 406 | 696 | 588 | 10025 | 7623 | 0 | 0 | 561 | 872 | 0 | 0 |
|  | between N. and S. Service Rd |  | 504 | 555 | 746 | 772 | 10926 | 9600 | 0 | 0 | 0 | 1623 | 0 | 0 |
|  | W. of S. Service Rd |  | 642 | 548 | 925 | 992 | 13683 | 12906 | 0 | - | 18 | 12 | 0 |  |
| Huron Church Line | between N. and S. Service Rd |  | 60 | 570 | 110 | 240 | 1426 | 5740 | 53 | 96 | 0 | 416 | 0 | 0 |
|  | W. of S. Service Rd |  | 300 | 630 | 756 | 386 | 8500 | 7407 | 120 | 122 | 422 | 455 | 0 | 0 |
| St Clair College | E. of N. Service Rd |  | 142 | 844 | 265 | 170 | 3288 | 8903 | 0 | 0 | 99 | 265 | 0 | 0 |
|  | between N. and S. Service Rd |  | 96 | 306 | 216 | 74 | 2706 | 2790 | 0 | 0 | 0 | 199 | 0 | 0 |
| Cousineau Dr | E. of N. Service Rd |  | 287 | 242 | 372 | 257 | 4478 | 3440 | 0 | 0 | 1038 | 941 | 0 | 0 |
|  | between N. and S. Service Rd |  | 361 | 458 | 740 | 499 | 7335 | 7920 | 0 | 3 | 1872 | 0 | 0 | 0 |
|  | W. of S. Service Rd |  | 364 | 539 | 1051 | 438 | 12218 | 8015 | 0 | 8 | 0 | 0 | 0 | 0 |
| Howard Ave | E. of N. Service Rd |  | 493 | 482 | 559 | 768 | 9089 | 10255 | 157 | 196 | 0 | 4 | 0 | 0 |
|  | betweem N. and S. Service Rd |  | 794 | 451 | 863 | 543 | 13368 | 8531 | 238 | 192 | 72 | 1 | 0 | 0 |
|  | W. of S. Service Rd |  | 733 | 953 | 900 | 980 | 13228 | 16593 | 284 | 375 | 34 | 46 | 0 | 0 |
|  | E. of Ojibway Pwy |  | 1335 | 480 | 1300 | 640 | 20977 | 9262 | 509 | 404 | 173 | 100 | 145 | 25 |
| EC Row Expressway | W. of Ojibway Pwy |  | 1585 | 835 | 1910 | 780 | 23810 | 13024 | 324 | 290 | 6487 | 0 | 163 | 0 |
|  | E. of Huron Church Rd |  | 3239 | 2837 | 2955 | 3666 | 44205 | 48832 | 932 | 1039 | 3586 | 4703 | 738 | 2009 |
|  | At Malden Rd |  | 2326 | 1860 | 2108 | 2450 | 30617 | 30207 | 642 | 678 | 2860 | 4783 | 1049 | 1550 |
|  | W. of Matchette |  | 1676 | 480 | 1433 | 640 | 24688 | 9262 | 591 | 404 | 190 | 100 | 160 | 25 |
| GN Booth Dr | W. of Ojibway Pwy |  | 27 | 10 | 13 | 44 | 345 | 448 | 7 | 8 | 5 | 5 | 0 |  |
| Sandwich St | W. of Ojibway Pwy |  | 82 | 89 | 121 | 107 | 1598 | 1499 | 156 | 102 | 21 | 29 | 0 | 0 |
| Prospect Ave |  |  | 29 | 33 | 9 | 21 | 331 | 425 | 7 | 5 | 4 | 7 | 0 |  |



| LOCATION | SECTION |  | Alternative 2A |  |  |  | 24 Hour AADT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | $2015$ |  |  |  | Local Cars |  | Local Trucks |  | International Cars |  | International Trucks |  |
|  |  |  | AM PEAK HOUR |  | PM PEAK HOUR |  |  |  |  |  |  |  |  |  |
|  | FROM | TO | NB/WB | SB/EB | NB/WB | SB/EB | NB/ WB | SB/EB | NB/WB | SB/EB | NB/WB | SB / EB | NB/WB | SB / EB |
| HC Road | Riverside | University |  |  |  |  | 6763 | 5390 | 173 | 85 | 3 | 1 | 0 | 3 |
|  | University | Wyandotte |  |  |  |  | 3111 | 3651 | 90 | 119 | 65 | 198 | 20 | 3 |
|  | Wyandotte | AMB Off Ramp |  |  |  |  | 2211 | 3045 | 0 | 0 | 44 | 131 | 0 | 0 |
|  | AMB Off Ramp | College |  |  |  |  | 8806 | 6251 | 231 | 96 | 6173 | 1 | 2162 | 0 |
|  | College St | Girardot St | 1797 | 834 | 1389 | 1656 | 17474 | 16649 | 522 | 486 | 6193 | 4360 | 192 | 2209 |
|  | Girardot St | Tecumseh Rd | 1720 | 817 | 1275 | 1536 | 17362 | 17487 | 621 | 567 | 5517 | 3775 | 176 | 2114 |
|  | Tecumseh Rd | Dorchester St | 1789 | 1119 | 1632 | 1717 | 20829 | 21671 | 768 | 720 | 5325 | 3452 | 181 | 1987 |
|  | Dorchester St | Prince Rd/Totten St | 1831 | 1163 | 1593 | 1799 | 21353 | 23522 | 685 | 653 | 4826 | 3173 | 156 | 1799 |
|  | Prince Rd/Totten St | Malden Rd | 1934 | 1384 | 1806 | 1952 | 23703 | 26772 | 763 | 750 | 4676 | 3162 | 177 | 1660 |
|  | Malden Rd | Industrial Rd | 1588 | 1142 | 1466 | 1623 | 18911 | 21704 | 573 | 566 | 4737 | 3352 | 7 | 1601 |
|  | Industrial Rd | EC Row N. Ramp Terminal | 1703 | 1199 | 1551 | 1743 | 20829 | 23841 | 635 | 632 | 4457 | 3024 | 0 | 1593 |
|  | EC Row N. Ramp Terminal | EC Row S. Ramp Terminal | 1190 | 1453 | 1223 | 2080 | 15536 | 30518 | 425 | 631 | 3900 | 2716 | 0 | 1307 |
|  | EC Row S. Ramp Terminal | Highway 401 Offramp | 1612 | 1082 | 1638 | 1555 | 23080 | 22647 | 449 | 380 | 2181 | 2159 | 0 | 1032 |
|  | Highway 401 Offramp | Spring Gdn Rd/Labelle St | 1055 | 412 | 406 | 430 | 10669 | 6674 | 78 | 97 | 654 | 564 | 0 | 0 |
|  | Spring Gdn Rd/Labelle St | Lambton St/Grand Marais Rd Ramp | 878 | 384 | 342 | 465 | 9076 | 6728 | 111 | 134 | 413 | 516 | 0 | 0 |
|  | Lambton St/Grand Marais Rd | Pulford St | 675 | 374 | 354 | 396 | 7863 | 6452 | 136 | 37 | 209 | 240 | 0 | 0 |
|  | Pulford St | Todd Ln/Cabana Rd | 648 | 422 | 359 | 417 | 8033 | 7308 | 152 | 50 | 0 | 37 | 0 | 0 |
|  | Todd Ln/Cabana Rd | Huron Church Line | 768 | 627 | 628 | 894 | 10252 | 12558 | 126 | 153 | 697 | 461 | 0 | 54 |
| Talbot Rd | Huron Church Line | St Clair College | 403 | 503 | 418 | 376 | 5522 | 6832 | 8 | 0 | 884 | 850 | 0 | 0 |
|  | St Clair College | Cousineau Dr | 858 | 303 | 464 | 429 | 8455 | 6404 | 10 | 0 | 1430 | 0 | 0 | 0 |
|  | Cousineau Dr | Howard Ave | 723 | 322 | 485 | 298 | 9730 | 5472 | 21 | 0 | 81 | 0 | 0 | 0 |
|  | Howard Ave | Highway 3 split | 982 | 789 | 862 | 711 | 14885 | 12943 | 291 | 278 | 0 | 25 | 0 | 0 |
| Ojibway Pwy | EC Row Expressway | GN Booth Dr | 665 | 409 | 570 | 790 | 9922 | 10623 | 136 | 139 | 32 | 15 | 90 | 430 |
|  | GN Booth Dr | Sandwich St | 650 | 411 | 576 | 765 | 9857 | 10411 | 136 | 136 | 32 | 15 | 91 | 432 |
|  | Sandwich St | Prospect Ave | 615 | 381 | 541 | 738 | 9360 | 9610 | 75 | 78 | 52 | 35 | 0 | 0 |
|  | N. of Prospect Ave |  | 608 | 370 | 540 | 725 | 9298 | 9402 | 74 | 76 | 52 | 34 | 0 | 0 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| CROSSING ROADS |  |  | WB | EB | WB | EB |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Wyandotte | W of HuronChurch |  |  |  |  |  | 4834 | 4441 | 0 | 0 | 358 | 431 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2850 | 4074 | 21 | 137 | 726 | 867 | 18 | 0 |
| University | W of HuronChurch |  |  |  |  |  | 1267 | 1128 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 1950 | 2014 | 119 | 90 | 67 | 21 | 3 | 20 |
| Riverside | W of HuronChurch |  |  |  |  |  | 3367 | 3642 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 6608 | 5719 | 0 | 0 | 132 | 45 | 0 | 0 |
| AMB Off Ramp | E of HuronChurch |  |  |  |  |  | 0 | 1643 | 0 | 60 | 0 | 6170 | 3 | 2162 |
| AMB On Ramp | E of HuronChurch |  |  |  |  |  | 317 | 0 | 12 | 0 | 5799 | 0 | 174 | 0 |
| Patricia | AMB | Wyandotte |  |  |  |  | 567 | 1642 | 22 | 58 | 3342 | 3241 | 171 | 231 |
| College St |  |  | 312 | 355 | 483 | 387 | 6465 | 5630 | 175 | 132 | 3 | 532 | 0 | 150 |
|  | E. of HC Road <br> W. of HC Road |  | 75 | 35 | 135 | 64 | 1566 | 746 | 0 | 0 | 210 | 59 | 0 | 0 |
| Girardot St |  |  | 51 | 86 | 86 | 45 | 1060 | 1052 | 0 | 0 | 68 | 109 | 0 | 0 |
|  | $\begin{array}{\|l} \hline \text { E. of HC Road } \\ \hline \text { W. of HC Road } \\ \hline \end{array}$ |  | 81 | 153 | 191 | 126 | 2178 | 2149 | 36 | 29 | 114 | 91 | 9 | 4 |
| Tecumseh Rd |  |  | 315 | 316 | 409 | 462 | 5638 | 6115 | 134 | 150 | 211 | 357 | 0 | 137 |
|  | E. of HC Road <br> W. of HC Road |  | 241 | 465 | 505 | 385 | 6296 | 6802 | 0 | 0 | 147 | 125 | 0 | 0 |
| Dorchester St | E. of HC Road |  | 75 | 84 | 132 | 93 | 1605 | 1382 | 0 | 0 | 102 | 155 | 0 | 0 |
|  | W. of HC Road |  | 76 | 46 | 86 | 52 | 1321 | 762 | 23 | 10 | 64 | 34 | 8 | 1 |
| Prince $\mathrm{Rd} /$ /otten St |  |  | 144 | 108 | 115 | 206 | 2040 | 2591 | 0 | 0 | 73 | 121 | 0 | 0 |
|  | $\begin{array}{\|l\|} \hline \text { E. of HC Road } \\ \hline \text { W. of HC Road } \\ \hline \end{array}$ |  | 235 | 299 | 322 | 337 | 4778 | 5165 | 0 | 0 | 82 | 73 | 0 | 0 |
| Malden Rd | E. of HC Road |  | 84 | 55 | 85 | 74 | 1175 | 931 | 0 | 0 | 201 | 183 | 0 | 0 |
|  | W. of HC Road |  | 373 | 403 | 479 | 432 | 6483 | 6223 | 368 | 337 | 484 | 36 | 168 | 431 |
| Industrial Rd | E. of HC Road |  | 244 | 142 | 202 | 257 | 3437 | 3215 | 46 | 55 | 146 | 164 | 6 | 16 |
|  | W. of HC Road |  | 290 | 92 | 148 | 275 | 3764 | 2787 | 147 | 191 | 0 | 0 | 0 | 179 |
| EC Row N. Ramp Termina | E-N/S Ramp |  | 843 | n/a | 870 | n/a | 12487 | n/a | 266 | n/a | 987 | n/a | 0 | n/a |
|  | N-W Ramp |  | n/a | 29 | n/a | 63 | n/a | 731 | n/a | 0 | n/a | 37 | n/a | 0 |
|  | S-W Ramp |  | 40 | n/a | 130 | n/a | 913 | n/a | 28 | n/a | 92 | n/a | 528 | n/a |
| EC Row S. Ramp Termina | W-N/S Ramp |  | n/a | 148 | n/a | 188 | n/a | 2461 | n/a | 47 | n/a | 297 | n/a | 0 |
|  | S/N-E Ramp |  | n/a | 941 | n/a | 1083 | n/a | 15627 | n/a | 296 | n/a | 762 | n/a | 224 |
| Spring Gdn Rd |  |  | 115 | 210 | 150 | 170 | 2320 | 3114 | 0 | 0 | 2 | 4 | 0 | 0 |
| Labelle St | E. of HC Road |  | 293 | 144 | 167 | 66 | 3338 | 1722 | 0 | 0 | 268 | 144 | 0 | 0 |
| Lambton St/Grand Marais | E. of HC Road |  | 420 | 225 | 234 | 285 | 5171 | 4240 | 0 | 0 | 95 | 198 | 0 | 0 |
| Pulford St | E. of HC Road |  | 203 | 128 | 207 | 91 | 3090 | 1739 | 0 | 0 | 211 | 186 | 0 | 0 |
| Cabana Rd | E. of HC Road |  | 550 | 324 | 467 | 444 | 7381 | 6096 | 0 | 0 | 646 | 518 | 0 | 0 |
|  | between HC Road and Hwy 401 NB Ramps |  | 1308 | 1175 | 928 | 1124 | 19803 | 17449 | 39 | 127 | 0 | 1440 | 0 | 66 |
| Todd Lane | between Hwy 401 NB and SB Ramps |  | 773 | 849 | 586 | 1122 | 12020 | 14962 | 24 | 113 | 0 | 1413 | 0 | 48 |
|  | W. of Hwy 401 SB Ramps |  | 675 | 527 | 613 | 664 | 11358 | 9885 | 0 | 0 | 12 | 10 | 0 | 0 |
| Huron Church Line | W. of HC Road |  | 251 | 454 | 614 | 306 | 7091 | 5657 | 97 | 80 | 234 | 294 | 22 | 0 |
| St Clair College | E. of Talbot Road |  | 150 | 805 | 234 | 267 | 2984 | 9148 | 0 | 0 | 160 | 451 | 0 | 0 |
| Cousineau Dr | E. of Talbot Road |  | 268 | 295 | 387 | 300 | 5464 | 4248 | 0 | 0 | 0 | 985 | 0 | 0 |
|  | W. of Talbot Road |  | 246 | 408 | 589 | 365 | 7233 | 6365 | 0 | 0 | 0 | 0 | 0 | 0 |
| Howard Ave | E. of Talbot Rd |  | 403 | 551 | 532 | 718 | 7723 | 10360 | 128 | 195 | 283 | 2 | 0 | 0 |
|  | betweem Talbot Road and Hwy 401 SB On-Ramp |  | 707 | 315 | 1015 | 526 | 13285 | 7183 | 208 | 150 | 951 | 0 | 0 | 0 |
|  | W. of Hwy 401 SB On-Ramp |  | 520 | 790 | 976 | 886 | 12302 | 14210 | 247 | 290 | 11 | 192 | 0 | 0 |
| E.C. Row Expressway | E. of Huron Church Rd |  | 2154 | 2268 | 2167 | 2873 | 31509 | 39193 | 678 | 807 | 2214 | 3455 | 390 | 1241 |
|  | At Malden Rd |  | 1380 | 1475 | 1490 | 1978 | 20975 | 25038 | 452 | 548 | 1325 | 3220 | 545 | 1052 |
|  | W. of Matchette |  | 960 | 412 | 980 | 520 | 15597 | 7773 | 364 | 401 | 23 | 0 | 66 | 0 |

[^2]

Table A-9 24-Hour Annual Average Daily Traffic (AADT) for Alternative 2A - Year 2025

| LOCATION | SECTION |  |  |  |  |  |  |  |  | 24 Hour | AADT |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 2025 |  |  |  | Local Cars |  | Local Trucks |  | International Cars |  | International Trucks |  |
|  |  |  | AM PEAK HOUR |  | PM PEAK HOUR |  |  |  |  |  |  |  |  |  |
|  | FROM | TO | NB / WB | SB/EB | NB/WB | SB/EB | NB / WB | SB / EB | NB / WB | SB/EB | NB / WB | SB/EB | NB / WB | SB/EB |
| HC Road | Riverside | University |  |  |  |  | 6879 | 5510 | 180 | 92 | 3 | 1 | 0 | 41 |
|  | University | Wyandotte |  |  |  |  | 3138 | 3726 | 88 | 124 | 67 | 191 | 56 | 41 |
|  | Wyandotte | AMB Off Ramp |  |  |  |  | 2188 | 3074 | 0 | 0 | 45 | 128 | 0 | 0 |
|  | AMB Off Ramp | College |  |  |  |  | 8649 | 6379 | 229 | 103 | 6543 | 1 | 2797 | 0 |
|  | College St | Girardot St | 1812 | 878 | 1461 | 1778 | 17875 | 17283 | 550 | 519 | 6510 | 4605 | 245 | 2918 |
|  | Girardot St | Tecumseh Rd | 1703 | 849 | 1348 | 1627 | 17721 | 18123 | 662 | 603 | 5770 | 3971 | 226 | 2741 |
|  | Tecumseh Rd | Dorchester St | 1772 | 1123 | 1711 | 1806 | 21215 | 22028 | 818 | 751 | 5543 | 3513 | 230 | 2420 |
|  | Dorchester St | Prince Rd/Totten St | 1853 | 1201 | 1676 | 1895 | 22001 | 24262 | 734 | 690 | 5064 | 3277 | 200 | 2247 |
|  | Prince Rd/Totten St | Malden Rd | 1952 | 1453 | 1905 | 2090 | 24456 | 28088 | 820 | 807 | 4895 | 3216 | 228 | 2022 |
|  | Malden Rd | Industrial Rd | 1607 | 1157 | 1550 | 1778 | 19557 | 22795 | 629 | 615 | 4986 | 3431 | 9 | 1979 |
|  | Industrial Rd | EC Row N. Ramp Terminal | 1744 | 1255 | 1710 | 1866 | 22312 | 25206 | 711 | 686 | 4773 | 3036 | 0 | 1963 |
|  | EC Row N. Ramp Terminal | EC Row S. Ramp Terminal | 1200 | 1653 | 1352 | 2350 | 16550 | 34662 | 480 | 730 | 4187 | 2860 | 0 | 1700 |
|  | EC Row S. Ramp Terminal | Highway 401 Offramp | 1777 | 1250 | 1848 | 1802 | 25613 | 26275 | 527 | 460 | 2559 | 2338 | 0 | 1316 |
|  | Highway 401 Offramp | Spring Gdn Rd/Labelle St | 1110 | 562 | 434 | 450 | 11292 | 8073 | 89 | 137 | 688 | 666 | 0 | 0 |
|  | Spring Gdn Rd/Labelle St | Lambton St/Grand Marais Rd Ramp | 927 | 535 | 372 | 465 | 9687 | 7975 | 134 | 190 | 431 | 596 | 0 | 0 |
|  | Lambton St/Grand Marais Rd | Pulford St | 720 | 475 | 354 | 445 | 8242 | 7776 | 152 | 50 | 183 | 246 | 0 | 0 |
|  | Pulford St | Todd Ln/Cabana Rd | 679 | 512 | 403 | 472 | 8646 | 8568 | 165 | 62 | 0 | 56 | 0 | 0 |
|  | Todd Ln/Cabana Rd | Huron Church Line | 839 | 737 | 768 | 954 | 11835 | 14051 | 155 | 158 | 801 | 470 | 0 | 66 |
| Talbot Rd | Huron Church Line | St Clair College | 440 | 501 | 444 | 403 | 5765 | 7109 | 13 | 0 | 1087 | 787 | 0 | 0 |
|  | St Clair College | Cousineau Dr | 881 | 319 | 516 | 484 | 8799 | 7016 | 12 | 0 | 1628 | 0 | 0 | 0 |
|  | Cousineau Dr | Howard Ave | 736 | 358 | 552 | 341 | 10300 | 6162 | 46 | 3 | 138 | 0 | 0 | 0 |
|  | Howard Ave | Highway 3 split | 1028 | 900 | 964 | 786 | 16103 | 14544 | 320 | 315 | 0 | 41 | 0 | 0 |
| Ojibway Pwy | EC Row Expressway | GN Booth Dr | 700 | 440 | 620 | 820 | 10615 | 10841 | 139 | 134 | 27 | 21 | 113 | 531 |
|  | GN Booth Dr | Sandwich St | 685 | 442 | 626 | 795 | 10550 | 10636 | 139 | 132 | 27 | 21 | 114 | 533 |
|  | Sandwich St | Prospect Ave | 646 | 403 | 582 | 760 | 9966 | 9995 | 74 | 74 | 48 | 41 | 0 | 0 |
|  | N. of Prospect Ave |  | 639 | 392 | 581 | 747 | 9904 | 9787 | 73 | 72 | 48 | 40 | 0 | 0 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| CROSSING ROADS |  |  | WB | EB | WB | EB |  |  |  |  |  |  |  |  |
| Wyandotte | W of HuronChurch |  |  |  |  |  | 4742 | 4425 | 0 | 0 | 370 | 436 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2809 | 4146 | 18 | 141 | 749 | 860 | 51 | 0 |
| University | W of HuronChurch |  |  |  |  |  | 1358 | 1254 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2083 | 2192 | 124 | 88 | 63 | 21 | 41 | 56 |
| Riverside | W of HuronChurch |  |  |  |  |  | 3547 | 3769 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 6822 | 5807 | 0 | 0 | 128 | 46 | 0 | 0 |
| AMB Off Ramp | E of HuronChurch |  |  |  |  |  | 0 | 1407 | 0 | 54 | 0 | 6540 | 41 | 2797 |
| AMB On Ramp | E of HuronChurch |  |  |  |  |  | 276 | 0 | 7 | 0 | 6098 | 0 | 223 | 0 |
| Patricia | AMB | Wyandotte |  |  |  |  | 490 | 1407 | 14 | 51 | 3456 | 3338 | 193 | 305 |
| College St | E. of HC Road |  | 321 | 356 | 487 | 404 | 6567 | 5748 | 178 | 133 | 3 | 534 | 0 | 193 |
|  | W. of HC Road |  | 84 | 43 | 163 | 67 | 1466 | 857 | 0 | 0 | 546 | 43 | 0 | 0 |
| Girardot St | E. of HC Road |  | 54 | 84 | 84 | 48 | 1065 | 1057 | 0 | 0 | 66 | 110 | 0 | 0 |
|  | W. of HC Road |  | 82 | 149 | 183 | 122 | 2131 | 2092 | 33 | 27 | 108 | 83 | 11 | 4 |
| Tecumseh Rd | E. of HC Road |  | 318 | 310 | 381 | 468 | 5430 | 5973 | 128 | 147 | 200 | 400 | 0 | 245 |
|  | W. of HC Road |  | 237 | 460 | 494 | 379 | 6273 | 6703 | 0 | 0 | 62 | 130 | 0 | 0 |
| Dorchester St | E. of HC Road |  | 77 | 85 | 134 | 95 | 1638 | 1406 | 0 | 0 | 101 | 156 | 0 | 0 |
|  | W. of HC Road |  | 76 | 46 | 86 | 52 | 1323 | 764 | 22 | 10 | 63 | 31 | 10 | 2 |
| Prince Rd/Totten St | E. of HC Road |  | 154 | 101 | 123 | 210 | 2177 | 2562 | 0 | 0 | 83 | 121 | 0 | 0 |
|  | W. of HC Road |  | 241 | 315 | 335 | 344 | 4937 | 5354 | 0 | 0 | 87 | 73 | 0 | 0 |
| Malden Rd | E. of HC Road |  | 102 | 63 | 105 | 94 | 1422 | 1098 | 0 | 0 | 259 | 250 | 0 | 0 |
|  | W. of HC Road |  | 398 | 413 | 486 | 421 | 6763 | 6083 | 368 | 315 | 461 | 90 | 206 | 562 |
| Industrial Rd | E. of HC Road |  | 267 | 151 | 213 | 296 | 3702 | 3605 | 47 | 56 | 146 | 180 | 8 | 21 |
|  | W. of HC Road |  | 299 | 103 | 185 | 297 | 4136 | 3003 | 171 | 199 | 0 | 0 | 0 | 250 |
| EC Row N. Ramp Termina | E-N/S Ramp |  | 993 | n/a | 1002 | n/a | 14751 | n/a | 313 | n/a | 1035 | n/a | 0 | n/a |
|  | N-W Ramp |  | n/a | 30 | n/a | 76 | n/a | 846 | n/a | 0 | n/a | 43 | n/a | 0 |
|  | S-W Ramp |  | 40 | n/a | 137 | n/a | 882 | n/a | 31 | n/a | 91 | n/a | 638 | $\mathrm{n} / \mathrm{a}$ |
| EC Row S. Ramp Terminal | W-N/S Ramp |  | n/a | 154 | n/a | 194 | n/a | 2478 | n/a | 63 | n/a | 369 | n/a | 0 |
|  | S/N-E Ramp |  | n/a | 1134 | n/a | 1194 | n/a | 17986 | n/a | 362 | n/a | 704 | n/a | 344 |
| Spring Gdn Rd | W. of HC Road |  | 115 | 210 | 150 | 170 | 2320 | 3114 | 0 | 0 | 2 | 4 | 0 | 0 |
| Labelle St | E. of HC Road |  | 330 | 174 | 168 | 56 | 3589 | 1905 | 0 | 0 | 294 | 154 | 0 | 0 |
| Lambton St/Grand Marais | E. of HC Road |  | 420 | 275 | 234 | 285 | 5182 | 4700 | 0 | 0 | 88 | 200 | 0 | 0 |
| Pulford St | E. of HC Road |  | 186 | 122 | 184 | 48 | 2788 | 1379 | 0 | 0 | 186 | 138 | 0 | 0 |
| Cabana Rd | E. of HC Road |  | 611 | 369 | 475 | 520 | 7859 | 7079 | 0 | 0 | 674 | 574 | 0 | 0 |
|  | between HC Road and Hwy 401 | 1 NB Ramps | 1444 | 1271 | 1033 | 1206 | 21920 | 18772 | 56 | 136 | 0 | 1574 | 0 | 82 |
| Todd Lane | between Hwy 401 NB and SB R | Ramps | 801 | 976 | 659 | 1186 | 12884 | 16383 | 33 | 123 | 0 | 1528 | 0 | 63 |
|  | W. of Hwy 401 SB Ramps |  | 689 | 533 | 694 | 682 | 12168 | 10089 | 0 | 0 | 14 | 10 | 0 | 0 |
| Huron Church Line | W. of HC Road |  | 324 | 508 | 671 | 444 | 8180 | 7131 | 100 | 102 | 278 | 378 | 29 | 0 |
| St Clair College | E. of Talbot Road |  | 150 | 785 | 246 | 281 | 3141 | 9151 | 0 | 0 | 112 | 396 | 0 | 0 |
| Cousineau Dr | E. of Talbot Road |  | 250 | 290 | 431 | 339 | 5707 | 4478 | 0 | 0 | 0 | 1034 | 0 | 0 |
|  | W. of Talbot Road |  | 265 | 448 | 679 | 402 | 8163 | 6968 | 4 | 31 | 0 | 0 | 0 | 0 |
| Howard Ave | E. of Talbot Rd |  | 453 | 607 | 552 | 778 | 8327 | 11296 | 138 | 218 | 289 | 2 | 10 | 0 |
|  | betweem Talbot Road and Hwy | y 401 SB On-Ramp | 747 | 350 | 1116 | 570 | 14389 | 7858 | 223 | 168 | 1019 | 1 | 15 | 0 |
|  | W. of Hwy 401 SB On-Ramp |  | 550 | 914 | 1075 | 981 | 13377 | 16078 | 261 | 331 | 15 | 219 | 0 |  |
| E.C. Row Expressway | E. of Huron Church Rd |  | 2769 | 2559 | 2604 | 3237 | 38745 | 43645 | 822 | 917 | 2922 | 4191 | 575 | 1644 |
|  | At Malden Rd |  | 1846 | 1579 | 1815 | 2237 | 25989 | 26982 | 552 | 588 | 2030 | 4004 | 810 | 1315 |
|  | W. of Matchette |  | 1310 | 450 | 1217 | 580 | 20264 | 8624 | 472 | 404 | 23 | 0 | 96 | 0 |

*     - For consistency, Huron Church Rd/Talbot Rd runs North-South and all crossing roads run East-West


| LOCATION | SECTION |  |  |  |  |  | 24 Hour AADT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 2035 |  |  |  | Local Cars |  | Local Trucks |  | International Cars |  | International Trucks |  |
|  |  |  | AM PEAK HOUR |  | PM PEAK HOUR |  |  |  |  |  |  |  |  |  |
|  | FROM | TO | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB / EB | NB/WB | SB/EB | NB/WB | SB / EB | NB/WB | SB/EB |
| HC Road | Riverside | University |  |  |  |  | 6841 | 5716 | 194 | 96 | 3 | 1 | 0 | 80 |
|  | University | Wyandotte |  |  |  |  | 3085 | 3887 | 88 | 127 | 77 | 186 | 62 | 80 |
|  | Wyandotte | AMB Off Ramp |  |  |  |  | 2237 | 3286 | 0 | 0 | 55 | 128 | 0 | 0 |
|  | AMB Off Ramp | College |  |  |  |  | 8264 | 6614 | 233 | 108 | 7649 | 1 | 3551 | 0 |
|  | College St | Girardot St | 1854 | 907 | 1500 | 1835 | 18116 | 16775 | 568 | 526 | 6811 | 5142 | 295 | 3625 |
|  | Girardot St | Tecumseh Rd | 1784 | 879 | 1380 | 1764 | 18031 | 18277 | 690 | 633 | 6165 | 4689 | 271 | 3498 |
|  | Tecumseh Rd | Dorchester St | 1835 | 1203 | 1740 | 1987 | 21405 | 22845 | 848 | 812 | 5878 | 4339 | 275 | 3082 |
|  | Dorchester St | Prince Rd/Totten St | 1914 | 1273 | 1697 | 2083 | 22173 | 25058 | 758 | 751 | 5347 | 4113 | 239 | 2888 |
|  | Prince Rd/Totten St | Malden Rd | 2007 | 1521 | 1942 | 2328 | 24722 | 29055 | 851 | 877 | 5231 | 4139 | 274 | 2590 |
|  | Malden Rd | Industrial Rd | 1652 | 1211 | 1576 | 2009 | 19807 | 23638 | 658 | 673 | 5197 | 4387 | 10 | 2592 |
|  | Industrial Rd | EC Row N. Ramp Terminal | 1777 | 1302 | 1682 | 2116 | 22107 | 26382 | 716 | 751 | 4838 | 3849 | 0 | 2513 |
|  | EC Row N. Ramp Terminal | EC Row S. Ramp Terminal | 1231 | 1752 | 1342 | 2482 | 16462 | 35843 | 481 | 792 | 4297 | 3423 | 0 | 2087 |
|  | EC Row S. Ramp Terminal | Highway 401 Offramp | 1878 | 1351 | 1952 | 1950 | 26863 | 28114 | 566 | 507 | 2834 | 2624 | 0 | 1586 |
|  | Highway 401 Offramp | Spring Gdn Rd/Labelle St | 1123 | 609 | 474 | 472 | 11642 | 8637 | 96 | 148 | 750 | 704 | 0 | 0 |
|  | Spring Gdn Rd/Labelle St | Lambton St/Grand Marais Rd Ramp | 934 | 580 | 401 | 531 | 9868 | 8846 | 142 | 210 | 506 | 664 | 0 | 0 |
|  | Lambton St/Grand Marais Rd | Pulford St | 729 | 510 | 402 | 443 | 8470 | 8099 | 133 | 55 | 344 | 226 | 0 | 0 |
|  | Pulford St | Todd Ln/Cabana Rd | 718 | 554 | 432 | 522 | 9190 | 9374 | 178 | 69 | 0 | 51 | 0 | 0 |
|  | Todd Ln/Cabana Rd | Huron Church Line | 952 | 846 | 868 | 1107 | 13388 | 16198 | 175 | 199 | 913 | 547 | 8 | 81 |
| Talbot Rd | Huron Church Line | St Clair College | 465 | 545 | 476 | 431 | 5966 | 7815 | 14 | 0 | 1266 | 734 | 0 | 0 |
|  | St Clair College | Cousineau Dr | 905 | 302 | 551 | 482 | 9054 | 6843 | 13 | 0 | 1791 | 0 | 0 | 0 |
|  | Cousineau Dr | Howard Ave | 751 | 324 | 588 | 375 | 10685 | 6137 | 68 | 4 | 157 | 0 | 0 | 0 |
|  | Howard Ave | Highway 3 split | 1058 | 981 | 1103 | 856 | 17511 | 15829 | 358 | 349 | 0 | 57 | 0 | 0 |
| Ojibway Pwy | EC Row Expressway | GN Booth Dr | 735 | 470 | 680 | 860 | 11381 | 11654 | 146 | 134 | 27 | 20 | 143 | 652 |
|  | GN Booth Dr | Sandwich St | 720 | 472 | 686 | 835 | 11316 | 11442 | 146 | 132 | 27 | 19 | 144 | 654 |
|  | Sandwich St | Prospect Ave | 679 | 424 | 632 | 800 | 10653 | 10523 | 76 | 72 | 51 | 44 | 0 | 0 |
|  | N. of Prospect Ave |  | 672 | 413 | 631 | 787 | 10590 | 10315 | 75 | 71 | 51 | 43 | 0 | 0 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| CROSSING ROADS |  |  | WB | EB | WB | EB |  |  |  |  |  |  |  |  |
| Wyandotte | W of HuronChurch |  |  |  |  |  | 4556 | 4436 | 0 | 0 | 381 | 446 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2767 | 4279 | 17 | 152 | 785 | 866 | 57 | 0 |
| University | W of HuronChurch |  |  |  |  |  | 1504 | 1357 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2193 | 2195 | 127 | 88 | 57 | 22 | 80 | 62 |
| Riverside | W of HuronChurch |  |  |  |  |  | 3634 | 3914 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 7125 | 5981 | 0 | 0 | 129 | 56 | 0 | 0 |
| AMB Off Ramp | E of HuronChurch |  |  |  |  |  | 0 | 1096 | 0 | 44 | 0 | 7646 | 80 | 3551 |
| AMB On Ramp | E of HuronChurch |  |  |  |  |  | 214 | 0 | 5 | 0 | 6435 | 0 | 273 | 0 |
| Patricia | AMB | Wyandotte |  |  |  |  | 375 | 1095 | 12 | 42 | 3544 | 3434 | 263 | 395 |
| College St |  |  | 333 | 359 | 482 | 423 | 6613 | 5752 | 182 | 139 | 3 | 636 | 0 | 253 |
|  | E. of HC Road |  | 90 | 50 | 184 | 71 | 1670 | 946 | 0 | 0 | 570 | 44 | 0 | 0 |
| Girardot St | E. of HC Road |  | 59 | 84 | 83 | 49 | 1071 | 1067 | 0 | 0 | 82 | 108 | 0 | 0 |
|  | W. of HC Road |  | 85 | 146 | 180 | 121 | 2156 | 2059 | 33 | 25 | 88 | 83 | 14 | 5 |
| Tecumseh Rd | E. of HC Road |  | 331 | 321 | 437 | 511 | 5994 | 6325 | 138 | 147 | 218 | 393 | 0 | 396 |
|  | W. of HC Road |  | 248 | 492 | 517 | 391 | 6561 | 7108 | 0 | 0 | 68 | 112 | 0 | 0 |
| Dorchester St | E. of HC Road |  | 78 | 86 | 135 | 97 | 1626 | 1428 | 0 | 0 | 119 | 158 | 0 | 0 |
|  | W. of HC Road |  | 76 | 46 | 86 | 52 | 1332 | 763 | 22 | 9 | 55 | 32 | 13 | 2 |
| Prince Rd/Totten St | E. of HC Road |  | 160 | 101 | 129 | 208 | 2265 | 2560 | 0 | 0 | 93 | 109 | 0 | 0 |
|  | W. of HC Road |  | 240 | 321 | 354 | 353 | 5081 | 5466 | 0 | 0 | 92 | 80 | 0 | 0 |
| Malden Rd | E. of HC Road |  | 117 | 70 | 113 | 111 | 1536 | 1195 | 0 | 0 | 323 | 347 | 0 | 0 |
|  | W. of HC Road |  | 429 | 431 | 485 | 432 | 6860 | 6247 | 378 | 324 | 603 | 45 | 246 | 701 |
| Industrial Rd | E. of HC Road |  | 300 | 160 | 232 | 305 | 4094 | 3789 | 50 | 59 | 163 | 155 | 10 | 27 |
|  | W. of HC Road |  | 306 | 113 | 198 | 304 | 4301 | 3096 | 180 | 222 | 0 | 0 | 0 | 279 |
| EC Row N. Ramp Termina | E-N/S Ramp |  | 1060 | n/a | 1009 | n/a | 15484 | n/a | 339 | n/a | 964 | n/a | 0 | n/a |
|  | N-W Ramp |  | n/a | 32 | n/a | 86 | n/a | 923 | n/a | 8 | n/a | 56 | n/a | 0 |
|  | S-W Ramp |  | 41 | n/a | 140 | n/a | 848 | n/a | 28 | n/a | 97 | n/a | 719 | n/a |
| EC Row S. Ramp Termina | W-N/S Ramp |  | n/a | 170 | n/a | 175 | n/a | 2475 | n/a | 72 | n/a | 313 | n/a | 0 |
|  | S/N-E Ramp |  | n/a | 1218 | n/a | 1317 | n/a | 19144 | n/a | 408 | n/a | 1181 | n/a | 480 |
| Spring Gdn Rd | W. of HC Road |  | 120 | 210 | 150 | 170 | 2365 | 3114 | 0 | 0 | 2 | 4 | 0 | 0 |
| Labelle St | E. of HC Road |  | 349 | 189 | 176 | 49 | 3780 | 1987 | 0 | 0 | 311 | 153 | 0 | 0 |
| Lambton St/Grand Marais | E. of HC Road |  | 420 | 285 | 234 | 351 | 5188 | 5305 | 0 | 0 | 83 | 238 | 0 | 0 |
| Pulford St | E. of HC Road |  | 187 | 117 | 189 | 51 | 2777 | 1362 | 0 | 0 | 226 | 132 | 0 | 0 |
| Cabana Rd | E. of HC Road |  | 650 | 437 | 529 | 588 | 8553 | 8024 | 0 | 0 | 723 | 778 | 0 | 0 |
|  | between HC Road and Hwy 401 NB Ramps |  | 1567 | 1415 | 1132 | 1350 | 23872 | 20643 | 69 | 167 | 0 | 2098 | 0 | 101 |
| Todd Lane | between Hwy 401 NB and SB Ramps |  | 880 | 1064 | 671 | 1290 | 13704 | 17541 | 40 | 147 | 0 | 1984 | 0 | 76 |
|  | W. of Hwy 401 SB Ramps |  | 723 | 538 | 670 | 705 | 12278 | 10327 | 0 | 0 | 14 | 11 | 0 | 0 |
| Huron Church Line | W. of HC Road |  | 395 | 601 | 803 | 528 | 9820 | 8436 | 129 | 117 | 353 | 464 | 38 | 5 |
| St Clair College | E. of Talbot Road |  | 124 | 807 | 261 | 286 | 3145 | 9460 | 0 | 0 | 70 | 342 | 0 | 0 |
| Cousineau Dr | E. of Talbot Road |  | 258 | 312 | 515 | 340 | 6502 | 4608 | 0 | 0 | 0 | 1120 | 0 | 0 |
|  | W. of Talbot Road |  | 284 | 490 | 737 | 439 | 8828 | 7603 | 4 | 47 | 0 | 0 | 0 | 0 |
| Howard Ave | E. of Talbot Rd |  | 499 | 650 | 543 | 839 | 8664 | 12148 | 149 | 234 | 288 | 3 | 11 | 0 |
|  | betweem Talbot Road and Hwy 401 SB On-Ramp |  | 858 | 376 | 1212 | 590 | 15994 | 8255 | 255 | 178 | 1085 | 1 | 18 | 0 |
|  | W. of Hwy 401 SB On-Ramp |  | 654 | 999 | 1180 | 1085 | 15057 | 17650 | 299 | 364 | 22 | 264 | 0 | 0 |
| E.C. Row Expressway | E. of Huron Church Rd |  | 3313 | 2955 | 2733 | 3551 | 42939 | 48782 | 909 | 1047 | 3516 | 4839 | 709 | 2001 |
|  | At Malden Rd |  | 2326 | 1907 | 1950 | 2409 | 29372 | 30979 | 615 | 673 | 2797 | 4237 | 984 | 1518 |
|  | W. of Matchette |  | 1676 | 480 | 1433 | 640 | 24849 | 9435 | 598 | 375 | 24 | 0 | 130 | 0 |

[^3]

| LOCATION | SECTION |  | Alternative 2A |  |  |  | 24 Hour AADT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 2015 |  |  |  | Local Cars |  | Local Trucks |  | International Cars |  | International Trucks |  |
|  |  |  | AM PEAK HOUR |  | PM PEAK HOUR |  |  |  |  |  |  |  |  |  |
|  | FROM | TO | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB/EB | NB / WB | SB / EB | NB/WB | SB / EB |
| HC Road | Riverside | University |  |  |  |  | 6763 | 5390 | 173 | 85 | 3 | 1 | 0 | 3 |
|  | University | Wyandotte |  |  |  |  | 3111 | 3651 | 90 | 119 | 65 | 198 | 20 | 3 |
|  | Wyandotte | AMB Off Ramp |  |  |  |  | 2211 | 3045 | 0 | 0 | 44 | 131 | 0 | 0 |
|  | AMB Off Ramp | College |  |  |  |  | 8806 | 6251 | 231 | 96 | 6173 | 1 | 2162 | 0 |
|  | College St | Girardot St | 1797 | 834 | 1389 | 1656 | 17474 | 16649 | 522 | 486 | 6193 | 4360 | 192 | 2209 |
|  | Girardot St | Tecumseh Rd | 1720 | 817 | 1275 | 1536 | 17362 | 17487 | 621 | 567 | 5517 | 3775 | 176 | 2114 |
|  | Tecumseh Rd | Dorchester St | 1789 | 1119 | 1632 | 1717 | 20829 | 21671 | 768 | 720 | 5325 | 3452 | 181 | 1987 |
|  | Dorchester St | Prince Rd/Totten St | 1831 | 1163 | 1593 | 1799 | 21353 | 23522 | 685 | 653 | 4826 | 3173 | 156 | 1799 |
|  | Prince Rd/Totten St | Malden Rd | 1934 | 1384 | 1806 | 1952 | 23703 | 26772 | 763 | 750 | 4676 | 3162 | 177 | 1660 |
|  | Malden Rd | Industrial Rd | 1588 | 1142 | 1466 | 1623 | 18911 | 21704 | 573 | 566 | 4737 | 3352 | 7 | 1601 |
|  | Industrial Rd | EC Row N. Ramp Terminal | 1703 | 1199 | 1551 | 1743 | 20829 | 23841 | 635 | 632 | 4457 | 3024 | 0 | 1593 |
|  | EC Row N. Ramp Terminal | EC Row S. Ramp Terminal | 1190 | 1453 | 1223 | 2080 | 15536 | 30518 | 425 | 631 | 3900 | 2716 | 0 | 1307 |
|  | EC Row S. Ramp Terminal | Highway 401 Offramp | 1612 | 1082 | 1638 | 1555 | 23080 | 22647 | 449 | 380 | 2181 | 2159 | 0 | 1032 |
|  | Highway 401 Offramp | Spring Gdn Rd/Labelle St | 1055 | 412 | 406 | 430 | 10669 | 6674 | 78 | 97 | 654 | 564 | 0 | 0 |
|  | Spring Gdn Rd/Labelle St | Lambton St/Grand Marais Rd Ramp | 878 | 384 | 342 | 465 | 9076 | 6728 | 111 | 134 | 413 | 516 | 0 | 0 |
|  | Lambton St/Grand Marais Rd P | Pulford St | 675 | 374 | 354 | 396 | 7863 | 6452 | 136 | 37 | 209 | 240 | 0 | 0 |
|  | Pulford St | Todd Ln/Cabana Rd | 648 | 422 | 359 | 417 | 8033 | 7308 | 152 | 50 | 0 | 37 | 0 | 0 |
|  | Todd Ln/Cabana Rd | Huron Church Line | 768 | 627 | 628 | 894 | 10252 | 12558 | 126 | 153 | 697 | 461 | 0 | 54 |
| Talbot Rd | Huron Church Line | St Clair College | 403 | 503 | 418 | 376 | 5522 | 6832 | 8 | 0 | 884 | 850 | 0 | 0 |
|  | St Clair College | Cousineau Dr | 858 | 303 | 464 | 429 | 8455 | 6404 | 10 | 0 | 1430 | 0 | 0 | 0 |
|  | Cousineau Dr | Howard Ave | 723 | 322 | 485 | 298 | 9730 | 5472 | 21 | 0 | 81 | 0 | 0 | 0 |
|  | Howard Ave | Highway 3 split | 982 | 789 | 862 | 711 | 14885 | 12943 | 291 | 278 | 0 | 25 | 0 | 0 |
| Ojibway Pwy | EC Row Expressway | GN Booth Dr | 665 | 409 | 570 | 790 | 9922 | 10623 | 136 | 139 | 32 | 15 | 90 | 430 |
|  | GN Booth Dr | Sandwich St | 650 | 411 | 576 | 765 | 9857 | 10411 | 136 | 136 | 32 | 15 | 91 | 432 |
|  | Sandwich St | Prospect Ave | 615 | 381 | 541 | 738 | 9360 | 9610 | 75 | 78 | 52 | 35 | 0 | 0 |
|  | N. of Prospect Ave |  | 608 | 370 | 540 | 725 | 9298 | 9402 | 74 | 76 | 52 | 34 | 0 | 0 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| CROSSING ROADS |  |  | WB | EB | WB | EB |  |  |  |  |  |  |  |  |
| Wyandotte | W of HuronChurch |  |  |  |  |  | 4834 | 4441 | 0 | 0 | 358 | 431 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2850 | 4074 | 21 | 137 | 726 | 867 | 18 | 0 |
| University | W of HuronChurch |  |  |  |  |  | 1267 | 1128 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 1950 | 2014 | 119 | 90 | 67 | 21 | 3 | 20 |
| Riverside | W of HuronChurch |  |  |  |  |  | 3367 | 3642 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 6608 | 5719 | 0 | 0 | 132 | 45 | 0 | 0 |
| AMB Off Ramp | E of HuronChurch |  |  |  |  |  | 0 | 1643 | 0 | 60 | 0 | 6170 | 3 | 2162 |
| AMB On Ramp | E of HuronChurch |  |  |  |  |  | 317 | 0 | 12 | 0 | 5799 | 0 | 174 | 0 |
| Patricia | AMB | Wyandotte |  |  |  |  | 567 | 1642 | 22 | 58 | 3342 | 3241 | 171 | 231 |
| College St | E. of HC Road |  | 312 | 355 | 483 | 387 | 6465 | 5630 | 175 | 132 | 3 | 532 | 0 | 150 |
|  | W. of HC Road |  | 75 | 35 | 135 | 64 | 1566 | 746 | 0 | 0 | 210 | 59 | 0 | 0 |
| Girardot St | E. of HC Road |  | 51 | 86 | 86 | 45 | 1060 | 1052 | 0 | 0 | 68 | 109 | 0 | 0 |
|  | W. of HC Road |  | 81 | 153 | 191 | 126 | 2178 | 2149 | 36 | 29 | 114 | 91 | 9 | 4 |
| Tecumseh Rd | E. of HC Road |  | 315 | 316 | 409 | 462 | 5638 | 6115 | 134 | 150 | 211 | 357 | 0 | 137 |
|  | W. of HC Road |  | 241 | 465 | 505 | 385 | 6296 | 6802 | 0 | 0 | 147 | 125 | 0 | 0 |
| Dorchester St |  |  | 75 | 84 | 132 | 93 | 1605 | 1382 | 0 | 0 | 102 | 155 | 0 | 0 |
|  | W. of HC Road |  | 76 | 46 | 86 | 52 | 1321 | 762 | 23 | 10 | 64 | 34 | 8 | 1 |
| Prince Rd/Totten St |  |  | 144 | 108 | 115 | 206 | 2040 | 2591 | 0 | 0 | 73 | 121 | 0 | 0 |
|  | E. of HC Road W. of HC Road |  | 235 | 299 | 322 | 337 | 4778 | 5165 | 0 | 0 | 82 | 73 | 0 | 0 |
| Malden Rd |  |  | 84 | 55 | 85 | 74 | 1175 | 931 | 0 | 0 | 201 | 183 | 0 | 0 |
|  | E. of HC Road <br> W. of HC Road |  | 373 | 403 | 479 | 432 | 6483 | 6223 | 368 | 337 | 484 | 36 | 168 | 431 |
| Industrial Rd | E. of HC Road |  | 244 | 142 | 202 | 257 | 3437 | 3215 | 46 | 55 | 146 | 164 | 6 | 16 |
|  | W. of HC Road |  | 290 | 92 | 148 | 275 | 3764 | 2787 | 147 | 191 | 0 | 0 | 0 | 179 |
| EC Row N. Ramp Terminal | E-N/S Ramp |  | 843 | n/a | 870 | n/a | 12487 | n/a | 266 | n/a | 987 | n/a | 0 | n/a |
|  | N-W Ramp |  | n/a | 29 | n/a | 63 | n/a | 731 | n/a | 0 | n/a | 37 | n/a | 0 |
|  | S-W Ramp |  | 40 | n/a | 130 | n/a | 913 | n/a | 28 | n/a | 92 | n/a | 528 | n/a |
| EC Row S. Ramp Terminal | W-N/S Ramp |  | n/a | 148 | n/a | 188 | n/a | 2461 | n/a | 47 | n/a | 297 | n/a | 0 |
|  | S/N-E Ramp |  | n/a | 941 | n/a | 1083 | n/a | 15627 | n/a | 296 | n/a | 762 | n/a | 224 |
| Spring Gdn Rd | W. of HC Road |  | 115 | 210 | 150 | 170 | 2320 | 3114 | 0 | 0 | 2 | 4 | 0 | 0 |
| Labelle St | E. of HC Road |  | 293 | 144 | 167 | 66 | 3338 | 1722 | 0 | 0 | 268 | 144 | 0 | 0 |
| Lambton St/Grand Marais Rd |  |  | 305 | 205 | 224 | 205 | 4221 | 3436 | 0 | 0 | 75 | 153 | 0 | 0 |
|  | W. of HC Road |  | 110 | 205 | 160 | 90 | 2283 | 2301 | 34 | 21 | 37 | 38 | 0 | 0 |
| Pulford St | E. of HC Road |  | 203 | 128 | 207 | 91 | 3090 | 1739 | 0 | 0 | 211 | 186 | 0 | 0 |
| Cabana Rd | E. of HC Road |  | 550 | 324 | 467 | 444 | 7381 | 6096 | 0 | 0 | 646 | 518 | 0 | 0 |
|  | between HC Road and Hwy 401 NB Ramps |  | 1308 | 1175 | 928 | 1124 | 19803 | 17449 | 39 | 127 | 0 | 1440 | 0 | 66 |
| Todd Lane | between Hwy 401 NB and SB Ramps |  | 773 | 849 | 586 | 1122 | 12020 | 14962 | 24 | 113 | 0 | 1413 | 0 | 48 |
|  | W. of Hwy 401 SB Ramps |  | 675 | 527 | 613 | 664 | 11358 | 9885 | 0 | 0 | 12 | 10 | 0 | 0 |
| Huron Church Line | W. of HC Road |  | 251 | 454 | 614 | 306 | 7091 | 5657 | 97 | 80 | 234 | 294 | 22 | 0 |
| St Clair College | E. of Talbot Road |  | 150 | 805 | 234 | 267 | 2984 | 9148 | 0 | 0 | 160 | 451 | 0 | 0 |
| Cousineau Dr | E. of Talbot Road |  | 268 | 295 | 387 | 300 | 5464 | 4248 | 0 | 0 | 0 | 985 | 0 | 0 |
|  | W. of Talbot Road |  | 246 | 408 | 589 | 365 | 7233 | 6365 | 0 | 0 | 0 | 0 | 0 | 0 |
| Howard Ave | E. of Talbot Rd |  | 403 | 551 | 532 | 718 | 7723 | 10360 | 128 | 195 | 283 | 2 | 0 | 0 |
|  | betweem Talbot Road and Hwy 401 SB On-Ramp |  | 707 | 315 | 1015 | 526 | 13285 | 7183 | 208 | 150 | 951 | 0 | 0 | 0 |
|  | W. of Hwy 401 SB On-Ramp |  | 520 | 790 | 976 | 886 | 12302 | 14210 | 247 | 290 | 11 | 192 | 0 | 0 |
| E.C. Row Expressway | E. of Huron Church Rd |  | 2154 | 2268 | 2167 | 2873 | 31509 | 39193 | 678 | 807 | 2214 | 3455 | 390 | 1241 |
|  | At Malden Rd |  | 1380 | 1475 | 1490 | 1978 | 20975 | 25038 | 452 | 548 | 1325 | 3220 | 545 | 1052 |
|  | W. of Matchette |  | 960 | 412 | 980 | 520 | 15597 | 7773 | 364 | 401 | 23 | 0 | 66 | 0 |

[^4]

| LOCATION | SECTION |  |  |  |  |  | 24 Hour AADT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 2025 |  |  |  | Local Cars |  | Local Trucks |  | International Cars |  | International Trucks |  |
|  |  |  | AM PEAK HOUR PM PEAK HOUR |  |  |  |  |  |  |  |  |  |  |  |
|  | FROM | TO | NB/ WB | SB / EB | NB / WB | SB/EB | NB/WB | SB/EB | NB/WB | SB/EB | NB/WB | SB / EB | NB/WB | SB/EB |
| HC Road | Riverside | University |  |  |  |  | 6879 | 5510 | 180 | 92 | 3 | 1 | 0 | 41 |
|  | University | Wyandotte |  |  |  |  | 3138 | 3726 | 88 | 124 | 67 | 191 | 56 | 41 |
|  | Wyandotte | AMB Off Ramp |  |  |  |  | 2188 | 3074 | 0 | 0 | 45 | 128 | 0 | 0 |
|  | AMB Off Ramp | College |  |  |  |  | 8649 | 6379 | 229 | 103 | 6543 | 1 | 2797 | 0 |
|  | College St | Girardot St | 1812 | 878 | 1461 | 1778 | 17875 | 17283 | 550 | 519 | 6510 | 4605 | 245 | 2918 |
|  | Girardot St | Tecumseh Rd | 1703 | 849 | 1348 | 1627 | 17721 | 18123 | 662 | 603 | 5770 | 3971 | 226 | 2741 |
|  | Tecumseh Rd | Dorchester St | 1772 | 1123 | 1711 | 1806 | 21215 | 22028 | 818 | 751 | 5543 | 3513 | 230 | 2420 |
|  | Dorchester St | Prince Rd/Totten St | 1853 | 1201 | 1676 | 1895 | 22001 | 24262 | 734 | 690 | 5064 | 3277 | 200 | 2247 |
|  | Prince Rd/Totten St | Malden Rd | 1952 | 1453 | 1905 | 2090 | 24456 | 28088 | 820 | 807 | 4895 | 3216 | 228 | 2022 |
|  | Malden Rd | Industrial Rd | 1607 | 1157 | 1550 | 1778 | 19557 | 22795 | 629 | 615 | 4986 | 3431 | 9 | 1979 |
|  | Industrial Rd | EC Row N. Ramp Terminal | 1744 | 1255 | 1710 | 1866 | 22312 | 25206 | 711 | 686 | 4773 | 3036 | 0 | 1963 |
|  | EC Row N. Ramp Terminal | EC Row S. Ramp Terminal | 1200 | 1653 | 1352 | 2350 | 16550 | 34662 | 480 | 730 | 4187 | 2860 | 0 | 1700 |
|  | EC Row S. Ramp Terminal | Highway 401 Offramp | 1777 | 1250 | 1848 | 1802 | 25613 | 26275 | 527 | 460 | 2559 | 2338 | 0 | 1316 |
|  | Highway 401 Offramp | Spring Gdn Rd/Labelle St | 1110 | 562 | 434 | 450 | 11292 | 8073 | 89 | 137 | 688 | 666 | 0 | 0 |
|  | Spring Gdn Rd/Labelle St | Lambton St/Grand Marais Rd Ramp | 927 | 535 | 372 | 465 | 9687 | 7975 | 134 | 190 | 431 | 596 | 0 | 0 |
|  | Lambton St/Grand Marais Rd Ra | Pulford St | 720 | 475 | 354 | 445 | 8242 | 7776 | 152 | 50 | 183 | 246 | 0 | 0 |
|  | Pulford St | Todd Ln/Cabana Rd | 679 | 512 | 403 | 472 | 8646 | 8568 | 165 | 62 | 0 | 56 | 0 | 0 |
|  | Todd Ln/Cabana Rd | Huron Church Line | 839 | 737 | 768 | 954 | 11835 | 14051 | 155 | 158 | 801 | 470 | 0 | 66 |
| Talbot Rd | Huron Church Line | St Clair College | 440 | 501 | 444 | 403 | 5765 | 7109 | 13 | 0 | 1087 | 787 | 0 | 0 |
|  | St Clair College | Cousineau Dr | 881 | 319 | 516 | 484 | 8799 | 7016 | 12 | 0 | 1628 | 0 | 0 | 0 |
|  | Cousineau Dr | Howard Ave | 736 | 358 | 552 | 341 | 10300 | 6162 | 46 | 3 | 138 | 0 | 0 | 0 |
|  | Howard Ave | Highway 3 split | 1028 | 900 | 964 | 786 | 16103 | 14544 | 320 | 315 | 0 | 41 | 0 | 0 |
| Ojibway Pwy | EC Row Expressway | GN Booth Dr | 700 | 440 | 620 | 820 | 10615 | 10841 | 139 | 134 | 27 | 21 | 113 | 531 |
|  | GN Booth Dr | Sandwich St | 685 | 442 | 626 | 795 | 10550 | 10636 | 139 | 132 | 27 | 21 | 114 | 533 |
|  | Sandwich St | Prospect Ave | 646 | 403 | 582 | 760 | 9966 | 9995 | 74 | 74 | 48 | 41 | 0 | 0 |
|  | N. of Prospect Ave |  | 639 | 392 | 581 | 747 | 9904 | 9787 | 73 | 72 | 48 | 40 | 0 | 0 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| CROSSING ROADS |  |  | WB | EB | WB | EB |  |  |  |  |  |  |  |  |
| Wyandotte | W of HuronChurch |  |  |  |  |  | 4742 | 4425 | 0 | 0 | 370 | 436 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2809 | 4146 | 18 | 141 | 749 | 860 | 51 | 0 |
| University | W of HuronChurch |  |  |  |  |  | 1358 | 1254 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2083 | 2192 | 124 | 88 | 63 | 21 | 41 | 56 |
| Riverside | W of HuronChurch |  |  |  |  |  | 3547 | 3769 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 6822 | 5807 | 0 | 0 | 128 | 46 | 0 | 0 |
| AMB Off Ramp | E of HuronChurch |  |  |  |  |  | 0 | 1407 | 0 | 54 | 0 | 6540 | 41 | 2797 |
| AMB On Ramp | E of HuronChurch |  |  |  |  |  | 276 | 0 | 7 | 0 | 6098 | - | 223 | 0 |
| Patricia | AMB | Wyandotte |  |  |  |  | 490 | 1407 | 14 | 51 | 3456 | 3338 | 193 | 305 |
| College St |  |  | 321 | 356 | 487 | 404 | 6567 | 5748 | 178 | 133 | 3 | 534 | 0 | 193 |
|  | W. of HC Road |  | 84 | 43 | 163 | 67 | 1466 | 857 | 0 | 0 | 546 | 43 | 0 | 0 |
| Girardot St |  |  | 54 | 84 | 84 | 48 | 1065 | 1057 | 0 | 0 | 66 | 110 | 0 | 0 |
|  | W. of HC Road |  | 82 | 149 | 183 | 122 | 2131 | 2092 | 33 | 27 | 108 | 83 | 11 | 4 |
| Tecumseh Rd | E. of HC Road |  | 318 | 310 | 381 | 468 | 5430 | 5973 | 128 | 147 | 200 | 400 | 0 | 245 |
|  | W. of HC Road |  | 237 | 460 | 494 | 379 | 6273 | 6703 | 0 | 0 | 62 | 130 | 0 | 0 |
| Dorchester St |  |  | 77 | 85 | 134 | 95 | 1638 | 1406 | 0 | 0 | 101 | 156 | 0 | 0 |
|  | W. of HC Road |  | 76 | 46 | 86 | 52 | 1323 | 764 | 22 | 10 | 63 | 31 | 10 | 2 |
| Prince Rd/Totten St |  |  | 154 | 101 | 123 | 210 | 2177 | 2562 | 0 | 0 | 83 | 121 | 0 | 0 |
|  | W. of HC Road |  | 241 | 315 | 335 | 344 | 4937 | 5354 | 0 | 0 | 87 | 73 | 0 | 0 |
| Malden Rd |  |  | 102 | 63 | 105 | 94 | 1422 | 1098 | 0 | 0 | 259 | 250 | 0 | 0 |
|  | W. of HC Road |  | 398 | 413 | 486 | 421 | 6763 | 6083 | 368 | 315 | 461 | 90 | 206 | 562 |
| Industrial Rd | E. of HC Road |  | 267 | 151 | 213 | 296 | 3702 | 3605 | 47 | 56 | 146 | 180 | 8 | 21 |
|  | W. of HC Road |  | 299 | 103 | 185 | 297 | 4136 | 3003 | 171 | 199 | 0 | 0 | 0 | 250 |
| EC Row N. Ramp Termina | E-N/S Ramp |  | 993 | n/a | 1002 | n/a | 14751 | n/a | 313 | n/a | 1035 | n/a | 0 | n/a |
|  | N-W Ramp |  | n/a | 30 | n/a | 76 | n/a | 846 | n/a | 0 | n/a | 43 | n/a | 0 |
|  | S-W Ramp |  | 40 | n/a | 137 | n/a | 882 | n/a | 31 | n/a | 91 | n/a | 638 | n/a |
| EC Row S. Ramp Terminal | W-N/S Ramp |  | n/a | 154 | n/a | 194 | n/a | 2478 | n/a | 63 | n/a | 369 | n/a | 0 |
|  | S/N-E Ramp |  | n/a | 1134 | n/a | 1194 | n/a | 17986 | n/a | 362 | n/a | 704 | n/a | 344 |
| Spring Gdn Rd | W. of HC Road |  | 115 | 210 | 150 | 170 | 2320 | 3114 | 0 | 0 | 2 | 4 | 0 | 0 |
| Labelle St | E. of HC Road |  | 330 | 174 | 168 | 56 | 3589 | 1905 | 0 | 0 | 294 | 154 | 0 | 0 |
| Lambton St/Grand Marais | E. of HC Road |  | 305 | 230 | 224 | 205 | 4229 | 3669 | 0 | 0 | 69 | 152 | 0 | 0 |
|  | W. of HC Road |  | 135 | 205 | 160 | 90 | 2492 | 2306 | 45 | 22 | 44 | 35 | 0 | 0 |
| Pulford St | E. of HC Road |  | 186 | 122 | 184 | 48 | 2788 | 1379 | 0 | 0 | 186 | 138 | 0 | 0 |
| Cabana Rd | E. of HC Road |  | 611 | 369 | 475 | 520 | 7859 | 7079 | 0 | 0 | 674 | 574 | 0 | 0 |
|  | between HC Road and Hwy 401 NB Ramps |  | 1444 | 1271 | 1033 | 1206 | 21920 | 18772 | 56 | 136 | 0 | 1574 | 0 | 82 |
| Todd Lane | between Hwy 401 NB and SB Ramps |  | 801 | 976 | 659 | 1186 | 12884 | 16383 | 33 | 123 | 0 | 1528 | 0 | 63 |
|  | W. of Hwy 401 SB Ramps |  | 689 | 533 | 694 | 682 | 12168 | 10089 | 0 | 0 | 14 | 10 | 0 | 0 |
| Huron Church Line | W. of HC Road |  | 324 | 508 | 671 | 444 | 8180 | 7131 | 100 | 102 | 278 | 378 | 29 | 0 |
| St Clair College | E. of Talbot Road |  | 150 | 785 | 246 | 281 | 3141 | 9151 | 0 | 0 | 112 | 396 | 0 | 0 |
| Cousineau Dr | E. of Talbot Road |  | 250 | 290 | 431 | 339 | 5707 | 4478 | 0 | 0 | 0 | 1034 | 0 | 0 |
|  | W. of Talbot Road |  | 265 | 448 | 679 | 402 | 8163 | 6968 | 4 | 31 | 0 | 0 | 0 | 0 |
| Howard Ave | E. of Talbot Rd |  | 453 | 607 | 552 | 778 | 8327 | 11296 | 138 | 218 | 289 | 2 | 10 | 0 |
|  | betweem Talbot Road and Hwy 401 SB On-Ramp |  | 747 | 350 | 1116 | 570 | 14389 | 7858 | 223 | 168 | 1019 | 1 | 15 | 0 |
|  | W. of Hwy 401 SB On-Ramp |  | 550 | 914 | 1075 | 981 | 13377 | 16078 | 261 | 331 | 15 | 219 | 0 | 0 |
| E.C. Row Expressway | E. of Huron Church Rd |  | 2769 | 2559 | 2604 | 3237 | 38745 | 43645 | 822 | 917 | 2922 | 4191 | 575 | 1644 |
|  | At Malden Rd |  | 1846 | 1579 | 1815 | 2237 | 25989 | 26982 | 552 | 588 | 2030 | 4004 | 810 | 1315 |
|  | W. of Matchette |  | 1310 | 450 | 1217 | 580 | 20264 | 8624 | 472 | 404 | 23 | 0 | 96 |  |

[^5]


[^6]

| LOCATION | SECTION |  | Alternative 3 |  |  |  | 24 Hour AADT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 2015 |  |  |  | Local Cars |  | Local Trucks |  | International Cars |  | International Trucks |  |
|  |  |  | AM PEAK HOUR |  | PM PEAK HOUR |  |  |  |  |  |  |  |  |  |
|  | FROM | TO | NB | SB | NB | SB | NB/ WB | SB / EB | NB/WB | SB/EB | NB/WB | SB / EB | NB/WB | SB/EB |
| HC Road | Riverside | University |  |  |  |  | 6736 | 5308 | 180 | 84 | 3 | 1 | 0 | 3 |
|  | University | Wyandotte |  |  |  |  | 3024 | 3558 | 91 | 118 | 58 | 241 | 20 | 3 |
|  | Wyandotte | AMB Off Ramp |  |  |  |  | 2222 | 2957 | 0 | 0 | 37 | 171 | 0 | 0 |
|  | AMB Off Ramp | College |  |  |  |  | 8545 | 6153 | 234 | 95 | 6352 | 1 | 2392 | 0 |
|  | College St | Girardot St | 1782 | 789 | 1381 | 1686 | 17294 | 16101 | 520 | 467 | 6189 | 4444 | 192 | 2437 |
|  | Girardot St | Tecumseh Rd | 1675 | 794 | 1245 | 1557 | 16858 | 17155 | 609 | 575 | 5424 | 3869 | 173 | 2365 |
|  | Tecumseh Rd | Dorchester St | 1734 | 1054 | 1585 | 1727 | 20136 | 20850 | 750 | 712 | 5219 | 3451 | 178 | 2173 |
|  | Dorchester St | Prince Rd/Totten St | 1801 | 1124 | 1573 | 1781 | 20995 | 22729 | 677 | 658 | 4796 | 3168 | 155 | 2001 |
|  | Prince Rd/Totten St | Malden Rd | 1907 | 1334 | 1787 | 1983 | 23361 | 26152 | 756 | 763 | 4660 | 3177 | 177 | 1962 |
|  | Malden Rd | Industrial Rd | 1588 | 1051 | 1413 | 1644 | 18327 | 20673 | 559 | 560 | 4752 | 3336 | 7 | 1819 |
|  | Industrial Rd | EC Row N. Ramp Terminal | 1700 | 1152 | 1513 | 1790 | 20368 | 23372 | 624 | 646 | 4486 | 3098 | 0 | 1866 |
|  | EC Row N. Ramp Terminal | EC Row S. Ramp Terminal | 1180 | 1443 | 1195 | 2230 | 15128 | 31177 | 429 | 668 | 3945 | 2935 | 0 | 1615 |
|  | EC Row S. Ramp Terminal | Spring Gdn Rd/Labelle St | 1629 | 1153 | 1531 | 1720 | 22018 | 24061 | 423 | 419 | 2198 | 2594 | 0 | 1381 |
|  | Spring Gdn Rd/Labelle St | Lambton St/Grand Marais Rd | 1327 | 379 | 1492 | 944 | 20998 | 10369 | 300 | 174 | 1392 | 745 | 0 | 0 |
|  | Lambton St/Grand Marais Rd | Pulford St | 849 | 457 | 528 | 941 | 9870 | 11158 | 149 | 154 | 740 | 677 | 0 | 0 |
|  | Pulford St | Todd Ln/Cabana Rd | 849 | 524 | 540 | 951 | 10093 | 12039 | 168 | 175 | 650 | 512 | 0 | 0 |
|  | Todd Ln/Cabana Rd | Huron Church Line | 706 | 492 | 608 | 862 | 10200 | 11413 | 103 | 95 | 311 | 235 | 0 | 0 |
| Talbot Rd | Huron Church Line | St Clair College | 423 | 703 | 406 | 847 | 6786 | 13447 | 54 | 76 | 0 | 82 | 0 | 0 |
|  | St Clair College | Cousineau Dr | 1130 | 389 | 744 | 508 | 12069 | 6110 | 125 | 80 | 2049 | 1326 | 0 | 153 |
|  | Cousineau Dr | Howard Ave | 400 | 346 | 410 | 288 | 5810 | 4825 | 95 | 93 | 528 | 488 | 0 | 176 |
|  | S. of Howard Ave |  | 679 | 746 | 940 | 670 | 13212 | 12253 | 279 | 252 | 0 | 0 | 0 | 0 |
| Ojibway Pwy | EC Row Expressway | GN Booth Dr | 634 | 409 | 570 | 790 | 9780 | 9982 | 134 | 135 | 0 | 14 | 0 | 425 |
|  | GN Booth Dr | Sandwich St | 618 | 413 | 577 | 763 | 9717 | 9790 | 133 | 132 | 0 | 14 | 0 | 429 |
|  | Sandwich St | Prospect Ave | 578 | 385 | 539 | 729 | 9064 | 9557 | 72 | 78 | 46 | 46 | 0 | 0 |
|  | N. of Prospect Ave |  | 571 | 376 | 538 | 715 | 9002 | 9359 | 72 | 76 | 46 | 45 | 0 | 0 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| CROSSING ROADS |  |  | WB | EB | WB | EB | NB/ WB | SB / EB | NB/WB | SB/EB | NB/WB | SB / EB | NB/WB | SB/EB |
| Wyandotte | W of HuronChurch |  |  |  |  |  | 4817 | 4431 | 0 | 0 | 359 | 438 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2818 | 4098 | 21 | 142 | 722 | 939 | 18 | 0 |
| University | W of HuronChurch |  |  |  |  |  | 1264 | 1185 | 0 | 0 | 0 | 0 | 0 | 0 |
| University | E of HuronChurch |  |  |  |  |  | 1938 | 1976 | 118 | 91 | 70 | 21 | 3 | 20 |
|  | W of HuronChurch |  |  |  |  |  | 3427 | 3539 | 0 | 0 | 0 | 0 | 0 | 0 |
| Riverside | E of HuronChurch |  |  |  |  |  | 6590 | 5621 | 0 | 0 | 171 | 37 | 0 | 0 |
| AMB Off Ramp | E of HuronChurch |  |  |  |  |  | 0 | 1379 | 0 | 56 | 0 | 6349 | 3 | 2392 |
| AMB On Ramp | E of HuronChurch |  |  |  |  |  | 305 | 0 | 11 | 0 | 5828 | 0 | 174 | 0 |
| Patricia | AMB Wyandotte <br> E. of HC Road  |  |  |  |  |  | 545 | 1376 | 22 | 55 | 3372 | 3417 | 171 | 267 |
| College St |  |  | 299 | 339 | 488 | 379 | 6421 | 5429 | 161 | 125 | 3 | 531 | 0 | 141 |
|  | W. of HC Road |  | 79 | 30 | 139 | 62 | 1594 | 709 | 0 | 0 | 245 | 43 | 0 | 0 |
| Girardot St | E. of HC Road |  | 51 | 86 | 86 | 45 | 989 | 996 | 0 | 0 | 131 | 160 | 0 | 0 |
|  | W. of HC Road |  | 81 | 154 | 191 | 126 | 2257 | 2224 | 41 | 25 | 48 | 33 | 0 | 0 |
| Tecumseh Rd | E. of HC Road |  | 309 | 322 | 381 | 456 | 5357 | 6111 | 139 | 146 | 196 | 357 | 0 | 154 |
|  | W. of HC Road |  | 244 | 448 | 504 | 392 | 6265 | 6733 | 0 | 0 | 187 | 124 | 0 | 0 |
| Dorchester St | E. of HC Road |  | 76 | 46 | 86 | 52 | 1142 | 717 | 0 | 0 | 154 | 127 | 0 | 0 |
|  | W. of HC Road |  | 75 | 84 | 132 | 93 | 1730 | 1419 | 32 | 18 | 34 | 20 | 0 | 0 |
| Prince Rd/Totten St | E. of HC Road |  | 138 | 102 | 115 | 202 | 1990 | 2507 | 0 | 0 | 77 | 117 | 0 | 0 |
|  | W. of HC Road |  | 235 | 255 | 301 | 313 | 4607 | 4619 | 0 | 0 | 76 | 68 | 0 | 0 |
| Malden Rd | E. of HC Road |  | 84 | 55 | 85 | 72 | 1174 | 896 | 0 | 0 | 202 | 198 | 0 | 0 |
|  | W. of HC Road |  | 373 | 403 | 456 | 391 | 6338 | 5887 | 362 | 323 | 455 | 28 | 150 | 429 |
| Industrial Rd | E. of HC Road |  | 247 | 146 | 207 | 257 | 3492 | 3232 | 52 | 59 | 149 | 177 | 7 | 17 |
|  | W. of HC Road |  | 292 | 93 | 121 | 275 | 3570 | 2785 | 132 | 193 | 0 | 0 | 0 | 190 |
| EC Row N. Ramp Terminal | E. of HC Road (E-N/S Off Ramp \& S-W On Ramp) |  | 873 | 50 | 940 | 123 | 13121 | 1365 | 274 | 0 | 1120 | 122 | 0 | 0 |
|  | W. of HC Road (N-W On Ramp) |  | 32 | n/a | 60 | n/a | 526 | n/a | 17 | n/a | 57 | n/a | 247 | n/a |
| EC Row S. Ramp Terminal | E. of HC Road (S-E On Ramp) |  | n/a | 525 | n/a | 420 | n/a | 7692 | n/a | 66 | n/a | 0 | n/a | 0 |
|  | W. of HC Road (N-E On Ramp \& W-N/S Off Ramp) |  | 360 | 146 | 630 | 204 | 7615 | 2560 | 257 | 74 | 452 | 295 | 273 | 0 |
| Spring Gdn Rd/Labelle St | E. of HC Road |  | 305 | 145 | 151 | 193 | 3548 | 2853 | 0 | 0 | 95 | 88 | 0 | 0 |
|  | W. of HC Road |  | 107 | 223 | 317 | 276 | 3654 | 4139 | 0 | 0 | 5 | 4 | 0 | 0 |
| Lambton St/Grand Marais Rd | E. of HC Road |  | 332 | 196 | 216 | 245 | 4140 | 3605 | 0 | 0 | 241 | 219 | 0 | 0 |
|  | W. of HC Road |  | 82 | 220 | 152 | 274 | 1953 | 3974 | 36 | 52 | 39 | 54 | 0 | 0 |
| Pulford St | E. of HC Road |  | 189 | 122 | 97 | 102 | 1922 | 1665 | 0 | 0 | 270 | 284 | 0 | 0 |
| Todd Ln/Cabana Rd | E. of HC Road |  | 525 | 441 | 703 | 560 | 10075 | 8720 | 0 | 0 | 95 | 46 | 0 | 0 |
|  | W. of HC Road |  | 513 | 543 | 998 | 698 | 13128 | 10310 | 0 | 0 | 8 | 7 | 0 | 0 |
| Huron Church Line | W. of HC Road |  | 288 | 431 | 650 | 371 | 7770 | 6118 | 96 | 81 | 218 | 247 | 0 | 0 |
| St Clair College | E. of Talbot Road |  | 121 | 774 | 248 | 288 | 2984 | 9189 | 0 | 0 | 94 | 321 | 0 | 0 |
| Cousineau Dr | E. of Talbot Road |  | 411 | 254 | 528 | 420 | 6000 | 4676 | 0 | 0 | 1836 | 1155 | 0 | 0 |
|  | W. of Talbot Road |  | 426 | 457 | 806 | 530 | 10717 | 6675 | 0 | 39 | 0 | 1625 | 0 | 0 |
| Howard Ave | E. of Talbot Road |  | 342 | 472 | 534 | 666 | 7519 | 9311 | 127 | 172 | 2 | 4 | 0 | 0 |
|  | betweem Talbot Road and Hwy 401 SB On Ramp |  | 451 | 337 | 837 | 449 | 10470 | 6729 | 189 | 154 | 111 | 2 | 0 | 0 |
|  | W. of Hwy 401 SB On Ramp |  | 397 | 683 | 754 | 747 | 9389 | 12226 | 205 | 284 | 54 | 56 | 0 | 0 |
| E.C. Row Expressway | E. of Huron Church Rd |  | 2171 | 1874 | 2247 | 2180 | 32171 | 30978 | 681 | 652 | 2331 | 2711 | 398 | 991 |
|  | At Malden Rd |  | 1380 | 1495 | 1490 | 1964 | 21005 | 24885 | 443 | 544 | 1327 | 3432 | 544 | 1020 |
|  | W. of Matchette |  | 960 | 412 | 980 | 520 | 15668 | 7785 | 365 | 388 | 0 | 0 | 0 | 0 |
| * For consistency, Huron Church Rd/Talbot Rd runs North-South and all crossing roads run East-West |  |  |  |  |  |  |  |  |  |  |  |  |  |  |



| LOCATION | SECTION |  | 2025 |  |  |  | 24 Hour AADT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Local Cars |  | Local Trucks |  | International Cars |  | International Trucks |  |
|  |  |  | AM PEAK HOUR PM PEAK HOUR |  |  |  |  |  |  |
|  | FROM | TO |  |  | NB | SB | NB | SB | NB / WB | SB / EB | NB / WB | SB / EB | NB / WB | SB / EB | NB / W | / EB |
| HC Road | Riverside | University |  |  |  |  | 6774 | 5399 | 184 | 90 | 3 | 1 | 0 | 41 |
|  | University | Wyandotte |  |  |  |  | 3000 | 3621 | 90 | 122 | 69 | 236 | 56 | 41 |
|  | Wyandotte | AMB Off Ramp |  |  |  |  | 2176 | 3010 | 0 | 0 | 48 | 166 | 0 | 0 |
|  | AMB Off Ramp | College |  |  |  |  | 8426 | 6251 | 234 | 101 | 7087 | 1 | 2816 | 0 |
|  | College St | Girardot St | 1803 | 848 | 1446 | 1778 | 17631 | 16549 | 546 | 493 | 6532 | 4968 | 246 | 2833 |
|  | Girardot St | Tecumseh Rd | 1733 | 872 | 1337 | 1666 | 17627 | 18139 | 664 | 624 | 5889 | 4413 | 227 | 2782 |
|  | Tecumseh Rd | Dorchester St | 1808 | 1163 | 1680 | 1836 | 20974 | 22098 | 814 | 779 | 5635 | 3919 | 229 | 2535 |
|  | Dorchester St | Prince Rd/Totten St | 1893 | 1221 | 1665 | 1925 | 21963 | 24148 | 736 | 717 | 5181 | 3654 | 201 | 2354 |
|  | Prince Rd/Totten St | Malden Rd | 1984 | 1383 | 1856 | 2150 | 24053 | 27222 | 810 | 813 | 4961 | 3616 | 225 | 2098 |
|  | Malden Rd | Industrial Rd | 1607 | 1077 | 1501 | 1801 | 19004 | 21544 | 618 | 608 | 5015 | 3810 | 9 | 2019 |
|  | Industrial Rd | EC Row N. Ramp Terminal | 1725 | 1103 | 1608 | 1980 | 21211 | 23946 | 680 | 665 | 4704 | 3444 | 0 | 1981 |
|  | EC Row N. Ramp Tern | EC Row S. Ramp Terminal | 1190 | 1498 | 1256 | 2451 | 15562 | 33206 | 464 | 721 | 4124 | 3269 | 0 | 1756 |
|  | EC Row S. Ramp Term | Spring Gdn Rd/Labelle St | 1818 | 1191 | 1788 | 1871 | 25166 | 25543 | 508 | 455 | 2586 | 2788 | 0 | 1424 |
|  | Spring Gdn Rd/Labelle | Lambton St/Grand Marais Rd | 1501 | 586 | 1745 | 955 | 24153 | 12203 | 360 | 196 | 1622 | 862 | 0 | 0 |
|  | Lambton St/Grand Ma | Pulford St | 897 | 678 | 697 | 980 | 11525 | 13341 | 184 | 167 | 848 | 812 | 0 |  |
|  | Pulford St | Todd Ln/Cabana Rd | 904 | 751 | 710 | 987 | 11793 | 14229 | 205 | 185 | 766 | 668 | 0 | 0 |
|  | Todd Ln/Cabana Rd | Huron Church Line | 739 | 718 | 773 | 939 | 11786 | 14038 | 123 | 104 | 372 | 311 | 0 | 0 |
| Talbot Rd | Huron Church Line | St Clair College | 332 | 761 | 452 | 844 | 6453 | 13839 | 76 | 89 | 1 | 179 | 0 | 0 |
|  | St Clair College | Cousineau Dr | 1149 | 407 | 796 | 491 | 12477 | 6120 | 128 | 78 | 2173 | 1318 | 0 | 189 |
|  | Cousineau Dr | Howard Ave | 435 | 340 | 458 | 252 | 6399 | 4536 | 101 | 80 | 585 | 430 | 0 | 196 |
|  | S. of Howard Ave |  | 728 | 844 | 1055 | 729 | 14581 | 13628 | 294 | 275 | 0 | 0 | 0 | 0 |
| Ojibway Pwy | EC Row Expressway | GN Booth Dr | 699 | 440 | 620 | 820 | 10717 | 10440 | 140 | 133 | 0 | 21 | 0 | 533 |
|  | GN Booth Dr | Sandwich St | 683 | 439 | 623 | 793 | 10619 | 10207 | 139 | 130 | 0 | 20 | 0 | 531 |
|  | Sandwich St | Prospect Ave | 643 | 409 | 579 | 750 | 9918 | 9959 | 73 | 74 | 47 | 46 | 0 | 0 |
|  | N. of Prospect Ave |  | 636 | 403 | 578 | 735 | 9856 | 9779 | 72 | 73 | 47 | 45 | 0 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| CROSSING ROADS |  |  | WB | EB | WB | EB | NB / WB | SB / EB | NB/ WB | SB / EB | NB / WB | SB / EB | NB / WB | / EB |
| Wyandotte | W of HuronChurch |  |  |  |  |  | 4743 | 4443 | 0 | 0 | 370 | 439 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2781 | 4182 | 18 | 143 | 754 | 932 | 51 | 0 |
| University | W of HuronChurch |  |  |  |  |  | 1375 | 1229 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2063 | 2040 | 122 | 90 | 70 | 21 | 41 | 56 |
| Riverside | W of HuronChurch |  |  |  |  |  | 3594 | 3728 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 6812 | 5752 | 0 | 0 | 167 | 48 | 0 | 0 |
| AMB Off Ramp | E of HuronChurch |  |  |  |  |  | 0 | 1287 | 0 | 54 | 0 | 7084 | 41 | 2816 |
| AMB On Ramp | E of HuronChurch |  |  |  |  |  | 242 | 0 | 6 | 0 | 6145 | 0 | 223 | 0 |
| Patricia College St | AMB | Wyandotte |  |  |  |  | 427 | 1285 | 12 | 52 | 3480 | 3359 | 216 | 305 |
|  | E. of HC Road |  | 307 | 347 | 493 | 389 | 6521 | 5512 | 166 | 129 | 4 | 554 | 0 | 195 |
|  | W. of HC Road |  | 84 | 49 | 164 | 66 | 1627 | 887 | 0 | 0 | 424 | 47 | 0 | 0 |
| Girardot St | E. of HC Road |  | 54 | 84 | 87 | 47 | 1011 | 987 | 0 | 0 | 139 | 166 | 0 | 0 |
|  | W. of HC Road |  | 82 | 150 | 183 | 122 | 2199 | 2159 | 42 | 26 | 47 | 32 | 0 | 0 |
| Tecumseh Rd | E. of HC Road |  | 315 | 313 | 375 | 465 | 5350 | 6031 | 137 | 145 | 197 | 384 | 0 | 204 |
|  | W. of HC Road |  | 244 | 456 | 515 | 397 | 6315 | 6830 | 0 | 0 | 222 | 130 | 0 | 0 |
| Dorchester St | E. of HC Road |  | 76 | 46 | 86 | 52 | 1138 | 711 | 0 | 0 | 159 | 132 | 0 | 0 |
|  | W. of HC Road |  | 76 | 85 | 135 | 94 | 1761 | 1434 | 34 | 19 | 36 | 20 | 0 | 0 |
| Prince Rd/Totten St | E. of HC Road |  | 147 | 98 | 123 | 203 | 2124 | 2482 | 0 | 0 | 82 | 115 | 0 | 0 |
|  | W. of HC Road |  | 236 | 264 | 305 | 332 | 4646 | 4854 | 0 | 0 | 80 | 70 | 0 | 0 |
| Malden Rd | E. of HC Road |  | 102 | 63 | 99 | 91 | 1356 | 1050 | 0 | 0 | 275 | 268 | 0 | 0 |
|  | W. of HC Road |  | 398 | 413 | 465 | 409 | 6562 | 6018 | 370 | 319 | 497 | 39 | 192 | 563 |
| Industrial Rd | E. of HC Road |  | 280 | 157 | 216 | 282 | 3818 | 3553 | 49 | 60 | 153 | 168 | 8 | 22 |
|  | W. of HC Road |  | 292 | 104 | 175 | 297 | 3994 | 3037 | 163 | 204 | 0 | 0 | 0 | 208 |
| EC Row N. Ramp 1 | E. of HC Road (E-N/S Off Ramp \& S-W On Ramp) |  | 982 | 50 | 1065 | 125 | 15016 | 1374 | 317 | 1 | 1165 | 128 | 0 | 0 |
|  | W. of HC Road (N-W On Ramp) |  | 33 | n/a | 70 | n/a | 577 | n/a | 17 | n/a | 60 | n/a | 294 | n/a |
| EC Row S. Ramp T | E. of HC Road (S-E On Ramp) |  | n/a | 717 | n/a | 620 | n/a | 10896 | n/a | 103 | n/a | 0 | n/a | 0 |
|  | W. of HC Road (N-E On Ramp \& W-N/S Off Ramp) |  | 377 | 159 | 661 | 215 | 7836 | 2684 | 273 | 77 | 558 | 371 | 343 | 0 |
| Spring Gdn Rd/Lab | E. of HC Road |  | 317 | 157 | 152 | 190 | 3622 | 2934 | 0 | 0 | 112 | 92 | 0 | 0 |
|  | W. of HC Road |  | 116 | 239 | 347 | 283 | 3989 | 4325 | 0 | 0 | 5 | 5 | 0 | 0 |
| Lambton St/Grand | E. of HC Road |  | 377 | 198 | 240 | 254 | 4709 | 3704 | 0 | 0 | 239 | 215 | 0 | 0 |
|  | W. of HC Road |  | 91 | 203 | 148 | 271 | 1999 | 3820 | 38 | 52 | 40 | 52 | 0 | 0 |
| Pulford St | E. of HC Road |  | 193 | 127 | 96 | 102 | 1931 | 1693 | 0 | 0 | 279 | 299 | 0 | 0 |
| Todd Ln/Cabana Re | E. of HC Road |  | 492 | 459 | 720 | 511 | 9987 | 8450 | 0 | 0 | 86 | 63 | 0 | 0 |
|  | W. of HC Road |  | 464 | 562 | 1017 | 682 | 12838 | 10322 | 0 | 0 | 8 | 6 | 0 | 0 |
| Huron Church Line | W. of HC Road |  | 400 | 476 | 729 | 530 | 9382 | 7721 | 116 | 105 | 271 | 324 | 0 | 0 |
| St Clair College | E. of Talbot Road |  | 121 | 770 | 262 | 280 | 3107 | 9102 | 0 | 0 | 94 | 304 | 0 | 0 |
| Cousineau Dr | E. of Talbot Road |  | 426 | 260 | 551 | 363 | 6182 | 4244 | 0 | 0 | 1977 | 1172 | 0 | 0 |
|  | W. of Talbot Road |  | 470 | 495 | 893 | 540 | 11856 | 6974 | 0 | 113 | 0 | 1681 | 0 | 0 |
| Howard Ave | E. of Talbot Road |  | 383 | 499 | 553 | 730 | 8048 | 10052 | 135 | 198 | 3 | 5 | 0 | 0 |
|  | betweem Talbot Road and Hwy 401 SB On Ramp |  | 497 | 358 | 884 | 474 | 11210 | 7126 | 198 | 161 | 116 | 2 | 0 | 0 |
|  | W. of Hwy 401 SB On Ramp |  | 436 | 802 | 816 | 886 | 10221 | 14457 | 214 | 328 | 53 | 49 | 0 | 0 |
| E.C. Row Expressw | E. of Huron Church Rd |  | 2735 | 2215 | 2684 | 2722 | 38998 | 37195 | 822 | 790 | 3058 | 3567 | 575 | 1422 |
|  | At Malden Rd |  | 1836 | 1657 | 1814 | 2317 | 25945 | 28015 | 543 | 613 | 2046 | 4258 | 774 | 1379 |
|  | W. of Matchette |  | 1300 | 450 | 1217 | 580 | 20276 | 8617 | 475 | 412 | 0 | 0 | 0 | 0 |

[^7]| LOCATION | Table A-15 Contd. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | 24 Hour AADT |  |  |  |  |  |  |  |
|  | SECTION |  | 2025 |  |  |  | Local Cars |  | Local Trucks |  | International Cars |  | International Trucks |  |
|  |  |  | AM PEAK HOUR PM PEAK HOUR |  |  |  |  |  |  |  |  |  |  |  |
|  | FROM | TO | NB | SB | NB | SB | NB / WB | SB/EB | NB/WB | SB / EB | NB / WB | SB/EB | NB / WB | B / EB |
| HIGHWAY 401 Mainline |  |  | 2025 |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  | AM PEAK HOUR PM PEAK HOUR |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  | NB | SB | NB | SB | NB / WB | SB / EB | NB / WB | SB / EB | NB / WB | SB / EB | NB / WBS | B / EB |
| S. of Hwy 3 merge/split |  |  | 1360 | 1180 | 1560 | 1860 | 14411 | 13155 | 412 | 301 | 3441 | 3388 | 6373 | 10127 |
| N. of Howard Ave |  |  | 1241 | 1132 | 1425 | 1862 | 11756 | 11978 | 386 | 291 | 4032 | 4059 | 6240 | 10124 |
| At Grand Marais Rd |  |  | 1050 | 1338 | 1387 | 1900 | 10989 | 12714 | 355 | 341 | 4051 | 4902 | 5092 | 10731 |
| At Malden Rd |  |  | 645 | 767 | 423 | 1129 | 1940 | 4241 | 63 | 275 | 1854 | 3308 | 5655 | 10254 |
| To/From Canadian Plaza |  |  | 1050 | 560 | 520 | 1800 | 1 | 5 | 3 | 3 | 5096 | 9206 | 7555 | 12338 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| HIGHWAY 401 Ramps |  |  | 2025 |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  | AM PEAK HOUR PM PEAK HOUR |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  | NB / WB | SB / EB | NB / WB | SB / EB | NB / WB | SB / EB NB / WESB / EB |  |  |
| 401 NB Off Ramp |  |  | 616 | - | 773 | - | 9160 | 0 | 201 | 0 | 1258 | 0 | 596 | 0 |
| 401 NB On Ramp |  |  | 518 |  | 668 |  | 9231 | 0 | 177 | 0 | 0 | 0 | 0 | 0 |
| 401 SB Off Ramp |  |  | - | 645 |  | 603 | 0 | 9747 | 0 | 182 | 0 | 0 | 0 | 0 |
| 401 SB On Ramp |  |  |  | 649 | 1 | 622 | 0 | 7792 | 0 | 196 | 0 | 2039 | 0 | 561 |
| At St. Clair College |  |  |  |  |  |  | NB / WB | SB / EB | NB / WB | SB / EB | NB / WB | SB / EB | NB / WESB / EB |  |
| 401 NB Off Ramp |  |  | 545 |  | 330 |  | 6356 | 0 | 58 | 0 | 0 | 0 | 0 | 0 |
| 401 NB On Ramp |  |  | 354 | - | 292 | - | 4004 | 0 | 37 | 0 | 687 | 0 | 0 | 0 |
| 401 SB Off Ramp |  |  | - | 374 | $\square$ | 425 | 0 | 4960 | 0 | 58 | 0 | 1145 | 0 | 168 |
| 401 SB On Ramp |  |  | $\square$ | 168 | $\square$ | 387 | 0 | 4398 | 0 | 25 | 0 | 0 | 0 | 0 |
| At Huron Church Rd |  |  |  |  |  |  | NB / WB | SB / EB | NB / WB | SB / EB | NB / WB | SB / EB | NB / WE | S / EB |
| 401 NB Off Ramp |  |  | 405 | - | 964 |  | 8942 | 0 | 303 | 0 | 2075 | 0 | 0 | 0 |
| 401 SB On Ramp |  |  |  | 571 | $\square$ | 771 | 0 | 7837 | 0 | 204 | 0 | 1688 | 0 | 1124 |
| Malden Rd IC |  |  |  |  |  |  | NB / WB | SB / EB | NB / WB | SB/EB | NB / WB | SB / EB | NB / WB | B / EB |
| 401 On Ramp |  |  | 275 | - | 235 |  | 2723 | 0 | 424 | 0 | 454 | 0 | 161 | 0 |
| 401 Off Ramp |  |  |  | 275 | - | 360 | 0 | 3551 | 0 | 517 | 0 | 887 | 0 | 0 |
| EC Row Expressway IC |  |  |  |  |  |  | NB / WB | SB / EB | NB / WB | SB / EB | NB / WB | SB / EB | NB / WE | S / EB |
| 401 SB Off Ramp |  |  | $\square$ | 770 | $\square$ | 1276 | 0 | 10804 | 0 | 265 | 0 | 3886 | 0 | 1169 |
| Ojibway Pkwy IC |  |  |  |  |  |  | NB / WB | SB / EB | NB / WB | SB / EB | NB / WB | SB / EB | NB / WE | B / EB |
| 401 NB Off Ramp |  |  | 370 | - | 295 | , | 4663 | 0 | 487 | 0 | 0 | 0 | 0 | 0 |
| 401 NB On Ramp |  |  | 230 | , | 40 | - | 0 | 0 | 0 | 0 | 1330 | 0 | 355 | 0 |
| 401 SB Off Ramp |  |  |  | 20 |  | 140 | 0 | 142 | 0 | 13 | 0 | 930 | 0 | 75 |
| 401 SB On Ramp |  |  |  | 1272 |  | 1105 | 0 | 18366 | 0 | 543 | 0 | 0 | 0 | 0 |
| EC Row Expressway IC |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 401 NB On Ramp |  |  | 270 | , | 117 | - | 0 | 0 | 0 | 0 | 1569 | 0 | 954 | 0 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | FROM | TO | NB | SB | NB | SB |  |  |  |  |  |  |  |  |
|  | S. of Hwy 3 merge/split | Hwy 3/ 401 NB Off Ramp | 1360 |  | 1560 | - | 14411 |  | 412 | , | 3441 | - | 6373 |  |
|  | Hwy 3/ 401 NB Off Ran | Hwy 3/401 NB On Ramp | 723 | - | 757 | - | 4501 | - | 190 | , | 2873 |  | 5419 |  |
|  | Hwy 3/401 NB On Ram, | St. Clair/401 NB Off Ramp | 1241 | - | 1425 | - | 11756 | $\bigcirc$ | 386 | , | 4032 | - | 6240 |  |
|  | St. Clair/401 NB Off Ra | St. Clair/401 NB On Ramp | 696 | - | 1095 | - | 7332 | - | 307 |  | 3281 | , | 4709 |  |
|  | St. Clair/401 NB On Rar | HC Rd/401 NB Off Ramp | 1050 | - | 1387 | - | 10989 | - | 355 | , | 4051 | , | 5092 |  |
|  | HC Rd/401 NB Off Ram | Malden/401 NB On Ramp | 645 | - | 423 | , | 1940 | - | 63 | - | 1854 | , | 5655 |  |
| $>$ | Malden/401 NB On Ram | Ojibway/401 NB Off Ramp | 920 | - | 658 | - | 4663 | - | 487 | - | 2308 | , | 5817 |  |
| $\leftrightarrows$ | Ojibway Pkway/401 NB | Ojibway Pkway/401 NB On Ramp | 550 | - | 363 | - | 0 | , | 0 | , | 2308 | $\bigcirc$ | 5817 |  |
| 露 | Ojibway Pkway/401 NB | EC ROW to 401 NB On Ramp | 780 | - | 403 | $\cdots$ | 0 | , | 0 | $\bigcirc$ | 3638 | , | 6171 |  |
| $\sum$ | EC ROW to 401 NB On | Canadian Plaza | 1050 | - | 520 | $\bigcirc$ | 0 | $\bigcirc$ | 0 | $\square$ | 5096 | $\cdots$ | 7555 | $\square$ |
| $\stackrel{\rightharpoonup}{q}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\underset{\text { J }}{7}$ | Canadian Plaza | Ojibway/401 SB Off Ramp | $\square$ | 560 | - | 1800 | - | 5 | - | 3 | - | 9206 | - | 12338 |
| $\sum_{0}^{0}$ | Ojibway/401 SB Off Rar | Ojibway/401 SB On Ramp | $\square$ | 540 | $\square$ | 1660 | - | 4 | - | 3 | - | 8557 | , | 11597 |
|  | Ojibway/401 SB On Ran | 401 to EC ROW SB Off Ramp |  | 1812 | $\square$ | 2765 |  | 18022 | - | 637 |  | 8202 |  | 12526 |
|  | 401 to EC ROW SB Off | Malden/401 SB Off Ramp | $\square$ | 1042 | $\square$ | 1489 | - | 5683 | - | 368 | - | 4395 | , | 13722 |
|  | Malden/401 SB Off Ram | HC Rd/401 SB On Ramp |  | 767 | $\square$ | 1129 | , | 3344 | , | 95 | , | 3394 | $\square$ | 11730 |
|  | HC Rd/401 SB On Ramp | St Clair/401 SB Off Ramp |  | 1338 | 7 | 1900 | , | 12714 | $\square$ | 341 | - | 4902 | - | 10731 |
|  | St Clair/401 SB Off Ran | St Clair/401 SB On Ramp | 7 | 964 | , | 1475 | - | 8444 | - | 266 | , | 3737 | - | 9521 |
|  | St Clair/401 SB On Ram | Hwy 3/401 SB Off Ramp |  | 1132 | $\square$ | 1862 | - | 11978 | - | 291 | , | 4059 | - | 10124 |
|  | Hwy 3/401 SB Off Ram, | Hwy 3/401 SB On Ramp | 7 | 487 | - | 1259 | $\bigcirc$ | 6636 | - | 156 | , | 2531 | , | 5970 |
|  | Hwy 3/401 SB On Ramp | S. of Hwy 3 merge/split | $\square$ | 1180 | $\square$ | 1860 | - | 13155 | - | 301 | - | 3388 | - | 10127 |
|  | FROM | TO | NB | SB | NB | SB |  |  |  |  |  |  |  |  |
|  | Chappus | 401 S. Ramp | 600 | 580 | 499 | 660 | 7519 | 9673 | 340 | 416 | 762 | 676 | 0 | 0 |
| Malden | 401 S. Ramp | 401 N. Ramp | 645 | 350 | 589 | 390 | 8505 | 5779 | 382 | 247 | 841 | 401 | 0 | 0 |
|  | N. of 401 N. Ramp |  | 415 | 395 | 409 | 445 | 5713 | 6556 | 255 | 281 | 554 | 456 | 0 | 0 |
|  | Chappus | EC Row S. Ramp | 508 | 395 | 504 | 560 | 8771 | 7728 | 0 | 0 | 144 | 269 | 0 | 0 |
| Matchette | EC Row S. Ramp | EC Row N. Ramp | 158 | 483 | 149 | 668 | 2577 | 9359 | 0 | 0 | 129 | 265 | 0 | 0 |
|  | EC Row N. Ramp | Carmichael | 382 | 171 | 331 | 253 | 6140 | 3359 | 0 | 0 | 159 | 207 | 0 | 0 |
| Montgomery | Surrey | Talbot | 4 | 15 | 8 | 7 | 97 | 188 | 1 | 2 | 2 | 4 | 0 | 3 |
| Surrey | Montgomery | Talbot | 5 | 5 | 4 | 22 | 70 | 217 | 1 | 3 | 2 | 8 | 0 | 1 |
| Grosvenor | Montgomery | Talbot | 11 | 7 | 5 | 19 | 122 | 211 | 2 | 3 | 4 | 8 | 0 | 1 |


| LOCATION | SECTION |  | 2035 |  |  |  | 24 Hour AADT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Local Cars |  | Local Trucks |  | International Cars |  | International Trucks |  |
|  |  |  | AM PEAK HOUR | PM PEAK HOUR |  |  |  |  |  |
|  | FROM | TO |  |  | NB | SB | NB | SB | NB/WB | SB/EB | NB/WB | SB / EB | NB/WB | SB/EB | NB/WB | SB/EB |
| HC Road | Riverside | University |  |  |  |  |  |  | 6678 | 5604 | 199 | 95 | 3 | 1 | 0 | 81 |
|  | University | Wyandotte |  |  |  |  | 2832 | 3759 | 92 | 125 | 67 | 241 | 63 | 81 |
|  | Wyandotte | AMB Off Ramp |  |  |  |  | 2061 | 3152 | 0 | 0 | 46 | 176 | 0 | 0 |
|  | AMB Off Ramp | College |  |  |  |  | 7911 | 6474 | 237 | 106 | 7717 | 1 | 3852 | 0 |
|  | College St | Girardot St | 1846 | 897 | 1476 | 1855 | 17762 | 16404 | 558 | 517 | 6849 | 5320 | 293 | 3861 |
|  | Girardot St | Tecumseh Rd | 1774 | 912 | 1384 | 1744 | 17897 | 18482 | 693 | 654 | 6255 | 4781 | 275 | 3762 |
|  | Tecumseh Rd | Dorchester St | 1836 | 1203 | 1740 | 1906 | 21270 | 22372 | 851 | 812 | 5982 | 3977 | 278 | 3334 |
|  | Dorchester St | Prince Rd/Totten St | 1913 | 1244 | 1727 | 1974 | 22277 | 24172 | 773 | 737 | 5498 | 3676 | 247 | 3066 |
|  | Prince Rd/Totten St | Malden Rd | 2012 | 1421 | 1956 | 2197 | 24702 | 27332 | 862 | 837 | 5362 | 3272 | 280 | 2766 |
|  | Malden Rd | Industrial Rd | 1652 | 1097 | 1560 | 1846 | 19459 | 22099 | 660 | 642 | 5334 | 3363 | 11 | 2746 |
|  | Industrial Rd | EC Row N. Ramp Terminal | 1755 | 1135 | 1640 | 2070 | 21452 | 24453 | 704 | 695 | 4898 | 3479 | 0 | 2656 |
|  | EC Row N. Ramp Terminal | EC Row S. Ramp Terminal | 1240 | 1582 | 1303 | 2598 | 15992 | 34770 | 489 | 774 | 4403 | 3352 | 0 | 2408 |
|  | EC Row S. Ramp Terminal | Spring Gdn Rd/Labelle St | 1897 | 1267 | 1891 | 1974 | 26279 | 26779 | 547 | 491 | 2836 | 2906 | 0 | 1848 |
|  | Spring Gdn Rd/Labelle St | Lambton St/Grand Marais Rd | 1606 | 627 | 1852 | 990 | 25464 | 12835 | 394 | 204 | 1862 | 889 | 0 | 0 |
|  | Lambton St/Grand Marais Rd | Pulford St | 963 | 752 | 751 | 1021 | 12242 | 14344 | 203 | 177 | 994 | 825 | 0 | 0 |
|  | Pulford St | Todd Ln/Cabana Rd | 975 | 831 | 765 | 1028 | 12527 | 15281 | 224 | 193 | 930 | 687 | 0 | 0 |
|  | Todd Ln/Cabana Rd | Huron Church Line | 797 | 767 | 798 | 1027 | 12297 | 15199 | 135 | 111 | 458 | 333 | 0 | 0 |
| Talbot Rd | Huron Church Line | St Clair College | 430 | 821 | 432 | 872 | 7028 | 14513 | 89 | 110 | 3 | 272 | 0 | 0 |
|  | St Clair College | Cousineau Dr | 1198 | 431 | 845 | 516 | 13092 | 6326 | 126 | 78 | 2294 | 1478 | 0 | 239 |
|  | Cousineau Dr | Howard Ave | 469 | 339 | 493 | 267 | 6884 | 4573 | 100 | 75 | 640 | 479 | 0 | 233 |
|  | S. of Howard Ave |  | 789 | 898 | 1114 | 793 | 15541 | 14634 | 324 | 305 | 0 | 0 | 0 | 0 |
| Ojibway Pwy | EC Row Expressway | GN Booth Dr | 734 | 470 | 700 | 830 | 11678 | 10724 | 149 | 131 | 0 | 19 | 0 | 648 |
|  | GN Booth Dr | Sandwich St | 720 | 468 | 703 | 803 | 11595 | 10483 | 148 | 128 | 0 | 19 | 0 | 645 |
|  | Sandwich St | Prospect Ave | 678 | 419 | 649 | 763 | 10793 | 10162 | 76 | 70 | 51 | 48 | 0 | 0 |
|  | N. of Prospect Ave |  | 671 | 406 | 648 | 750 | 10731 | 9936 | 75 | 68 | 51 | 47 | 0 | 0 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| CROSSING ROADS |  |  | WB | EB | WB | EB | NB/ WB | SB / EB | NB/ WB | SB / EB | NB/WB | SB / EB | NB/WB | SB / EB |
| Wyandotte | W of HuronChurch |  |  |  |  |  | 4596 | 4453 | 0 | 0 | 381 | 447 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2776 | 4331 | 17 | 153 | 770 | 948 | 58 | 0 |
| University | W of HuronChurch |  |  |  |  |  | 1513 | 1313 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 2206 | 2073 | 125 | 92 | 65 | 21 | 81 | 63 |
| Riverside | W of HuronChurch |  |  |  |  |  | 3708 | 4018 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | E of HuronChurch |  |  |  |  |  | 7079 | 5910 | 0 | 0 | 177 | 46 | 0 | 0 |
| AMB Off Ramp | E of HuronChurch |  |  |  |  |  | 0 | 932 | 0 | 43 | 0 | 7713 | 81 | 3852 |
| AMB On Ramp | E of HuronChurch |  |  |  |  |  | 221 | 0 | 6 | 0 | 6520 | 0 | 273 | 0 |
| Patricia | AMB | Wyandotte |  |  |  |  | 387 | 930 | 12 | 41 | 3574 | 3467 | 233 | 394 |
| College St | E. of HC Road |  | 322 | 352 | 496 | 399 | 6657 | 5523 | 173 | 130 | 4 | 590 | 0 | 289 |
|  | W. of HC Road |  | 90 | 47 | 185 | 77 | 1737 | 984 | 0 | 0 | 522 | 38 | 0 | 0 |
| Girardot St | E. of HC Road |  | 59 | 84 | 84 | 48 | 1017 | 992 | 0 | 0 | 142 | 168 | 0 | 0 |
|  | W. of HC Road |  | 85 | 147 | 180 | 121 | 2202 | 2125 | 42 | 27 | 47 | 33 | 0 | 0 |
| Tecumseh Rd | E. of HC Road |  | 329 | 325 | 416 | 505 | 5805 | 6274 | 141 | 146 | 204 | 457 | 0 | 345 |
|  | W. of HC Road |  | 253 | 479 | 541 | 414 | 6419 | 7211 | 0 | 0 | 381 | 110 | 0 | 0 |
| Dorchester St | E. of HC Road |  | 76 | 46 | 86 | 52 | 1136 | 708 | 0 | 0 | 160 | 134 | 0 | 0 |
|  | W. of HC Road |  | 78 | 85 | 135 | 96 | 1779 | 1449 | 34 | 20 | 36 | 21 | 0 | 0 |
| Prince $\mathrm{Rd} /$ /otten St | E. of HC Road |  | 153 | 97 | 129 | 204 | 2220 | 2485 | 0 | 0 | 83 | 112 | 0 | 0 |
|  | W. of HC Road |  | 240 | 269 | 316 | 331 | 4780 | 4880 | 0 | 0 | 76 | 73 | 0 | 0 |
| Malden Rd | E. of HC Road |  | 117 | 70 | 114 | 106 | 1550 | 964 | 0 | 0 | 318 | 498 | 0 | 0 |
|  | W. of HC Road |  | 429 | 431 | 479 | 426 | 6891 | 6225 | 380 | 323 | 529 | 14 | 232 | 699 |
| Industrial Rd | E. of HC Road |  | 307 | 166 | 227 | 295 | 3646 | 3711 | 45 | 59 | 695 | 195 | 8 | 29 |
|  | W. of HC Road |  | 309 | 113 | 189 | 304 | 4258 | 3117 | 174 | 210 | 0 | 0 | 0 | 267 |
| EC Row N. Ramp Terminal | E. of HC Road (E-N/S Off Ram | \& S-W On Ramp) | 1049 | 52 | 1104 | 140 | 15875 | 1474 | 340 | 3 | 1188 | 171 | 0 | 0 |
|  | W. of HC Road (N-W On Ramp) |  | 35 | n/a | 80 | n/a | 587 | n/a | 15 | n/a | 66 | n/a | 403 | n/a |
| EC Row S. Ramp Terminal | E. of HC Road (S-E On Ramp) |  | n/a | 758 | n/a | 680 | n/a | 11723 | n/a | 118 | n/a | 0 | n/a | 0 |
|  | W. of HC Road (N-E On Ramp | W-N/S Off Ramp) | 390 | 176 | 698 | 221 | 8199 | 2878 | 294 | 76 | 479 | 362 | 535 | 0 |
| Spring Gdn Rd/Labelle St | E. of HC Road |  | 283 | 169 | 161 | 191 | 3170 | 3133 | 0 | 0 | 265 | 27 | 0 | 0 |
|  | W. of HC Road |  | 126 | 214 | 367 | 296 | 4250 | 4243 | 0 | 0 | 5 | 4 | 0 | 0 |
| Lambton St/Grand Marais Rd | E. of HC Road |  | 425 | 206 | 254 | 263 | 5172 | 3873 | 0 | 0 | 246 | 200 | 0 | 0 |
|  | W. of HC Road |  | 100 | 202 | 144 | 266 | 2047 | 3766 | 39 | 53 | 40 | 53 | 0 | 0 |
| Pulford St | E. of HC Road |  | 198 | 132 | 94 | 102 | 1943 | 1730 | 0 | 0 | 284 | 308 | 0 | 0 |
| Todd Ln/Cabana Rd | E. of HC Road |  | 518 | 512 | 797 | 538 | 10844 | 9168 | 0 | 0 | 97 | 60 | 0 | 0 |
|  | W. of HC Road |  | 478 | 586 | 1051 | 720 | 13248 | 10838 | 0 | 0 | 13 | 7 | 0 | 0 |
| Huron Church Line | W. of HC Road |  | 445 | 571 | 836 | 617 | 10658 | 9046 | 131 | 124 | 293 | 420 | 0 | 0 |
| St Clair College | E. of Talbot Road |  | 121 | 769 | 268 | 269 | 3164 | 9013 | 0 | 0 | 91 | 293 | 0 | 0 |
| Cousineau Dr | E. of Talbot Road |  | 449 | 278 | 611 | 365 | 6652 | 4360 | 0 | 0 | 2223 | 1244 | 0 | 0 |
|  | W. of Talbot Road |  | 517 | 538 | 987 | 582 | 13081 | 7324 | 0 | 120 | 0 | 2039 | 0 | 0 |
| Howard Ave | E. of Talbot Road |  | 420 | 530 | 598 | 801 | 8753 | 10898 | 149 | 213 | 3 | 6 | 0 | 0 |
|  | betweem Talbot Road and Hwy | 01 SB On Ramp | 542 | 375 | 947 | 526 | 12085 | 7704 | 216 | 178 | 121 | 2 | 0 | 0 |
|  | W. of Hwy 401 SB On Ramp |  | 476 | 867 | 869 | 985 | 10974 | 15844 | 234 | 363 | 54 | 57 | 0 | 0 |
| E.C. Row Expressway | E. of Huron Church Rd |  | 3288 | 2489 | 3014 | 3034 | 44794 | 41237 | 944 | 890 | 3783 | 4209 | 748 | 1758 |
|  | At Malden Rd |  | 2326 | 1907 | 2130 | 2575 | 30926 | 31214 | 645 | 689 | 2820 | 5162 | 1041 | 1553 |
|  | W. of Matchette |  | 1676 | 480 | 1453 | 640 | 25148 | 9398 | 594 | 413 | 0 | 0 | 0 | 0 |

[^8]
 RESULTS

# MEMORANDUM 



Hyper Anemo*

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To: Ms. Abby Salb, SENES
From: Tom Darlington
Date: December 8, 2005
Subject: Emission Rates for Windsor/Detroit Crossing Project

This memo details the models, inputs, and procedures used to estimate on-road mobile source emission rates for various vehicle types for the Windsor/Detroit Crossing project.

This memo focuses on the emission rates from all vehicle types except idle emissions from heavy-duty diesel vehicles. The idle emissions from heavy-duty diesel vehicles are described in a separate AIR memo. [1]

This memo is divided into the following sections:

- Background
- Models
- Seasons and ambient temperatures
- Vehicle speeds
- Fuel inputs
- Results


## Background

Detroit and Windsor are studying the possibility of adding a second Detroit River facility to augment the current Ambassador Bridge and tunnel. Such a crossing would change emissions of vehicles on both sides of the crossing. For example, heavy-duty diesel trucks may experience reduced idle times if the crossing were added. Light duty vehicles may also experience reduced idle times and somewhat higher average speeds in the vicinity of the crossings. At the same time, cross-border traffic could increase, as the time it takes to cross the border is reduced.

A key part of the study is to estimate the impact of a new crossing on traffic flow on both sides, and the resultant impact on vehicle emissions. To estimate these emission impacts requires detailed information about emission rates at idle, and at various speeds, for all the different vehicle types, and also detailed projections of traffic flow, and the projected impact of the crossing on traffic flow in the vicinity of the crossing.

The purpose of this memo is to describe the methods used to estimate emissions on both the U.S. and Canadian side of the crossing. SENES contracted with AIR to estimate vehicle emissions for all of the various vehicle types, for both sides of the crossing. AIR assisted EPA in the development of the MOBILE6 model, and also developed the MOBILE6.2C model for Environment Canada. These models estimate emissions for a number of different vehicle types. The emissions are estimated in units of $\mathrm{g} / \mathrm{mi}$ for vehicles not at idle, and in units of $\mathrm{g} / \mathrm{hr}$ for vehicles at idle.

## Models Used

AIR used EPA's MOBILE6.2 model for the Detroit side, and used Environment Canada's M6C25PPM model for Windsor. The M6C25PPM model is a Canadian version of the MOBILE6 model that incorporates fuel changes and many other changes that are specific to the Canadian fleet. Both models estimate all of the pollutants needed in this evaluation, however, AIR utilized more up-to-date procedures for estimating emissions from idling heavy-duty diesel trucks.

The following pollutants were estimated:

- VOC
- CO
- NOx
- SO 2
- PM2.5
- CO 2
- Methane
- 1,3 butadiene
- Acrolein
- Formaldehyde
- Acetaldehyde
- Benzene

The above pollutants were estimated for a base year, 2004, and two projection years, 2013 and 2023.

## Seasons and Ambient Temperatures

Emissions are estimated for the fours seasons. Average minimum and maximum temperatures for these seasons were determined for both locations using 30 years of data from the National Weather Service for the US, and from Environment Canada for Canada. The ambient temperatures for the two locations are shown in Table 1 below.

| Table 1. Average Minimum and Average Maximum Temperatures (F) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Season | Detroit |  | Windsor |  |  |
| Winter | 22.8 | 35.6 | 19.7 | 32.4 |  |
| Spring | 38.8 | 57.7 | 37.3 | 55.4 |  |
| Summer | 67.1 | 88.9 | 60.4 | 79.9 |  |
| Autumn | 43.4 | 60.9 | 46.7 | 60.2 |  |

## Vehicle Speeds

Vehicle speed inputs were obtained from SENES. Emissions were estimated for the following speeds: Idle ( 2.5 mph ), 15.5, 31.1, 46.6, and 62.1 mph . The same speeds were used for both sides of the border.

## Fuel Inputs

Both models used default gasoline and diesel fuel sulfur levels for Canada and the U.S. Detailed gasoline inputs are also needed to compute toxics emission rates. Ontario fuel property data was obtained from Natural Resources Canada. [2] Data for Detroit was obtained from The Alliance of Automobile Manufacturers. [3] Fuel characteristics are shown in Table 2.

| Table 3. Non-Sulphur Gasoline Characteristics |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| City | Season | RVP <br> (psi) | E200 <br> $(\%)$ | E300 <br> $(\%)$ | Arom. <br> $(\%)$ | Olef. <br> $(\%)$ | Benzene <br> $(\%)$ | \% with <br> ETOH | ETOH <br> Concen. |  |  |
| Detroit | Winter | 14.4 | 53.8 | 82.7 | 26.8 | 6.9 | 1.7 | $25 \%$ | $9.75 \%$ |  |  |
|  | Spring | 11.0 | 47.7 | 81.2 | 29.4 | 8.5 | 1.6 | $25 \%$ | $9.75 \%$ |  |  |
|  | Summer | 7.6 | 41.6 | 79.6 | 32.0 | 10.0 | 1.5 | $25 \%$ | $9.75 \%$ |  |  |
|  | Fall | 11.0 | 47.7 | 81.2 | 29.4 | 8.5 | 1.6 | $25 \%$ | $9.75 \%$ |  |  |
| Windsor | Winter | 14.6 | 53.9 | 84.4 | 25.1 | 9.0 | 0.73 | $100 \%$ | $1.92 \%$ |  |  |
|  | Spring | 12.1 | 50.9 | 83.4 | 26.9 | 9.3 | 0.73 | $100 \%$ | $1.92 \%$ |  |  |
|  | Summer | 9.7 | 47.9 | 82.4 | 28.8 | 9.7 | 0.73 | $100 \%$ | $1.92 \%$ |  |  |
|  | Fall | 12.1 | 50.9 | 83.4 | 26.9 | 9.3 | 0.73 | $100 \%$ | $1.92 \%$ |  |  |

Gasoline and diesel sulphur levels that are contained in both models for 2003, 2013, and 2023 are shown in Table 4.

| Table 4. Sulphur Levels |  |  |  |
| :---: | :---: | :---: | :---: |
| Fuel | Year | Sulphur Level (ppm) - <br> Windsor | Sulphur Level (ppm) - <br> Detroit |
| Gasoline | 2004 | 52 | $170-180$ ppm ,depending <br> on season |
|  | 2013 | 25 | 30 |
|  | 2023 | 25 | 30 |
| Diesel | 2004 | 320 | 365 |
|  | 2013 | 15 | 11 |
|  | 2023 | 15 | 11 |

## Technologies and Emission Standards

Both models used in this analysis include the effects of all currently adopted regulatory programs for light duty vehicles and light duty trucks, as follows:

## Light Duty Vehicles

- National LEV program starting in 2001
- Onboard vapor recovery requirements for all gasoline cars, trucks, and SUVs
- Onboard diagnostic requirements for all vehicles
- Tier 2 exhaust emission standards
- Tier 2 evaporative emission standards

Technologies which are being used to meet the Tier 2 exhaust emission standards are closer air/fuel ratio control, increased previous metal loadings on catalysts, closer-coupled catalysts, reduced cold-start emissions, and dual oxygen sensors. Technologies being used to meet the Tier 2 evaporative standards are larger and redesigned charcoal canisters, very low permeation hoses and fuel tanks, and other technologies designed to reduced vapor generation from the fuel tanks and lines during engine operation.

## Heavy-Duty Vehicles

- 2004 HC+NOx standards
- 2007-2010 HC, NOx and PM standards
- 2010 NOx standards

The 2007-2010 heavy-duty standards assume the use of catalyzed PM traps to meet the $0.01 \mathrm{~g} / \mathrm{bhp}-\mathrm{hr}$ PM standard, and either engine controls like aggressive EGR, or aftertreatment (or both) needed to obtain a $50 \%$ NOx reduction. The 2010 heavy-duty NOx standards are a $90 \%$ reduction from 2006 NOx, and currently it is thought that this can only be met with aftertreament and aggressive EGR. Currently the aftertreatment choices to meet the 2010 NOx standard of 0.2 $\mathrm{g} / \mathrm{bhp}-\mathrm{hr}$ is either selective catalytic reduction (SCR), or a NOx adsorber.

EPA is planning to propose a mobile source toxics rule to apply to future light duty gasoline vehicles and trucks. That rule will probably reduce toxics from motor vehicles further, but the rule is not reflected in these emission rates because it has not been either proposed or adopted.

## Heavy-Duty Fleet Turnover Comparison

The figure below shows a comparison of registration fractions versus age for both Detroit and Windsor. The Detroit fleet appears to be somewhat newer with the highest registration fractions in the 1-5 year old age group, but there also is a significantly higher fraction in the 25+ year old category for Detroit. Windsor appears to have a somewhat older fleet on average, in that the highest registrations fractions are for vehicles that are 6-9 years of age.


## Results

All results are shown in spreadsheet format in two different files, "Detroit.xls", and "Windsor.xls".

## References

[1] "Idle Emission Rates for Diesel Trucks", Memo from Tom Darlington at AIR to Dan Hrebenyk at SENES, November 9, 2005.
[2] Natural Resources Canada
[3] Alliance of Automobile Manufacturers Fuel Survey for Detroit for 2003.

# MEMORANDUM 



Hyper Anemo"

To: Dan Hrebenyk, SENES
From: Tom Darlington
Date: November 9, 2005
Subject: Idle Emission Rates for Diesel Trucks

This memo develops heavy duty diesel emission idle and "creep" emission rates for use in Vancouver.

Method
We are unsure of the duty cycle of heavy-duty trucks which are waiting in line at them loading terminal. Therefore, we have developed two sets of emissions rates - one is an idle emission rate, if the duty cycle is almost all idle, and the second estimate is based on a "creep" cycle, which was developed by the California Air Resources Board and West Virginia University.

The idle emission rates we recommend using in Vancouver come from a recent ARB staff report on requirements to reduce idling emissions from new and in-use trucks. The report lists idle emissions by model year for heavy-duty diesel trucks that are weighted by the fraction of time spent at low idle and high idle. The emission rates are also weighted by summer and winter fractions.

We obtained the separate winter and summer idle emission emission rates, at both low and high idle. For Vancouver, we have developed separate summer and winter emission rates, but we have used the ARB low and high idle fractions in each season. Idle emission rates were developed for three years: 2003, 2011, and 2020. Idle emission rates were developed for NOx, PM10, VOC, CO, and $\mathrm{CO}_{2}$.

The emission rates based on the creep cycle have been developed from raw data obtained from the Coordinating Research Council's E55/57 testing program (the idle emission rates also ultimately come from this testing program). The creep cycle is a very low average speed cycle, where speed is varied between 0 and 8 mph and 0 and 3 mph , with an idle period in between.

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## ARB's Idle Emission Rates

In the recent idle emissions staff report, ARB lists the idle emissions for heavy-duty diesel trucks in g/hr. [1] These emission rates are shown in Table 1.

| Table 1. HDDT Idle Emissions (grams/hour) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Calendar <br> Year | Model Year | NOx | ROG | PM | $\mathrm{CO}_{2}$ |
| 2010 | Pre-1991 | 39.8 | 20.2 | 5.3 | 6228 |
|  | $1991-2006$ | 115.3 | 9.4 | 1.9 | 6228 |
|  | $2007+$ | 115.3 | 8.3 | 0.16 | 6228 |
| 2020 | Pre-1991 | 39.8 | 20.1 | 5.2 | 6228 |
|  | $1991-2006$ | 115.3 | 8.9 | 1.8 | 6228 |
|  | $2007+$ | 115.3 | 8.3 | 0.16 | 6228 |

The above emission factors were developed by the ARB from recent tests conducted by West Virginia University as a part of the Coordinating Research Councils' E55/E57 testing program. [2] The above numbers include typical accessory loads for both summer and winter (summer is weighted 7/12 and winter is weighted 5/12), and also include both low and high idle operation. The low/high idle weighting factors are $61 \%$ low idle, $39 \%$ high idle.

The PM emission rates are much lower for 2007 and later trucks, due to fact that 2007 and later trucks are subject to much lower PM standards ( $0.01 \mathrm{~g} / \mathrm{bhp}-\mathrm{hr}$ ). While the NOx standards are also lower in 2007 and 2010 ( $1.2 \mathrm{~g} / \mathrm{bhp}-\mathrm{hr}$ and $0.2 \mathrm{~g} / \mathrm{bhp}-\mathrm{hr}$, respectively), ARB does not expect this technology to reduce idle NOx emissions, because idle temperatures are much lower than when the engine is under load, and the expected NOx emission control technology is expected to be less efficient at lower temperatures than at high temperatures. [1]

Idle Emissions for Vancouver
Since the climate is much different in Vancouver than in California, we recommend the use of separate winter and summer emission rates in Vancouver. The emission rates still need to utilize the ARB fractions of high and low idle operation.

We obtained the separate summer and winter high and low idle emission rates and high idle correction factors from the ARB, and these are shown in Attachments 1 and 2. We then weighted the low idle baseline with the summer high idle and winter high idle emission rates. The results are shown in Tables 2 and 3.

| Table 2. Summer Vancouver Idle Emission Rates (g/hr) for HDDTs |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Model Year | PM | NOx | CO | HC | CO2 |
| $2007+$ | 0.13 | 119.0 | 33.7 | 7.8 | 6594 |
| $2004-2006$ | 1.35 | 119.0 | 33.7 | 7.8 | 6594 |
| $1998-2003$ | 1.35 | 119.0 | 33.7 | 7.8 | 6594 |
| $1994-1997$ | 1.80 | 119.0 | 37.4 | 9.7 | 6594 |
| $1991-1993$ | 2.38 | 119.0 | 41.6 | 12.0 | 6594 |
| 1990 | 3.17 | 119.0 | 46.2 | 14.9 | 6594 |
| $1987-1989$ | 3.17 | 41.1 | 46.2 | 14.9 | 6594 |
| $1984-1986$ | 4.21 | 41.1 | 51.2 | 18.5 | 6594 |
| $1980-1983$ | 5.60 | 41.1 | 56.9 | 22.9 | 6594 |
| $1977-1979$ | 7.42 | 41.1 | 63.2 | 28.4 | 6594 |
| $1975-1976$ | 9.08 | 41.1 | 68.1 | 33.0 | 6594 |
| Pre-1975 | 10.68 | 41.1 | 72.3 | 37.4 | 6594 |

Table 3. Winter Vancouver Idle Emission Rates (g/hr) for HDDTs

| Model Year | PM | NOx | CO | HC | CO2 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $2007+$ | 0.19 | 110.2 | 63.9 | 9.0 | 5714 |
| $2004-2006$ | 1.95 | 110.2 | 63.9 | 9.0 | 5714 |
| $1998-2003$ | 1.95 | 110.2 | 63.9 | 9.0 | 5714 |
| $1994-1997$ | 2.59 | 110.2 | 70.9 | 11.1 | 5714 |
| $1991-1993$ | 3.44 | 110.2 | 78.8 | 13.8 | 5714 |
| 1990 | 4.58 | 110.2 | 87.5 | 17.1 | 5714 |
| $1987-1989$ | 4.58 | 38.0 | 87.5 | 17.1 | 5714 |
| $1984-1986$ | 6.07 | 38.0 | 97.2 | 21.2 | 5714 |
| $1980-1983$ | 8.08 | 38.0 | 107.9 | 26.3 | 5714 |
| $1977-1979$ | 10.72 | 38.0 | 119.8 | 32.5 | 5714 |
| $1975-1976$ | 13.11 | 38.0 | 129.1 | 37.9 | 5714 |
| Pre-1975 | 15.42 | 38.0 | 137.1 | 42.8 | 5714 |

As shown in Table 2 and 3, the winter PM, CO, and HC emission rates are higher than the summer emission rates, and the NOx and $\mathrm{CO}_{2}$ emission rates are lower.

Heavy-duty truck registration distributions were obtained for British Columbia from modeling we have done for Environment Canada. The registration distributions are shown in Attachment 3. These registration distributions were used with the idle emission rates in Table 2 and 3 to develop fleet idle emission rates for three years: 2003, 2011, and 2020. The final fleet idle emission rates for summer and winter for 2003, 2011 and 2020 are shown in Table 4.

| Table 4. HDDT Fleet Idle Emission Rates (g/hr) for HDDTs |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Season | PM | NOx | CO | HC | $\mathrm{CO}_{2}$ |
| 2003 | Summer | 2.26 | 110 | 39.4 | 11.3 | 6594 |
|  | Winter | 3.26 | 102 | 74.7 | 12.9 | 5714 |
| 2011 | Summer | 1.26 | 111 | 36.1 | 9.2 | 6594 |
|  | Winter | 1.82 | 110 | 68.5 | 10.5 | 5714 |
| 2020 | Summer | 0.52 | 119 | 34.0 | 8.0 | 6594 |
|  | Winter | 0.75 | 110 | 64.5 | 9.2 | 5714 |

"Creep" Emission Rates
The CRC testing referenced earlier also included a "Creep" cycle. This cycle was 0.13 miles long, with an average speed of 1.6 mph . The driving cycle is shown in Attachment 4. The cycle is intended to develop emissions for situations in which trucks wait in lines for long periods of time with idle and very slow speed operation, like at borders and toll collections, etc. Trucks were tested with normal accessory loads (compressor fan and alternator, but not a/c or heater).

AIR estimated average creep emissions in g/mi for pre-1991 and 1991 and later trucks, as shown in Table 5. These were estimated in both g/mi (first two columns), and in $\mathrm{g} / \mathrm{hr}$ (second two columns).

| Table 5. "Creep" Emission Rates |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | $\mathrm{g} / \mathrm{mi}$ |  | $\mathrm{g} / \mathrm{hr}$ |  |
| Pollutant | Pre-1991 | $1991+$ | Pre-1991 | $1991+$ |
| NOx | 38.6 | 71.7 | 62.7 | 116.1 |
| HC | 15.5 | 9.2 | 25.1 | 14.9 |
| PM | 7.2 | 3.5 | 11.7 | 5.7 |
| CO | 30.9 | 20.2 | 50.0 | 32.7 |

For NOx, the $\mathrm{g} / \mathrm{hr}$ emission rates in Table 5 are similar to the NOx and CO emission rates in Table 4. However the creep cycle HC and PM rates appear to be higher than the rates in Table 4. This is due to the acceleration periods from idle in this cycle (see Attachment 4). Starting in model year 2007, however, PM emission rates must be reduced by $90 \%$. Therefore, we propose the use of a $0.57 \mathrm{~g} / \mathrm{hr}$ emission rate for 2007 and later heavy-duty trucks. While NOx emissions may also be reduced because of lower NOx standards, for this analysis we will assume they remain the same as $1991+$ creep emission rates. We also propose the use of a $12 \%$ reduction in VOC emissions, similar to the ARB in Table 1 ( $13.1 \mathrm{~g} / \mathrm{hr}$ ).

Using estimates of HDDV VMT fractions in Attachment 3, the 2003, 2011 and 2020 fleet "creep" emissions are shown in Table 6.

| Table 6. HDDT Fleet Idle Emission Rates (g/hr) for HDDTs |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Year | PM | NOx | CO | HC |
| 2003 | 6.94 | 105 | 36 | 17 |
| 2011 | 5.04 | 116 | 36 | 16 |
| 2020 | 2.19 | 116 | 33 | 14 |

The HC and PM emission rates in Table 6 are somewhat higher than those in Table 5. These may be the most realistic emission rates to use for Vancouver, if the duty cycle includes idle punctuated by slow movement.

## $\mathrm{SO}_{2}$ Emission Rates

$\mathrm{SO}_{2}$ emission rates can be estimated from the very low speed fuel consumption estimates from the creep cycle data (fuel consumption is not available from the idle emission tests). Idle $\mathrm{SO}_{2}$ emission rates in $\mathrm{g} / \mathrm{hr}$ can be estimated with the following expression:
$\mathrm{SO}_{2}(\mathrm{~g} / \mathrm{hr})=($ cycle miles $/ \mathrm{mpg}) * 4.44 \mathrm{~L} / \mathrm{gal} * 850 \mathrm{~g} / \mathrm{L} *$ Sulphur ppm * $(64 / 32) /\left(\mathrm{hr} * 10^{6}\right)$
Where:
Cycle miles $=0.13$ miles
$\mathrm{Mpg}=$ average of 2.32 mpg
$850=$ typical density of diesel fuel
sulphur ppm = 365 ppm in 2003, 15 ppm in other years
$64 / 32=$ molecular weight ratio of $\mathrm{SO}_{2}$ to S
$\mathrm{hr}=$ cycle time in hours, or 0.08 hrs
Using the above expression, the $\mathrm{SO}_{2}$ emission rates in $\mathrm{g} / \mathrm{hr}$ are shown in Table 6 below.

| Table 6. SO $\mathbf{2}_{2}$ Emission Rates (g/hr) |  |  |
| :---: | :---: | :---: |
| Year | Sulphur in Diesel fuel (ppm) | $\mathrm{SO}_{2}$ Emission Rate (g/hr) |
| 2003 | 365 | 1.93 |
| 2011 | 15 | 0.08 |
| 2020 | 15 | 0.08 |

## EPA Guidance on PM and NOx

Finally, we note EPA’s 2002 guidance recommends a NOx emission rate of $135 \mathrm{~g} / \mathrm{hr}$, and a PM emission rates that vary by model year from $3.68 \mathrm{~g} / \mathrm{hr}$ for 2006 and earlier vehicles down to $0.33 \mathrm{~g} / \mathrm{hr}$ for 2029 vehicles. [3] EPA does not provide CO, HC , or $\mathrm{SO}_{2}$ emission rates. EPA developed these emission rates from a variety of sources including the CRC data, but the guidance does not explain how EPA arrived at these emission rates.

## Uncertainties

The major uncertainty with the above emission rates is ARB's assumption that the NOx idle emission rates will not be lower in with lower NOx standards in the 2007 and later model years. The ARB is proposing to adopt controls that would either (1) require new engines to shutoff after a period of time, or (2) emit at below $30 \mathrm{~g} / \mathrm{hr}$. If these controls are adopted by the ARB, they could also be adopted by the EPA. If they are adopted by the EPA, it is likely that Environment Canada will implement an memorandum of understanding to require the controls in Canada as well. But even if none of this happens, it is likely that the 2007-2010 NOx emission reduction strategies will have some effect at reducing idle emissions from 2007 and later trucks. Thus, the idle NOx emission rates for 2020 in Table 4 are probably quite high.

Another uncertainty is whether the idle emission rates properly represent the duty cycle at the terminal. The creep emission rates indicate that the NOx emissions are probably appropriate, but if the duty cycle is more like the creep cycle than the idle cycle, then PM and HC emission rate will be somewhat higher.

## References

1. "Staff Report: Initial Statement of reasons, Notice of Public Hearing to Consider Requirements to Reduce Idling Emissions from New and In-Use Trucks, Beginning in 2008", September 1, 2005, California EPA, Air Resources Board.
2. "Heavy-Duty Vehicle Chassis Dynamometer Testing for Emission Inventory", CRC Project No. E-55/59, http:crcao.com
3. "Guidance for Quantifying and Using Long Duration Truck Idling Emission Reductions in State Implementation Plans and Transportation Conformity", EPA420-B-04-001, January 2004.

## Attachment 1

Low Idle and High Idle Emission Rates

| LOW IDLE | PM |  | NOx |  | CO |  | HC |  | CO 2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BASELINE |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |
| $2007+$ | 0.09 | 83.73 | 18.40 | 6.12 | 4366 |  |  |  |  |
| $2004-2006$ | 0.85 | 83.73 | 18.40 | 6.12 | 4366 |  |  |  |  |
| $1998-2003$ | 0.85 | 83.73 | 18.40 | 6.12 | 4366 |  |  |  |  |
| $1994-1997$ | 1.13 | 83.73 | 20.44 | 7.59 | 4366 |  |  |  |  |
| $1991-1993$ | 1.50 | 83.73 | 22.70 | 9.39 | 4366 |  |  |  |  |
| 1990 | 2.00 | 83.73 | 25.21 | 11.65 | 4366 |  |  |  |  |
| $1987-1989$ | 2.00 | 28.91 | 25.21 | 11.65 | 4366 |  |  |  |  |
| $1984-1986$ | 2.65 | 28.91 | 28.00 | 14.42 | 4366 |  |  |  |  |
| $1980-1983$ | 3.53 | 28.91 | 31.10 | 17.89 | 4366 |  |  |  |  |
| $1977-1979$ | 4.68 | 28.91 | 34.53 | 22.14 | 4366 |  |  |  |  |
| $1975-1976$ | 5.72 | 28.91 | 37.21 | 25.79 | 4366 |  |  |  |  |
| Pre-1975 | 6.73 | 28.91 | 39.51 | 29.15 | 4366 |  |  |  |  |


| High Idle Summer | PM | NOx | CO | HC | CO2 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2007+ | 0.213 | 174 | 57.6 | 10.5 | 10081 |
| 2004-2006 | 2.131 | 174 | 57.6 | 10.5 | 10081 |
| 1998-2003 | 2.131 | 174 | 57.6 | 10.5 | 10081 |
| 1994-1997 | 2.837 | 174 | 64.0 | 13.1 | 10081 |
| 1991-1993 | 3.761 | 174 | 71.0 | 16.2 | 10081 |
| 1990 | 5.007 | 174 | 78.9 | 20.1 | 10081 |
| 1987-1989 | 5.007 | 60 | 78.9 | 20.1 | 10081 |
| 1984-1986 | 6.639 | 60 | 87.6 | 24.8 | 10081 |
| 1980-1983 | 8.838 | 60 | 97.3 | 30.8 | 10081 |
| 1977-1979 | 11.719 | 60 | 108.1 | 38.1 | 10081 |
| 1975-1976 | 14.336 | 60 | 116.5 | 44. | 10081 |
| Pre-1975 | 16.863 | 60 | 123.6 | 50.2 | 10081 |
| High Idle Winter | PM | NOx | CO | HC | CO2 |
| 2007+ | 0.367 | 151.5 | 135.0 | 13.5 | 7823 |
| 2004-2006 | 3.666 | 151.5 | 135.0 | 13.5 | 7823 |
| 1998-2003 | 3.666 | 151.5 | 135.0 | 13.5 | 7823 |
| 1994-1997 | 4.880 | 151.5 | 149.9 | 16.7 | 7823 |
| 1991-1993 | 6.471 | 151.5 | 166.5 | 20.7 | 7823 |
| 1990 | 8.613 | 151.5 | 184.9 | 25.6 | 7823 |
| 1987-1989 | 8.613 | 52.3 | 184.9 | 25.6 | 7823 |
| 1984-1986 | 11.421 | 52.3 | 205.3 | 31.7 | 7823 |
| 1980-1983 | 15.203 | 52.3 | 228.1 | 39.4 | 7823 |
| 1977-1979 | 20.159 | 52.3 | 253.2 | 48.7 | 7823 |
| 1975-1976 | 24.661 | 52.3 | 272.9 | 56.7 | 7823 |
| Pre-1975 | 29.008 | 52.3 | 289.7 | 64.1 | 7823 |

## Attachment 2

Idle Correction Factors
High Idle Correction Factors

|  |  | PM | NOX | CO | HC | CO2 |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |
| Summer CF | CF1 | 2.51 | 2.08 | 3.13 | 1.72 | 2.31 |
| Winter CF | CF2 | 4.31 | 1.81 | 7.33 | 2.20 | 1.79 |

## Attachment 3

British Columbia HDDT Registration Distributions

| Age | Calendar Year 2000 <br> (used for 2003) | Calendar Year 2010 <br> (used for 2011) | Calendar Year 2020 <br> (used for 2020) |
| :---: | :---: | :---: | :---: |
| 1 | 0.079 | 0.0816 | 0.0835 |
| 2 | 0.086 | 0.0733 | 0.075 |
| 3 | 0.086 | 0.0685 | 0.0701 |
| 4 | 0.065 | 0.0641 | 0.0655 |
| 5 | 0.055 | 0.0599 | 0.0612 |
| 6 | 0.074 | 0.052 | 0.0515 |
| 7 | 0.066 | 0.0486 | 0.0482 |
| 8 | 0.044 | 0.0455 | 0.045 |
| 9 | 0.040 | 0.0426 | 0.0422 |
| 10 | 0.039 | 0.0397 | 0.0394 |
| 11 | 0.062 | 0.0372 | 0.0368 |
| 12 | 0.050 | 0.0348 | 0.0344 |
| 13 | 0.047 | 0.0325 | 0.0322 |
| 14 | 0.034 | 0.0305 | 0.0301 |
| 15 | 0.029 | 0.0284 | 0.0281 |
| 16 | 0.022 | 0.0267 | 0.0263 |
| 17 | 0.013 | 0.0249 | 0.0245 |
| 18 | 0.004 | 0.0233 | 0.023 |
| 19 | 0.008 | 0.0218 | 0.0215 |
| 20 | 0.018 | 0.0204 | 0.0201 |
| 21 | 0.017 | 0.0191 | 0.0188 |
| 22 | 0.013 | 0.0179 | 0.0176 |
| 23 | 0.007 | 0.0166 | 0.0164 |
| 24 | 0.009 | 0.0156 | 0.0154 |
| 25 | 0.034 | 0.0745 | 0.0734 |

* Only the 1997-2020 year data were used in this analysis for 2003, 2011, and 2020.

Attachment 4
CREEP Cycle Used in CRC E55/E57 (the second cycle is the same as the first, but has 4 repeats )



Table B-1 Summary of Creep Emission Factors for Windsor and Detroit for $\mathrm{NO}_{\mathrm{x}}(\mathrm{g} / \mathrm{veh} / \mathrm{hr})$

| Year | Windsor | Detroit |
| :---: | :---: | :---: |
| 2006 | 106.3 | 103.4 |
| 2007 | 109.1 | 106.1 |
| 2015 | 115.0 | 113.8 |
| 2025 | 116.2 | 116.4 |
| 2035 | 116.2 | 116.4 |

Table B-2 Summary of Creep Emission Factors for Windsor and Detroit for $\mathbf{P M}_{2.5}(\mathrm{~g} / \mathrm{veh} / \mathrm{hr})$

| Year | Windsor | Detroit |
| :---: | :---: | :---: |
| 2006 | 6.6 | 7.0 |
| 2007 | 6.3 | 6.6 |
| 2015 | 3.2 | 3.3 |
| 2025 | 1.0 | 1.4 |
| 2035 | 0.6 | 0.6 |

Table B-3 Working sheet showing Creep Emission Factor Calculation for Windsor and Detroit - Year 2015

|  |  |  |  |  | Creep Emission Factors |  |  | Windsor |  | Detroit |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | HDDT8 Distribution |  | year | Model Year | NOx EF | PM EF | PM2.5 EF | Weighted | Weighted | Weighted | Weighted |
| (Years) | Windsor | Detroit |  |  | g/hr | g/hr | g/hr | NOx | PM2.5 | NOx | PM2.5 |
| 0 | 0.016 | 0.03 | 2015 | 2007+ | 116.1 | 0.57 | 0.5529 |  |  |  |  |
| 1 | 0.071 | 0.08 | 2014 | 2007+ | 116.1 | 0.57 | 0.5529 | 8.2431 | 0.0392559 | 9.288 | 0.044232 |
| 2 | 0.058 | 0.067 | 2013 | 2007+ | 116.1 | 0.57 | 0.5529 | 6.7338 | 0.0320682 | 7.7787 | 0.0370443 |
| 3 | 0.049 | 0.076 | 2012 | 2007+ | 116.1 | 0.57 | 0.5529 | 5.6889 | 0.0270921 | 8.8236 | 0.0420204 |
| 4 | 0.044 | 0.066 | 2011 | 2007+ | 116.1 | 0.57 | 0.5529 | 5.1084 | 0.0243276 | 7.6626 | 0.0364914 |
| 5 | 0.046 | 0.08 | 2010 | 2007+ | 116.1 | 0.57 | 0.5529 | 5.3406 | 0.0254334 | 9.288 | 0.044232 |
| 6 | 0.072 | 0.06 | 2009 | 2007+ | 116.1 | 0.57 | 0.5529 | 8.3592 | 0.0398088 | 6.966 | 0.033174 |
| 7 | 0.078 | 0.044 | 2008 | 2007+ | 116.1 | 0.57 | 0.5529 | 9.0558 | 0.0431262 | 5.1084 | 0.0243276 |
| 8 | 0.087 | 0.036 | 2007 | 2007+ | 116.1 | 0.57 | 0.5529 | 10.1007 | 0.0481023 | 4.1796 | 0.0199044 |
| 9 | 0.074 | 0.036 | 2006 | 1991+ | 116.1 | 5.7 | 5.529 | 8.5914 | 0.409146 | 4.1796 | 0.199044 |
| 10 | 0.067 | 0.042 | 2005 | 1991+ | 116.1 | 5.7 | 5.529 | 7.7787 | 0.370443 | 4.8762 | 0.232218 |
| 11 | 0.064 | 0.048 | 2004 | 1991+ | 116.1 | 5.7 | 5.529 | 7.4304 | 0.353856 | 5.5728 | 0.265392 |
| 12 | 0.045 | 0.045 | 2003 | 1991+ | 116.1 | 5.7 | 5.529 | 5.2245 | 0.248805 | 5.2245 | 0.248805 |
| 13 | 0.03 | 0.04 | 2002 | 1991+ | 116.1 | 5.7 | 5.529 | 3.483 | 0.16587 | 4.644 | 0.22116 |
| 14 | 0.03 | 0.038 | 2001 | 1991+ | 116.1 | 5.7 | 5.529 | 3.483 | 0.16587 | 4.4118 | 0.210102 |
| 15 | 0.035 | 0.031 | 2000 | 1991+ | 116.1 | 5.7 | 5.529 | 4.0635 | 0.193515 | 3.5991 | 0.171399 |
| 16 | 0.02 | 0.026 | 1999 | 1991+ | 116.1 | 5.7 | 5.529 | 2.322 | 0.11058 | 3.0186 | 0.143754 |
| 17 | 0.022 | 0.013 | 1998 | 1991+ | 116.1 | 5.7 | 5.529 | 2.5542 | 0.121638 | 1.5093 | 0.071877 |
| 18 | 0.015 | 0.01 | 1997 | 1991+ | 116.1 | 5.7 | 5.529 | 1.7415 | 0.082935 | 1.161 | 0.05529 |
| 19 | 0.013 | 0.01 | 1996 | 1991+ | 116.1 | 5.7 | 5.529 | 1.5093 | 0.071877 | 1.161 | 0.05529 |
| 20 | 0.013 | 0.01 | 1995 | 1991+ | 116.1 | 5.7 | 5.529 | 1.5093 | 0.071877 | 1.161 | 0.05529 |
| 21 | 0.013 | 0.015 | 1994 | 1991+ | 116.1 | 5.7 | 5.529 | 1.5093 | 0.071877 | 1.7415 | 0.082935 |
| 22 | 0.009 | 0.015 | 1993 | 1991+ | 116.1 | 5.7 | 5.529 | 1.0449 | 0.049761 | 1.7415 | 0.082935 |
| 23 | 0.007 | 0.015 | 1992 | 1991+ | 116.1 | 5.7 | 5.529 | 0.8127 | 0.038703 | 1.7415 | 0.082935 |
| 24 | 0.016 | 0.05 | 1991 | 1991+ | 116.1 | 5.7 | 5.529 | 1.8576 | 0.088464 | 5.805 | 0.27645 |
| 25 | 0.023 | 0.05 | 1990 | pre 1991 | 62.7 | 11.7 | 11.349 | 1.4421 | 0.261027 | 3.135 | 0.56745 |
|  | 1.001 | 1.003 |  |  |  |  |  | 115.0 | 3.2 | 113.8 | 3.3 |

Table B-4 Working sheet showing Creep Emission Factor Calculation for Windsor and Detroit - Year 2025

|  |  |  |  |  | Creep Emission Factors |  |  | Windsor |  | Detroit |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | HDDT8 | Distribution | year | Model Year | NOx EF | PM EF | PM2.5 EF | Weighted | Weighted | Weighted | Weighted |
| (Years) | Windsor | Detroit |  |  | g/hr | g/hr | g/hr | NOx | PM2.5 | NOx | PM2.5 |
| 0 | 0.016 | 0.03 | 2025 | 2007+ | 116.1 | 0.57 | 0.5529 |  |  |  |  |
| 1 | 0.071 | 0.08 | 2024 | 2007+ | 116.1 | 0.57 | 0.5529 | 8.2431 | 0.0392559 | 9.288 | 0.044232 |
| 2 | 0.058 | 0.067 | 2023 | 2007+ | 116.1 | 0.57 | 0.5529 | 6.7338 | 0.0320682 | 7.7787 | 0.0370443 |
| 3 | 0.049 | 0.076 | 2022 | 2007+ | 116.1 | 0.57 | 0.5529 | 5.6889 | 0.0270921 | 8.8236 | 0.0420204 |
| 4 | 0.044 | 0.066 | 2021 | 2007+ | 116.1 | 0.57 | 0.5529 | 5.1084 | 0.0243276 | 7.6626 | 0.0364914 |
| 5 | 0.046 | 0.08 | 2020 | 2007+ | 116.1 | 0.57 | 0.5529 | 5.3406 | 0.0254334 | 9.288 | 0.044232 |
| 6 | 0.072 | 0.06 | 2019 | 2007+ | 116.1 | 0.57 | 0.5529 | 8.3592 | 0.0398088 | 6.966 | 0.033174 |
| 7 | 0.078 | 0.044 | 2018 | 2007+ | 116.1 | 0.57 | 0.5529 | 9.0558 | 0.0431262 | 5.1084 | 0.0243276 |
| 8 | 0.087 | 0.036 | 2017 | 2007+ | 116.1 | 0.57 | 0.5529 | 10.1007 | 0.0481023 | 4.1796 | 0.0199044 |
| 9 | 0.074 | 0.036 | 2016 | 2007+ | 116.1 | 0.57 | 0.5529 | 8.5914 | 0.0409146 | 4.1796 | 0.0199044 |
| 10 | 0.067 | 0.042 | 2015 | 2007+ | 116.1 | 0.57 | 0.5529 | 7.7787 | 0.0370443 | 4.8762 | 0.0232218 |
| 11 | 0.064 | 0.048 | 2014 | 2007+ | 116.1 | 0.57 | 0.5529 | 7.4304 | 0.0353856 | 5.5728 | 0.0265392 |
| 12 | 0.045 | 0.045 | 2013 | 2007+ | 116.1 | 0.57 | 0.5529 | 5.2245 | 0.0248805 | 5.2245 | 0.0248805 |
| 13 | 0.03 | 0.04 | 2012 | 2007+ | 116.1 | 0.57 | 0.5529 | 3.483 | 0.016587 | 4.644 | 0.022116 |
| 14 | 0.03 | 0.038 | 2011 | 2007+ | 116.1 | 0.57 | 0.5529 | 3.483 | 0.016587 | 4.4118 | 0.0210102 |
| 15 | 0.035 | 0.031 | 2010 | 2007+ | 116.1 | 0.57 | 0.5529 | 4.0635 | 0.0193515 | 3.5991 | 0.0171399 |
| 16 | 0.02 | 0.026 | 2009 | 2007+ | 116.1 | 0.57 | 0.5529 | 2.322 | 0.011058 | 3.0186 | 0.0143754 |
| 17 | 0.022 | 0.013 | 2008 | 2007+ | 116.1 | 0.57 | 0.5529 | 2.5542 | 0.0121638 | 1.5093 | 0.0071877 |
| 18 | 0.015 | 0.01 | 2007 | 2007+ | 116.1 | 0.57 | 0.5529 | 1.7415 | 0.0082935 | 1.161 | 0.005529 |
| 19 | 0.013 | 0.01 | 2006 | 1991+ | 116.1 | 5.7 | 5.529 | 1.5093 | 0.071877 | 1.161 | 0.05529 |
| 20 | 0.013 | 0.01 | 2005 | 1991+ | 116.1 | 5.7 | 5.529 | 1.5093 | 0.071877 | 1.161 | 0.05529 |
| 21 | 0.013 | 0.015 | 2004 | 1991+ | 116.1 | 5.7 | 5.529 | 1.5093 | 0.071877 | 1.7415 | 0.082935 |
| 22 | 0.009 | 0.015 | 2003 | 1991+ | 116.1 | 5.7 | 5.529 | 1.0449 | 0.049761 | 1.7415 | 0.082935 |
| 23 | 0.007 | 0.015 | 2002 | 1991+ | 116.1 | 5.7 | 5.529 | 0.8127 | 0.038703 | 1.7415 | 0.082935 |
| 24 | 0.016 | 0.05 | 2001 | 1991+ | 116.1 | 5.7 | 5.529 | 1.8576 | 0.088464 | 5.805 | 0.27645 |
| 25 | 0.023 | 0.05 | 2000 | 1991+ | 116.1 | 5.7 | 5.529 | 2.6703 | 0.127167 | 5.805 | 0.27645 |
|  | 1.001 | 1.003 |  |  |  |  |  | 116.2 | 1.0 | 116.4 | 1.4 |

Table B-5 Working sheet showing Creep Emission Factor Calculation for Windsor and Detroit - Year 2035

|  |  |  |  |  | Creep Emission Factors |  |  | Windsor |  | Detroit |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | HDDT8 Distribution |  | year | Model Year | NOx EF | PM EF | PM2.5 EF | Weighted | Weighted | Weighted | Weighted |
| (Years) | Windsor | Detroit |  |  | $\mathrm{g} / \mathrm{hr}$ | g/hr | g/hr | NOx | PM2.5 | NOx | PM2.5 |
| 0 | 0.016 | 0.03 | 2035 | 2007+ | 116.1 | 0.57 | 0.5529 |  |  |  |  |
| 1 | 0.071 | 0.08 | 2034 | 2007+ | 116.1 | 0.57 | 0.5529 | 8.2431 | 0.0392559 | 9.288 | 0.044232 |
| 2 | 0.058 | 0.067 | 2033 | 2007+ | 116.1 | 0.57 | 0.5529 | 6.7338 | 0.0320682 | 7.7787 | 0.0370443 |
| 3 | 0.049 | 0.076 | 2032 | 2007+ | 116.1 | 0.57 | 0.5529 | 5.6889 | 0.0270921 | 8.8236 | 0.0420204 |
| 4 | 0.044 | 0.066 | 2031 | 2007+ | 116.1 | 0.57 | 0.5529 | 5.1084 | 0.0243276 | 7.6626 | 0.0364914 |
| 5 | 0.046 | 0.08 | 2030 | 2007+ | 116.1 | 0.57 | 0.5529 | 5.3406 | 0.0254334 | 9.288 | 0.044232 |
| 6 | 0.072 | 0.06 | 2029 | 2007+ | 116.1 | 0.57 | 0.5529 | 8.3592 | 0.0398088 | 6.966 | 0.033174 |
| 7 | 0.078 | 0.044 | 2028 | 2007+ | 116.1 | 0.57 | 0.5529 | 9.0558 | 0.0431262 | 5.1084 | 0.0243276 |
| 8 | 0.087 | 0.036 | 2027 | 2007+ | 116.1 | 0.57 | 0.5529 | 10.1007 | 0.0481023 | 4.1796 | 0.0199044 |
| 9 | 0.074 | 0.036 | 2026 | 2007+ | 116.1 | 0.57 | 0.5529 | 8.5914 | 0.0409146 | 4.1796 | 0.0199044 |
| 10 | 0.067 | 0.042 | 2025 | 2007+ | 116.1 | 0.57 | 0.5529 | 7.7787 | 0.0370443 | 4.8762 | 0.0232218 |
| 11 | 0.064 | 0.048 | 2024 | 2007+ | 116.1 | 0.57 | 0.5529 | 7.4304 | 0.0353856 | 5.5728 | 0.0265392 |
| 12 | 0.045 | 0.045 | 2023 | 2007+ | 116.1 | 0.57 | 0.5529 | 5.2245 | 0.0248805 | 5.2245 | 0.0248805 |
| 13 | 0.03 | 0.04 | 2022 | 2007+ | 116.1 | 0.57 | 0.5529 | 3.483 | 0.016587 | 4.644 | 0.022116 |
| 14 | 0.03 | 0.038 | 2021 | 2007+ | 116.1 | 0.57 | 0.5529 | 3.483 | 0.016587 | 4.4118 | 0.0210102 |
| 15 | 0.035 | 0.031 | 2020 | 2007+ | 116.1 | 0.57 | 0.5529 | 4.0635 | 0.0193515 | 3.5991 | 0.0171399 |
| 16 | 0.02 | 0.026 | 2019 | 2007+ | 116.1 | 0.57 | 0.5529 | 2.322 | 0.011058 | 3.0186 | 0.0143754 |
| 17 | 0.022 | 0.013 | 2018 | 2007+ | 116.1 | 0.57 | 0.5529 | 2.5542 | 0.0121638 | 1.5093 | 0.0071877 |
| 18 | 0.015 | 0.01 | 2017 | 2007+ | 116.1 | 0.57 | 0.5529 | 1.7415 | 0.0082935 | 1.161 | 0.005529 |
| 19 | 0.013 | 0.01 | 2016 | 2007+ | 116.1 | 0.57 | 0.5529 | 1.5093 | 0.0071877 | 1.161 | 0.005529 |
| 20 | 0.013 | 0.01 | 2015 | 2007+ | 116.1 | 0.57 | 0.5529 | 1.5093 | 0.0071877 | 1.161 | 0.005529 |
| 21 | 0.013 | 0.015 | 2014 | 2007+ | 116.1 | 0.57 | 0.5529 | 1.5093 | 0.0071877 | 1.7415 | 0.0082935 |
| 22 | 0.009 | 0.015 | 2013 | 2007+ | 116.1 | 0.57 | 0.5529 | 1.0449 | 0.0049761 | 1.7415 | 0.0082935 |
| 23 | 0.007 | 0.015 | 2012 | 2007+ | 116.1 | 0.57 | 0.5529 | 0.8127 | 0.0038703 | 1.7415 | 0.0082935 |
| 24 | 0.016 | 0.05 | 2011 | 2007+ | 116.1 | 0.57 | 0.5529 | 1.8576 | 0.0088464 | 5.805 | 0.027645 |
| 25 | 0.023 | 0.05 | 2010 | 2007+ | 116.1 | 0.57 | 0.5529 | 2.6703 | 0.0127167 | 5.805 | 0.027645 |
|  | 1.001 | 1.003 |  |  |  |  |  | 116.2 | 0.6 | 116.4 | 0.6 |

Table B-8 Working sheet showing Idle Emission Factor Calculation for Windsor and Detroit - Year 2015

|  |  |  |  | Model Year | Idle Emission Factors |  |  |  |  | Windsor |  |  | Detroit |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | HDDT8 Distribution |  | year |  | NOx EF | PM EF | PM2.5 EF | CO2 EF | CO EF | Weighted | Weighted | Weighted | Weighted | Weighted | Weighted |
| (Years) | Windsor | Detroit |  |  | g/hr | g/hr | g/hr | g/hr | g/hr | NOx | PM2.5 | CO | NOx | PM2.5 | CO |
| 0 | 0.016 | 0.03 | 2015 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 |  |  |  |  |  |  |
| 1 | 0.071 | 0.08 | 2014 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 8.1863 | 0.0110192 | 3.6420515 | 9.224 | 0.012416 | 4.10372 |
| 2 | 0.058 | 0.067 | 2013 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 6.6874 | 0.0090016 | 2.975197 | 7.7251 | 0.0103984 | 3.4368655 |
| 3 | 0.049 | 0.076 | 2012 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 5.6497 | 0.0076048 | 2.5135285 | 8.7628 | 0.0117952 | 3.898534 |
| 4 | 0.044 | 0.066 | 2011 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 5.0732 | 0.0068288 | 2.257046 | 7.6098 | 0.0102432 | 3.385569 |
| 5 | 0.046 | 0.08 | 2010 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 5.3038 | 0.0071392 | 2.359639 | 9.224 | 0.012416 | 4.10372 |
| 6 | 0.072 | 0.06 | 2009 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 8.3016 | 0.0111744 | 3.693348 | 6.918 | 0.009312 | 3.07779 |
| 7 | 0.078 | 0.044 | 2008 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 8.9934 | 0.0121056 | 4.001127 | 5.0732 | 0.0068288 | 2.257046 |
| 8 | 0.087 | 0.036 | 2007 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 10.0311 | 0.0135024 | 4.4627955 | 4.1508 | 0.0055872 | 1.846674 |
| 9 | 0.074 | 0.036 | 2006 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 8.5322 | 0.136382 | 3.795941 | 4.1508 | 0.066348 | 1.846674 |
| 10 | 0.067 | 0.042 | 2005 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 7.7251 | 0.123481 | 3.4368655 | 4.8426 | 0.077406 | 2.154453 |
| 11 | 0.064 | 0.048 | 2004 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 7.3792 | 0.117952 | 3.282976 | 5.5344 | 0.088464 | 2.462232 |
| 12 | 0.045 | 0.045 | 2003 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 5.1885 | 0.082935 | 2.3083425 | 5.1885 | 0.082935 | 2.3083425 |
| 13 | 0.03 | 0.04 | 2002 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 3.459 | 0.05529 | 1.538895 | 4.612 | 0.07372 | 2.05186 |
| 14 | 0.03 | 0.038 | 2001 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 3.459 | 0.05529 | 1.538895 | 4.3814 | 0.070034 | 1.949267 |
| 15 | 0.035 | 0.031 | 2000 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 4.0355 | 0.064505 | 1.7953775 | 3.5743 | 0.057133 | 1.5901915 |
| 16 | 0.02 | 0.026 | 1999 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 2.306 | 0.03686 | 1.02593 | 2.9978 | 0.047918 | 1.333709 |
| 17 | 0.022 | 0.013 | 1998 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 2.5366 | 0.040546 | 1.128523 | 1.4989 | 0.023959 | 0.6668545 |
| 18 | 0.015 | 0.01 | 1997 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 56.97065 | 1.7295 | 0.027645 | 0.85455975 | 1.153 | 0.01843 | 0.5697065 |
| 19 | 0.013 | 0.01 | 1996 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 56.97065 | 1.4989 | 0.023959 | 0.74061845 | 1.153 | 0.01843 | 0.5697065 |
| 20 | 0.013 | 0.01 | 1995 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 56.97065 | 1.4989 | 0.023959 | 0.74061845 | 1.153 | 0.01843 | 0.5697065 |
| 21 | 0.013 | 0.015 | 1994 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 56.97065 | 1.4989 | 0.023959 | 0.74061845 | 1.7295 | 0.027645 | 0.85455975 |
| 22 | 0.009 | 0.015 | 1993 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 63.26325 | 1.0377 | 0.016587 | 0.56936925 | 1.7295 | 0.027645 | 0.94894875 |
| 23 | 0.007 | 0.015 | 1992 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 63.26325 | 0.8071 | 0.012901 | 0.44284275 | 1.7295 | 0.027645 | 0.94894875 |
| 24 | 0.016 | 0.05 | 1991 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 63.26325 | 1.8448 | 0.029488 | 1.012212 | 5.765 | 0.09215 | 3.1631625 |
| 25 | 0.023 | 0.05 | 1990 | pre 1991 | 39.8 | 5.3 | 5.141 | 6228 | 70.2641 | 0.9154 | 0.118243 | 1.6160743 | 1.99 | 0.25705 | 3.513205 |
|  | 1.001 | 1.003 |  |  |  |  |  |  |  | 113.7 | 1.1 | 52.5 | 111.9 | 1.2 | 53.6 |

Table B-9 Working sheet showing Idle Emission Factor Calculation for Windsor and Detroit - Year 2025

|  |  |  |  | Model <br> Year | Idle Emission Factors |  |  |  |  | Windsor |  |  | Detroit |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | HDDT8 Distribution |  | year |  | NOx EF | PM EF | PM2.5 EF | CO2 EF | CO EF | Weighted | Weighted | Weighted | Weighted | Weighted | Weighted |
| (Years) | Windsor | Detroit |  |  | g/hr | $\mathrm{g} / \mathrm{hr}$ | g/hr | g/hr | g/hr | NOx | PM2.5 | CO | NOx | PM2.5 | CO |
| 0 | 0.016 | 0.03 | 2025 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 |  |  |  |  |  |  |
| 1 | 0.071 | 0.08 | 2024 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 8.1863 | 0.0110192 | 3.6420515 | 9.224 | 0.012416 | 4.10372 |
| 2 | 0.058 | 0.067 | 2023 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 6.6874 | 0.0090016 | 2.975197 | 7.7251 | 0.0103984 | 3.4368655 |
| 3 | 0.049 | 0.076 | 2022 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 5.6497 | 0.0076048 | 2.5135285 | 8.7628 | 0.0117952 | 3.898534 |
| 4 | 0.044 | 0.066 | 2021 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 5.0732 | 0.0068288 | 2.257046 | 7.6098 | 0.0102432 | 3.385569 |
| 5 | 0.046 | 0.08 | 2020 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 5.3038 | 0.0071392 | 2.359639 | 9.224 | 0.012416 | 4.10372 |
| 6 | 0.072 | 0.06 | 2019 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 8.3016 | 0.0111744 | 3.693348 | 6.918 | 0.009312 | 3.07779 |
| 7 | 0.078 | 0.044 | 2018 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 8.9934 | 0.0121056 | 4.001127 | 5.0732 | 0.0068288 | 2.257046 |
| 8 | 0.087 | 0.036 | 2017 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 10.0311 | 0.0135024 | 4.4627955 | 4.1508 | 0.0055872 | 1.846674 |
| 9 | 0.074 | 0.036 | 2016 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 8.5322 | 0.0114848 | 3.795941 | 4.1508 | 0.0055872 | 1.846674 |
| 10 | 0.067 | 0.042 | 2015 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 7.7251 | 0.0103984 | 3.4368655 | 4.8426 | 0.0065184 | 2.154453 |
| 11 | 0.064 | 0.048 | 2014 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 7.3792 | 0.0099328 | 3.282976 | 5.5344 | 0.0074496 | 2.462232 |
| 12 | 0.045 | 0.045 | 2013 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 5.1885 | 0.006984 | 2.3083425 | 5.1885 | 0.006984 | 2.3083425 |
| 13 | 0.03 | 0.04 | 2012 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 3.459 | 0.004656 | 1.538895 | 4.612 | 0.006208 | 2.05186 |
| 14 | 0.03 | 0.038 | 2011 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 3.459 | 0.004656 | 1.538895 | 4.3814 | 0.0058976 | 1.949267 |
| 15 | 0.035 | 0.031 | 2010 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 4.0355 | 0.005432 | 1.7953775 | 3.5743 | 0.0048112 | 1.5901915 |
| 16 | 0.02 | 0.026 | 2009 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 2.306 | 0.003104 | 1.02593 | 2.9978 | 0.0040352 | 1.333709 |
| 17 | 0.022 | 0.013 | 2008 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 2.5366 | 0.0034144 | 1.128523 | 1.4989 | 0.0020176 | 0.6668545 |
| 18 | 0.015 | 0.01 | 2007 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 1.7295 | 0.002328 | 0.7694475 | 1.153 | 0.001552 | 0.512965 |
| 19 | 0.013 | 0.01 | 2006 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 1.4989 | 0.023959 | 0.6668545 | 1.153 | 0.01843 | 0.512965 |
| 20 | 0.013 | 0.01 | 2005 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 1.4989 | 0.023959 | 0.6668545 | 1.153 | 0.01843 | 0.512965 |
| 21 | 0.013 | 0.015 | 2004 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 1.4989 | 0.023959 | 0.6668545 | 1.7295 | 0.027645 | 0.7694475 |
| 22 | 0.009 | 0.015 | 2003 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 1.0377 | 0.016587 | 0.4616685 | 1.7295 | 0.027645 | 0.7694475 |
| 23 | 0.007 | 0.015 | 2002 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 0.8071 | 0.012901 | 0.3590755 | 1.7295 | 0.027645 | 0.7694475 |
| 24 | 0.016 | 0.05 | 2001 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 1.8448 | 0.029488 | 0.820744 | 5.765 | 0.09215 | 2.564825 |
| 25 | 0.023 | 0.05 | 2000 | 1991+ | 115.3 | 1.9 | 1.843 | 6228 | 51.2965 | 2.6519 | 0.042389 | 1.1798195 | 5.765 | 0.09215 | 2.564825 |
|  | 1.001 | 1.003 |  |  |  |  |  |  |  | 115.4 | 0.3 | 51.3 | 115.6 | 0.4 | 51.5 |

Table B-10 Working sheet showing Idle Emission Factor Calculation for Windsor and Detroit - Year 2035

|  |  |  |  |  | Idle Emission Factors |  |  |  |  | Windsor |  |  | Detroit |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\frac{\text { Age }}{\text { (Years) }}$ | HDDT8 Distribution |  | year | Model Year | NOx EF | $\begin{array}{\|r\|} \hline \text { PM EF } \\ \hline \mathbf{g} / \mathbf{h r} \\ \hline \end{array}$ | $\begin{gathered} \text { PM2.5 EF } \\ \hline \mathbf{g} / \mathbf{h r} \\ \hline \end{gathered}$ | $\begin{array}{\|c\|} \mathrm{CO} 2 \mathrm{EF} \\ \hline \mathbf{g} / \mathbf{h r} \\ \hline \end{array}$ | $\begin{array}{\|r} \hline \mathrm{CO} \mathrm{EF} \\ \hline \mathbf{g} / \mathbf{h r} \end{array}$ | $\begin{gathered} \text { Weighted } \\ \hline \text { NOx } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Weighted } \\ \hline \text { PM2.5 } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Weighted } \\ \hline \mathrm{CO} \\ \hline \end{gathered}$ | $\frac{\text { Weighted }}{} \frac{\text { NOx }}{}$ | $\begin{gathered} \text { Weighted } \\ \hline \text { PM2.5 } \end{gathered}$ | $\begin{gathered} \text { Weighted } \\ \hline \text { CO } \\ \hline \end{gathered}$ |
|  | Windsor | Detroit |  |  | g/hr |  |  |  |  |  |  |  |  |  |  |
| 0 | 0.016 | 0.03 | 2035 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 |  |  |  |  |  |  |
| 1 | 0.071 | 0.08 | 2034 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 8.1863 | 0.0110192 | 3.6420515 | 9.224 | 0.012416 | 4.10372 |
| 2 | 0.058 | 0.067 | 2033 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 6.6874 | 0.0090016 | 2.975197 | 7.7251 | 0.0103984 | 3.4368655 |
| 3 | 0.049 | 0.076 | 2032 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 5.6497 | 0.0076048 | 2.5135285 | 8.7628 | 0.0117952 | 3.898534 |
| 4 | 0.044 | 0.066 | 2031 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 5.0732 | 0.0068288 | 2.257046 | 7.6098 | 0.0102432 | 3.385569 |
| 5 | 0.046 | 0.08 | 2030 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 5.3038 | 0.0071392 | 2.359639 | 9.224 | 0.012416 | 4.10372 |
| 6 | 0.072 | 0.06 | 2029 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 8.3016 | 0.0111744 | 3.693348 | 6.918 | 0.009312 | 3.07779 |
| 7 | 0.078 | 0.044 | 2028 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 8.9934 | 0.0121056 | 4.001127 | 5.0732 | 0.0068288 | 2.257046 |
| 8 | 0.087 | 0.036 | 2027 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 10.0311 | 0.0135024 | 4.4627955 | 4.1508 | 0.0055872 | 1.846674 |
| 9 | 0.074 | 0.036 | 2026 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 8.5322 | 0.0114848 | 3.795941 | 4.1508 | 0.0055872 | 1.846674 |
| 10 | 0.067 | 0.042 | 2025 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 7.7251 | 0.0103984 | 3.4368655 | 4.8426 | 0.0065184 | 2.154453 |
| 11 | 0.064 | 0.048 | 2024 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 7.3792 | 0.0099328 | 3.282976 | 5.5344 | 0.0074496 | 2.462232 |
| 12 | 0.045 | 0.045 | 2023 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 5.1885 | 0.006984 | 2.3083425 | 5.1885 | 0.006984 | 2.3083425 |
| 13 | 0.03 | 0.04 | 2022 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 3.459 | 0.004656 | 1.538895 | 4.612 | 0.006208 | 2.05186 |
| 14 | 0.03 | 0.038 | 2021 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 3.459 | 0.004656 | 1.538895 | 4.3814 | 0.0058976 | 1.949267 |
| 15 | 0.035 | 0.031 | 2020 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 4.0355 | 0.005432 | 1.7953775 | 3.5743 | 0.0048112 | 1.5901915 |
| 16 | 0.02 | 0.026 | 2019 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 2.306 | 0.003104 | 1.02593 | 2.9978 | 0.0040352 | 1.333709 |
| 17 | 0.022 | 0.013 | 2018 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 2.5366 | 0.0034144 | 1.128523 | 1.4989 | 0.0020176 | 0.6668545 |
| 18 | 0.015 | 0.01 | 2017 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 1.7295 | 0.002328 | 0.7694475 | 1.153 | 0.001552 | 0.512965 |
| 19 | 0.013 | 0.01 | 2016 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 1.4989 | 0.0020176 | 0.6668545 | 1.153 | 0.001552 | 0.512965 |
| 20 | 0.013 | 0.01 | 2015 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 1.4989 | 0.0020176 | 0.6668545 | 1.153 | 0.001552 | 0.512965 |
| 21 | 0.013 | 0.015 | 2014 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 1.4989 | 0.0020176 | 0.6668545 | 1.7295 | 0.002328 | 0.7694475 |
| 22 | 0.009 | 0.015 | 2013 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 1.0377 | 0.0013968 | 0.4616685 | 1.7295 | 0.002328 | 0.7694475 |
| 23 | 0.007 | 0.015 | 2012 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 0.8071 | 0.0010864 | 0.3590755 | 1.7295 | 0.002328 | 0.7694475 |
| 24 | 0.016 | 0.05 | 2011 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 1.8448 | 0.0024832 | 0.820744 | 5.765 | 0.00776 | 2.564825 |
| 25 | 0.023 | 0.05 | 2010 | 2007+ | 115.3 | 0.16 | 0.1552 | 6228 | 51.2965 | 2.6519 | 0.0035696 | 1.1798195 | 5.765 | 0.00776 | 2.564825 |
|  | 1.001 | 1.003 |  |  |  |  |  |  |  | 115.4 | 0.2 | 51.3 | 115.6 | 0.2 | 51.5 |



## SAMPLE CALCULATIONS

## $\mathbf{P M}_{2.5}$ Emissions

Emissions of particulate (TSP, $\mathrm{PM}_{10}$, and $\mathrm{PM}_{2.5}$ ) from vehicle travel on roadways results both from tailpipe emissions and recirculation of road dust.

## 1. Tailpipe Emissions

Tailpipe emissions from vehicle travel were calculated by applying a fleet averaged emission factor from the Mobile 6C Emissions model for each horizon year. For the public roads, traffic data on AADT levels was supplied by IBI Group. The emission factors output from Mobile 6C have been included below in the following tables. As indicates in the tables, the emission factors are dependent upon vehicle type, country of origin (of vehicle), vehicle speed and analysis year. The $\mathrm{PM}_{2.5}$ and NOx emission factors have been highlighted, as they are the two contaminants that have been assessed at this point in time. All contaminants will be included in the final analysis.

As both cars and trucks travel on the same roadways, an average fleet tailpipe emission factor must be calculated.
(a) $\mathrm{VKT}_{\text {Total }}=\mathrm{VKT}_{\text {CDN_car }}+\mathrm{VKT}_{\text {CDN_truck }}+\mathrm{VKT}_{\mathrm{US}_{\_} \text {ar }}+\mathrm{VKT}_{\text {US_Truck }}$
(b) Fleet Average $\mathrm{EF}_{(\mathrm{g} / \mathrm{VKT})}=$

$$
E F_{C D N_{-} \text {car }} * \frac{V K T_{\text {CDN_car }}}{V K T_{\text {Total }}}+E F_{C D N_{-} \text {truck }} \frac{V K T_{C D N_{-} \text {truck }}}{V K T_{\text {car }}}+E F_{U S_{-} \text {car }} \frac{V K T_{U S_{-} \text {car }}}{V K T_{\text {Total }}}+E F_{U S_{-} \text {truck }} \frac{V K T_{\text {US_truck }}}{V K T_{\text {Total }}}
$$

Table 1a - 2015 Canadian Car Tailpipe Emissions (g/VKT)

| Speed (km/h) | PM | $\mathrm{PM}_{10}$ | $\mathrm{PM}_{2.5}$ | NOx | SOx | CO | $\mathrm{CO}_{2}$ | VOC | Bn | Ac | Fm | Bu | Acr |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Idle | 0.0161 | g .0161 | 0.0086 | 1.32 | 0.0108 | 29.3 | 1398.5 | 2.70 | 0.0532 | 0.0084 | 0.0196 | 0.0058 | 0.0014 |
| 25 | 0.0040 | 0.0040 | 0.0021 | 0.44 | 0.0047 | 6.4 | 347.6 | 0.38 | 0.0108 | 0.0020 | 0.0047 | 0.0012 | 0.0003 |
| 50 | 0.0040 | 0.0040 | 0.0021 | 0.40 | 0.0047 | 5.9 | 347.6 | 0.28 | 0.0087 | 0.0014 | 0.0033 | 0.0010 | 0.0002 |
| 75 | 0.0040 | 0.0040 | 0.0021 | 0.49 | 0.0047 | 6.6 | 347.6 | 0.27 | 0.0085 | 0.0013 | 0.0031 | 0.0009 | 0.0002 |
| 100 | 0.0040 | 0.0040 | 0.0021 | 0.49 | 0.0047 | 6.6 | 347.6 | 0.27 | 0.0085 | 0.0013 | 0.0031 | 0.0009 | 0.0002 |

$\mathrm{Bn}=$ Benzene, $\mathrm{Ac}=$ Acetaldehyde, $\mathrm{Fm}=$ Formaldehyde, $\mathrm{Bu}=1,3$ Butadiene, $\mathrm{Acr}=$ Acrolein

Table 1b - 2015 CanadianTruck Tailpipe Emissions (g/VKT)

| Speed (km/h) | PM | $\mathrm{PM}_{10}$ | $\mathrm{PM}_{2.5}$ | NOx | SOx | CO | $\mathrm{CO}_{2}$ | VOC | Bn | Ac | Fm | Bu | Acr |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Idle | 1.1015 | 1.1015 | 1.07 | 113.68 | 0.08 | 52.50 | 6228 | 1.02 | 0.0113 | 0.0309 | 0.0838 | 0.0065 | 0.0050 |
| 25 | 0.0191 | 0.0191 | 0.01 | 2.35 | 0.007 | 0.96 | 960 | 0.33 | 0.0036 | 0.0099 | 0.0268 | 0.0021 | 0.0016 |
| 50 | 0.0191 | 0.0191 | 0.01 | 2.02 | 0.007 | 0.49 | 960 | 0.19 | 0.0020 | 0.0056 | 0.0152 | 0.0012 | 0.0009 |
| 75 | 0.0191 | 0.0191 | 0.01 | 2.91 | 0.007 | 0.51 | 960 | 0.16 | 0.0018 | 0.0048 | 0.0131 | 0.0010 | 0.0008 |
| 100 | 0.0191 | 0.0191 | 0.01 | 2.91 | 0.007 | 0.51 | 960 | 0.16 | 0.0018 | 0.0048 | 0.0131 | 0.0010 | 0.0008 |

$\mathrm{Bn}=$ Benzene, $\mathrm{Ac}=$ Acetaldehyde, $\mathrm{Fm}=$ Formaldehyde, $\mathrm{Bu}=1,3$ Butadiene, Acr = Acrolein

Table 1c-2015 American Car Tailpipe Emissions (g/VKT)

| Speed (km/h) | PM | $\mathrm{PM}_{10}$ | $\mathrm{PM}_{2.5}$ | NOx | SOx | CO | $\mathrm{CO}_{2}$ | VOC | Bn | Ac | Fm | Bu | Acr |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Idle | 0.0158 | 0.0158 | 0.0086 | 1.20 | 0.0123 | 25.0 | 1405 | 2.34 | 0.0577 | 0.0080 | 0.0174 | 0.0050 | 0.0012 |
| 25 | 0.0039 | 0.0039 | 0.0021 | 0.40 | 0.0055 | 5.5 | 349 | 0.33 | 0.0118 | 0.0019 | 0.0043 | 0.0011 | 0.0003 |
| 50 | 0.0039 | 0.0039 | 0.0021 | 0.36 | 0.0056 | 5.1 | 349 | 0.25 | 0.0096 | 0.0013 | 0.0029 | 0.0008 | 0.0002 |
| 75 | 0.0039 | 0.0039 | 0.0021 | 0.44 | 0.0056 | 5.7 | 349 | 0.24 | 0.0094 | 0.0013 | 0.0028 | 0.0008 | 0.0002 |
| 100 | 0.0039 | 0.0039 | 0.0021 | 0.44 | 0.0056 | 5.7 | 349 | 0.24 | 0.0094 | 0.0013 | 0.0028 | 0.0008 | 0.0002 |

$\mathrm{Bn}=$ Benzene, $\mathrm{Ac}=$ Acetaldehyde, $\mathrm{Fm}=$ Formaldehyde, $\mathrm{Bu}=1,3$ Butadiene, $\mathrm{Acr}=$ Acrolein

Table 1d-2015 American Truck Tailpipe Emissions (g/VKT)

| Speed (km/h) | PM | $\mathrm{PM}_{10}$ | $\mathrm{PM}_{2.5}$ | NOx | SOx | CO | $\mathrm{CO}_{2}$ | VOC | Bn | Ac | Fm | Bu | Acr |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Idle | 1.1901 | 1.1901 | 1.1543 | 111.9 | 0.0800 | 53.60 | 6228 | 1.00 | 0.0111 | 0.0303 | 0.0822 | 0.0064 | 0.0049 |
| 25 | 0.0181 | 0.0181 | 0.0119 | 1.9 | 0.0066 | 0.83 | 960 | 0.32 | 0.0035 | 0.0097 | 0.0263 | 0.0021 | 0.0016 |
| 50 | 0.0181 | 0.0181 | 0.0119 | 1.7 | 0.0066 | 0.43 | 960 | 0.18 | 0.0020 | 0.0055 | 0.0149 | 0.0012 | 0.0009 |
| 75 | 0.0181 | 0.0181 | 0.0119 | 2.4 | 0.0066 | 0.44 | 960 | 0.16 | 0.0017 | 0.0047 | 0.0128 | 0.0010 | 0.0008 |
| 100 | 0.0181 | 0.0181 | 0.0119 | 2.4 | 0.0066 | 0.44 | 960 | 0.16 | 0.0017 | 0.0047 | 0.0128 | 0.0010 | 0.0008 |

[^9]Table 2a-2025 Canadian Car Tailpipe Emissions (g/VKT)

| Speed (km/h) | PM | $\mathrm{PM}_{10}$ | $\mathrm{PM}_{2.5}$ | NOx | SOx | CO | $\mathrm{CO}_{2}$ | VOC | Bn | Ac | Fm | Bu | Acr |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Idle | 0.0141 | 0.0141 | 0.0066 | 0.63 | 0.0108 | 26.56 | 1411 | 2.26 | 0.0433 | 0.0069 | 0.0159 | 0.0048 | 0.0011 |
| 25 | 0.0035 | 0.0035 | 0.0016 | 0.20 | 0.0048 | 5.77 | 351 | 0.31 | 0.0087 | 0.0016 | 0.0038 | 0.0010 | 0.0003 |
| 50 | 0.0035 | 0.0035 | 0.0016 | 0.18 | 0.0048 | 5.34 | 351 | 0.23 | 0.0071 | 0.0012 | 0.0027 | 0.0008 | 0.0002 |
| 75 | 0.0035 | 0.0035 | 0.0016 | 0.21 | 0.0048 | 6.00 | 351 | 0.21 | 0.0070 | 0.0011 | 0.0025 | 0.0008 | 0.0002 |
| 100 | 0.0035 | 0.0035 | 0.0016 | 0.21 | 0.0048 | 6.00 | 351 | 0.21 | 0.0070 | 0.0011 | 0.0025 | 0.0008 | 0.0002 |

$\mathrm{Bn}=$ Benzene, $\mathrm{Ac}=$ Acetaldehyde, $\mathrm{Fm}=$ Formaldehyde, $\mathrm{Bu}=1,3$ Butadiene, Acr = Acrolein
Table 2b - 2025 Canadian Truck Tailpipe Emissions (g/VKT)

| Speed (km/h) | PM | $\mathrm{PM}_{10}$ | $\mathrm{PM}_{2.5}$ | NOx | SOx | CO | $\mathrm{CO}_{2}$ | VOC | Bn | Ac | Fm | Bu | Acr |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Idle | 0.0476 | 0.0476 | 0.3140 | 115.42 | 0.0800 | 51.30 | 6228 | 0.8575 | 0.0094 | 0.0259 | 0.0702 | 0.0055 | 0.0042 |
| 25 | 0.0118 | 0.0118 | 0.0062 | 0.46 | 0.0071 | 0.31 | 960 | 0.2740 | 0.0030 | 0.0083 | 0.0225 | 0.0018 | 0.0013 |
| 50 | 0.0118 | 0.0118 | 0.0062 | 0.39 | 0.0071 | 0.16 | 960 | 0.1553 | 0.0017 | 0.0047 | 0.0128 | 0.0010 | 0.0008 |
| 75 | 0.0118 | 0.0118 | 0.0062 | 0.57 | 0.0071 | 0.16 | 960 | 0.1336 | 0.0015 | 0.0040 | 0.0110 | 0.0009 | 0.0007 |
| 100 | 0.0118 | 0.0118 | 0.0062 | 0.57 | 0.0071 | 0.16 | 960 | 0.1336 | 0.0015 | 0.0040 | 0.0110 | 0.0009 | 0.0007 |

$\mathrm{Bn}=$ Benzene, $\mathrm{Ac}=$ Acetaldehyde, $\mathrm{Fm}=$ Formaldehyde, $\mathrm{Bu}=1,3$ Butadiene, Acr = Acrolein
Table 2c-2025 American Car Tailpipe Emissions (g/VKT)

| Speed (km/h) | PM | $\mathrm{PM}_{10}$ | $\mathrm{PM}_{2.5}$ | NOx | SOx | CO | $\mathrm{CO}_{2}$ | VOC | Bn | Ac | Fm | Bu | Acr |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Idle | 0.0141 | 0.0141 | 0.0067 | 0.59 | 0.0123 | 22.0 | 1417 | 1.88 | 0.0454 | 0.0064 | 0.0141 | 0.0040 | 0.0010 |
| 25 | 0.0035 | 0.0035 | 0.0016 | 0.19 | 0.0056 | 4.8 | 352 | 0.26 | 0.0092 | 0.0015 | 0.0035 | 0.0009 | 0.0002 |
| 50 | 0.0035 | 0.0035 | 0.0016 | 0.17 | 0.0057 | 4.5 | 352 | 0.19 | 0.0076 | 0.0011 | 0.0024 | 0.0007 | 0.0002 |
| 75 | 0.0035 | 0.0035 | 0.0016 | 0.20 | 0.0057 | 5.0 | 352 | 0.18 | 0.0075 | 0.0010 | 0.0022 | 0.0007 | 0.0002 |
| 100 | 0.0035 | 0.0035 | 0.0016 | 0.20 | 0.0057 | 5.0 | 352 | 0.18 | 0.0075 | 0.0010 | 0.0022 | 0.0007 | 0.0002 |

$\mathrm{Bn}=$ Benzene, $\mathrm{Ac}=$ Acetaldehyde, $\mathrm{Fm}=$ Formaldehyde, $\mathrm{Bu}=1,3$ Butadiene, Acr = Acrolein
Table 2d - 2025 American Truck Tailpipe Emissions (g/VKT)

| Speed (km/h) | PM | $\mathrm{PM}_{10}$ | $\mathrm{PM}_{2.5}$ | NOx | SOx | CO | $\mathrm{CO}_{2}$ | VOC | Bn | Ac | Fm | Bu | Acr |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Idle | 0.0483 | 0.0483 | 0.4342 | 115.65 | 0.0800 | 51.50 | 6228 | 0.86 | 0.0095 | 0.0261 | 0.0708 | 0.0055 | 0.0042 |
| 25 | 0.0120 | 0.0120 | 0.0063 | 0.50 | 0.0066 | 0.32 | 960 | 0.28 | 0.0030 | 0.0083 | 0.0226 | 0.0018 | 0.0014 |
| 50 | 0.0120 | 0.0120 | 0.0063 | 0.43 | 0.0066 | 0.16 | 960 | 0.16 | 0.0017 | 0.0047 | 0.0128 | 0.0010 | 0.0008 |
| 75 | 0.0120 | 0.0120 | 0.0063 | 0.63 | 0.0066 | 0.17 | 960 | 0.13 | 0.0015 | 0.0041 | 0.0111 | 0.0009 | 0.0007 |
| 100 | 0.0120 | 0.0120 | 0.0063 | 0.63 | 0.0066 | 0.17 | 960 | 0.13 | 0.0015 | 0.0041 | 0.0111 | 0.0009 | 0.0007 |

$\mathrm{Bn}=$ Benzene, $\mathrm{Ac}=$ Acetaldehyd, $\mathrm{Fm}=$ Formaldehyde, $\mathrm{Bu}=1,3$ Butadiene, Acr = Acrolein

Table 3a-2035 Canadian Car Tailpipe Emissions (g/VKT)

| Speed (km/h) | PM | $\mathrm{PM}_{10}$ | $\mathrm{PM}_{2.5}$ | NOx | SOx | CO | $\mathrm{CO}_{2}$ | VOC | Bn | Ac | Fm | Bu | Acr |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Idle | 0.0139 | 0.0139 | 0.0065 | 0.58 | 0.0108 | 26.4 | 1411 | 2.24 | 0.0425 | 0.0068 | 0.0156 | 0.0047 | 0.0011 |
| 25 | 0.0034 | 0.0034 | 0.0016 | 0.18 | 0.0048 | 5.7 | 351 | 0.30 | 0.0086 | 0.0016 | 0.0038 | 0.0010 | 0.0003 |
| 50 | 0.0034 | 0.0034 | 0.0016 | 0.17 | 0.0048 | 5.3 | 351 | 0.22 | 0.0070 | 0.0011 | 0.0026 | 0.0008 | 0.0002 |
| 75 | 0.0034 | 0.0034 | 0.0016 | 0.19 | 0.0048 | 6.0 | 351 | 0.21 | 0.0069 | 0.0011 | 0.0025 | 0.0007 | 0.0002 |
| 100 | 0.0034 | 0.0034 | 0.0016 | 0.19 | 0.0048 | 6.0 | 351 | 0.21 | 0.0069 | 0.0011 | 0.0025 | 0.0007 | 0.0002 |

$\mathrm{n}=$ Formaldehyde, $\mathrm{Bu}=1,3$ Butadiene, Acr = Acrolein
Table 3b-2035 Canadian Truck Tailpipe Emissions (g/VKT)

| Speed $(\mathrm{km} / \mathrm{h})$ | PM | $\mathrm{PM}_{10}$ | $\mathrm{PM}_{2.5}$ | NOx | SOx | CO | $\mathrm{CO}_{2}$ | VOC | Bn | Ac | Fm | Bu | Acr |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Idle | 0.0458 | 0.0458 | 0.1554 | 115.42 | 0.0800 | 51.30 | 6228 | 0.85 | 0.0093 | 0.0255 | 0.0693 | 0.0054 | 0.0041 |
| 25 | 0.0114 | 0.0114 | 0.0058 | 0.34 | 0.0071 | 0.26 | 960 | 0.27 | 0.0030 | 0.0082 | 0.0222 | 0.0017 | 0.0013 |
| 50 | 0.0114 | 0.0114 | 0.0058 | 0.29 | 0.0071 | 0.13 | 960 | 0.15 | 0.0017 | 0.0046 | 0.0126 | 0.0010 | 0.0008 |
| 75 | 0.0114 | 0.0114 | 0.0058 | 0.43 | 0.0071 | 0.14 | 960 | 0.13 | 0.0015 | 0.0040 | 0.0108 | 0.0008 | 0.0006 |
| 100 | 0.0114 | 0.0114 | 0.0058 | 0.43 | 0.0071 | 0.14 | 960 | 0.13 | 0.0015 | 0.0040 | 0.0108 | 0.0008 | 0.0006 |

n = Formaldehyde, $\mathrm{Bu}=1,3$ Butadiene, Acr = Acrolein
Table 3c - 2035 American Car Tailpipe Emissions (g/VKT)

| Speed (km/h) | PM | $\mathrm{PM}_{10}$ | $\mathrm{PM}_{2.5}$ | NOx | SOx | CO | $\mathrm{CO}_{2}$ | VOC | Bn | Ac | Fm | Bu | Acr |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Idle | 0.0139 | 0.0139 | 0.0065 | 0.52 | 0.0123 | 21.8 | 1417 | 1.85 | 0.0443 | 0.0062 | 0.0136 | 0.0039 | 0.0009 |
| 25 | 0.0034 | 0.0034 | 0.0016 | 0.16 | 0.0056 | 4.8 | 352 | 0.25 | 0.0090 | 0.0015 | 0.0034 | 0.0008 | 0.0002 |
| 50 | 0.0034 | 0.0034 | 0.0016 | 0.15 | 0.0057 | 4.4 | 352 | 0.19 | 0.0074 | 0.0011 | 0.0023 | 0.0006 | 0.0002 |
| 75 | 0.0034 | 0.0034 | 0.0016 | 0.17 | 0.0057 | 4.9 | 352 | 0.18 | 0.0073 | 0.0010 | 0.0022 | 0.0006 | 0.0001 |
| 100 | 0.0034 | 0.0034 | 0.0016 | 0.17 | 0.0057 | 4.9 | 352 | 0.18 | 0.0073 | 0.0010 | 0.0022 | 0.0006 | 0.0001 |

$\mathrm{n}=$ Formaldehyde, $\mathrm{Bu}=1,3$ Butadiene, $\mathrm{Acr}=$ Acrolein

Table 3d-2035 American Truck Tailpipe Emissions (g/VKT)

| Speed $(\mathrm{km} / \mathrm{h})$ | PM | $\mathrm{PM}_{10}$ | $\mathrm{PM}_{2.5}$ | NOx | SOx | CO | $\mathrm{CO}_{2}$ | VOC | Bn | Ac | Fm | Bu | Acr |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Idle | 0.0458 | 0.0458 | 0.1557 | 115.65 | 0.0800 | 51.50 | 6228 | 0.85 | 0.0093 | 0.0255 | 0.0693 | 0.0054 | 0.0041 |
| 25 | 0.0114 | 0.0114 | 0.0058 | 0.34 | 0.0066 | 0.26 | 960 | 0.27 | 0.0030 | 0.0082 | 0.0222 | 0.0017 | 0.0013 |
| 50 | 0.0114 | 0.0114 | 0.0058 | 0.29 | 0.0066 | 0.13 | 960 | 0.15 | 0.0017 | 0.0046 | 0.0126 | 0.0010 | 0.0008 |
| 75 | 0.0114 | 0.0114 | 0.0058 | 0.43 | 0.0066 | 0.14 | 960 | 0.13 | 0.0015 | 0.0040 | 0.0108 | 0.0008 | 0.0006 |
| 100 | 0.0114 | 0.0114 | 0.0058 | 0.43 | 0.0066 | 0.14 | 960 | 0.13 | 0.0015 | 0.0040 | 0.0108 | 0.0008 | 0.0006 |

$\mathrm{n}=$ Formaldehyde, $\mathrm{Bu}=1,3$ Butadiene, Acr = Acrolein

## 2. Road Dust Emissions

Emissions of road dust (TSP, $\mathrm{PM}_{10}$, and $\mathrm{PM}_{2.5}$ ) resulting from vehicular travel on paved roads were estimated using the empirical expression (Equation 1) and parameters (Tables 13.2.1-1 and 13.2.1-2) provided in Section 13.2.1: Paved Roads of the U.S. EPA AP-42 document.

$$
\mathrm{EF}_{(g \mathrm{VKT})}=k *\left(\frac{s L}{2}\right)^{0.65} *\left(\frac{W}{3}\right)^{1.5}-C
$$

where,
EF = particle emission factor (having units matching the units of k )
$\mathrm{k} \quad=$ particle size multiplier (see Table 1)
$\mathrm{sL} \quad=$ road surface silt content $\left(\mathrm{g} / \mathrm{m}^{2}\right)$ (see Table 2)
$\mathrm{W} \quad=$ average weight (tons) of the vehicles traveling the road
C $\quad=$ emission factor for 1980 's vehicle fleet exhaust, brake wear and tire wear (see Table 1)
VKT = vehicle kilometers travelled

Table 4-Paved Road Parameters

| Constant | TSP | $\mathrm{PM}_{10}$ | $\mathrm{PM}_{2.5}$ |
| :---: | :---: | :---: | :---: |
| $\mathrm{k}(\mathrm{g} / \mathrm{VKT})$ | 24 | 4.6 | 0.66 |
| $\mathrm{C}(\mathrm{g} / \mathrm{VKT})$ | 0.1317 | 0.1317 | 0.1005 |

Table 5 - Silt Loading Default Values

| Constant |  | Average Travel (No. of Vehicles) |  |  |
| ---: | :---: | :---: | :---: | :---: |
|  | $<500$ | $5,000-10,000$ | $>5000$ |  |
| sL | 0.6 | 0.06 | 0.03 |  |

## Estimating the W

(a) To calculate W , the car and truck contributions to the total VKT must first be determined.
$V K T_{\text {Total }}=V K T_{\text {car }}+V K T_{\text {truck }}$
(b) the weight of each type of vehicle must be determined

Average weight of car=3.5 tons
Average weight of truck $=20$ tons
(c) the average weight (tons) of the vehicles traveling the road can be determined:

$$
\mathrm{W}=W_{\text {car }} * \frac{V K T_{\text {car }}}{V K T_{\text {Total }}}+W_{\text {truck }} \frac{V K T_{\text {truck }}}{V K T_{\text {car }}}
$$

## 3. Total PM ${ }_{2.5}$ Emissions

Total_PM $\sum_{2} E R_{(g / s)}=\left[\operatorname{TailpipeEE~}_{(g / V K T)}+\right.$ RoadDustEF $\left._{(g / V K T)}\right] X V K T_{\text {Total }(k g / h r)} x \frac{1 \mathrm{hr}}{3600 \mathrm{~s}}$

## $\mathrm{NO}_{\mathrm{x}}$ Emissions

Emissions of $\mathrm{NO}_{\mathrm{x}}$ from vehicle travel on roadways results solely from tailpipe emissions. The NOx tailpipe emissions were estimated in the same manner as the $\mathrm{PM}_{2.5}$ tailpipe emissions, and using the emission factors included above in Tables 1a through 3c.
(a) $\mathrm{VKT}_{\text {Total }}=\mathrm{VKT}_{\text {CDN_car }}+\mathrm{VKT}_{\text {CDN_truck }}+\mathrm{VKT}_{\mathrm{US} \text { _car }}+\mathrm{VKT}_{\text {US_Truck }}$
(b) Fleet Average $\mathrm{EF}_{(\mathrm{g} / \mathrm{VKT})}=$
$E F_{C D N_{-} \text {car }} * \frac{V K T_{C D N_{-} \text {car }}}{V K T_{\text {Total }}}+E F_{C D N_{-} \text {truck }} \frac{V K T_{\text {CDN_truck }}}{V K T_{\text {car }}}+E F_{\text {US_car }} \frac{V K T_{U S_{-} \text {car }}}{V K T_{\text {Total }}}+E F_{U S_{-} \text {truck }} \frac{V K T_{U S_{-} \text {truck }}}{V K T_{\text {Total }}}$
(c) NOXTailpipeE $R_{(g / s)}=[$ TailpipeEF $(\mathrm{g} / \mathrm{VKr})] \times V K T_{\text {Total(kg/hr) }} \times \frac{1 \mathrm{hr}}{3600 \mathrm{~s}}$

## Queueing at the Customs/Inspection Plazas

Key assumptions:

- Inbound vehicles at customs plaza will queue at inspection booths.
- Outbound vehicles at customs plaza will not queue.
- Queuing traffic volume is same as free-flowing traffic volume.
- There is always queuing (idling) at the booth due to the one vehicle in the booth being inspected.
- Inspection times for cars and trucks are 45 seconds and 60 seconds, respectively.


## Customs Plaza Queuing Algorithm:

Groups of queue links were set up for each plaza based on an equal distribution of free flow traffic through each booth that is open during a given hour. Then each queue link was manually "turned on" or "off" by calculating the number of vehicles queued. This modeling approach represents the actual situation because not all groups of queue links actually experience queuing for a given hour.

The amount of queuing at each booth was calculated manually for each group of queue links and for each hour using the hourly free flow traffic volume and the number of booths that are open during each hour, which varies by demand.

1. For each hour, the number of booths that are open is calculated using the hourly free flow traffic volume and the inspection time for each vehicle.
2. The number of vehicles passing through each booth is then back calculated.
3. The calculated number from Step 2 is then compared with the capacity of each booth, i.e., 80 for cars and 60 for trucks. If the number is less than its capacity, then no queuing in this hour; if greater than its capacity, then queuing will occur and the difference is the number of vehicles queued at the booth during that hour.
4. Based on the results obtained from Step 3, the queue links are either "turned on" (with queuing) or "off" (no queuing).
5. If there is queuing, and the queue length per booth exceeds 4 trucks or 6 cars, an additional booth is opened, if possible.
6. If there are no more booths to open, the queue length extends far enough back to accommodate the number of vehicles waiting at the plaza. The locations depend on the physical configuration of each plaza; if the number of vehicles queued determined from Step 3 exceeds the physical length of the queue link, then the next corresponding group of queue links will be "turned on", and so on.

For example, for an hour with 1004 truck traffic, the number of booths that are needed is 1004 / $60=17$. Then the number of trucks passing through each booth is back calculated: $1004 / 17=59$. Since this number is less than the capacity of each booth ( 60 trucks per hour), there will be no queuing at each booth except for the one truck that is in the booth and being inspected.

For an hour with 443 truck traffic, the number of booths that are needed is $443 / 60=7$. Similarly, the back-calculated number of trucks passing through each booth is $443 / 7=63.3$. Theoretically there will be 3.3 trucks queuing at each booth, in addition to the one truck that is in the booth and being inspected. If the group of queue links right next to the booths are set up such that only 2 trucks can wait in line, then 7 of the next group of queue links will be "turned on" and on each link there will be 1.3 trucks queuing.

Summary of CAL3QHCR Model Inputs:

|  | Cars | Trucks |
| :--- | :--- | :--- |
| Number of queuing lanes | 1 | 1 |
| Light cycle time | 45 seconds | 60 seconds |
| Yellow time | 0 seconds | 0 seconds |
| Red duration time | 40 seconds | 55 seconds |
| Saturated flow volume (veh/hr/lane) | 1200 | 1200 |
| Signal type | 2 | 2 |
| Arrival rate | 1 | 1 |
| Maximum number of booths at each plaza | 20 | 19 |


[^0]:    ${ }^{1}$ Note that there is a significant difference between the $\mathrm{NO}_{x}$ concentrations of the 1 B Option 2 route alignment in comparison to Alternative 3 at 50 m away.

[^1]:    * For consistency, Huron Church Rd/Talbot Rd runs North-South and all crossing roads run East-West

[^2]:    *     - For consistency, Huron Church Rd/Talbot Rd runs North-South and all crossing roads run East-West

[^3]:    *     - For consistency, Huron Church Rd/Talbot Rd runs North-South and all crossing roads run East-West

[^4]:    * For consistency, Huron Church Rd/Talbot Rd runs North-South and all crossing roads run East-West

[^5]:    *     - For consistency, Huron Church Rd/Talbot Rd runs North-South and all crossing roads run East-West

[^6]:    *     - For consistency, Huron Church Rd/Talbot Rd runs North-South and all crossing roads run East-West

[^7]:    - For consistency, Huron Church Rd/Talbot Rd runs North-South and all crossing roads run East-West

[^8]:    *     - For consistency, Huron Church Rd/Talbot Rd runs North-South and all crossing roads run East-West

[^9]:    $\mathrm{Bn}=$ Benzene, $\mathrm{Ac}=$ Acetaldehyde, $\mathrm{Fm}=$ Formaldehyde, $\mathrm{Bu}=1,3$ Butadiene, $\mathrm{Acr}=$ Acrolein

